

Document Control

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Glossary

AoO - advice on operations – Contained within the conservation advice packages from Natural England and JNCC, the AoO details the pressure/gear combinations a feature may be sensitive to.

Attribute - Selected characteristic of an interest feature/sub-feature which contributes to the overall condition of the feature to which it applies.

Broad-scale habitat – A categorisation of habitats based on a shared set of ecological requirements. Broad-scale habitats are one type of MCZ feature, the other being Features of Conservation Interest (FOCI). More information can be found in the Ecological Network Guidance (Marine Conservation Zone Project) section 4.2.3¹.

Catch recording service - MMO's catch recording service was developed to allow fishers to create and submit records of daily catches for English and Welsh under 10 metre flag vessels that fish in UK waters.

Cefas - Centre for Environment, Fisheries and Aquaculture Science. Cefas is a government agency that carries out research, consultancy and advisory work.

Conservation objectives - Conservation objectives are set for each designated feature of an MPA, to either maintain or restore a designated feature of the protected site.

Designated features – Habitats or species within an MPA which have been designated as protected features.

EMS – European marine site. Any special protection areas (SPAs) and special areas of conservation (SACs) that are covered by tidal waters.

Exposure - The level at which a designated feature or its supporting habitat is open to a distressing influence resulting from the possible/likely effects of operations arising from human activities (e.g. fishing) currently occurring on the site. The assessment of exposure can include the spatial extent, frequency, duration and intensity of the pressure(s) associated with the activities, where this information is available.

Fishermap - In 2012 the Fishermap project mapped the activities of the commercial fishing fleet, by interviewing skippers and collating data to show fishing activity and gear types used in map grid cells.

¹ <https://hub.jncc.gov.uk/assets/94f961af-0bfc-4787-92d7-0c3bcf0fd083>

Habitats Directive – Council Directive 92/43/EEC on the conservation of natural habitats and of wild fauna and flora².

Habitats Regulations – The Conservation of Habitats and Species Regulations 2017 (S.I. 2010/490)³.

HOCl – habitat of conservation importance. Habitats that are threatened, rare, or declining. More information can be found in the Ecological Network Guidance (Marine Conservation Zone Project) section 4.2.3⁴.

IFCA – Inshore Fisheries Conservation Authority. IFCAs are responsible for fisheries management from 0 to 6 nautical miles (nm). There are 10 IFCAs in England, each one funded by local authorities.

ICES – International Council for the Exploration of the Sea. ICES is an intergovernmental marine science organisation, providing evidence on the state and sustainable use of our seas and oceans.

JNCC – Joint Nature Conservation Committee. A public body that advises the government on UK and international nature conservation. This includes aspects related to the marine environment from 12 nm to 200 nm.

Marine plans – MMO marine plans have been designed to help manage the seas around England⁵.

MCRS – minimum conservation reference size. MCRS is the minimum size at which an ocean species can be landed for human consumption. MCRS for many species are listed in the annexes of the Technical Conservation Regulations (EU) 2019/1241⁶. Several pieces of domestic legislation also implement MCRS for certain species.

MCZ – marine conservation zone. Marine conservation zones are a type of MPA in English, Welsh and Northern Irish waters designated under the Marine and Coastal Access Act 2009⁷ (for England and Wales) or The Marine Act (Northern Ireland) 2013⁸ (for Northern Ireland).

MPA – marine protected area. Marine protected areas are protected sites with a marine element, this includes special areas of conservation (SAC), special protection areas (SPA) and marine conservation zones (MCZ).

² <https://www.legislation.gov.uk/eudr/1992/43/contents>

³ <https://www.legislation.gov.uk/ukxi/2017/1012/contents/made>

⁴ <https://hub.jncc.gov.uk/assets/94f961af-0bfc-4787-92d7-0c3bcf0fd083>

⁵ <https://www.gov.uk/government/collections/marine-planning-in-england>

⁶ <https://www.legislation.gov.uk/eur/2019/1241/contents>

⁷ <https://www.legislation.gov.uk/ukpga/2009/23/contents>

⁸ <https://www.legislation.gov.uk/nia/2013/10/contents>

MPA assessment – MPA site level assessments are carried out in a manner consistent with the requirements of Article 6 (3) of the Habitats Regulations for EMSs and the requirements of section 126 of the Marine and Coastal Access Act 2009 for MCZs. For EMSs the assessments will determine whether, in light of the site's conservation objectives, fishing activities are having an adverse effect on the integrity of the site. For MCZs the assessments will determine whether there is a significant risk of fishing activities hindering the conservation objectives and general management approach of the site.

Natural England - Government advisor for the environment in England. This includes aspects of the marine environment of 0 to 12 nm.

Offshore Habitats Regulations – The Conservation of Offshore Marine Habitats and Species Regulations 2017 (S.I. 2007/1842)⁹.

PAD – Pressure Activity Database. This JNCC database supports the advice on operations for UK offshore MPAs and is used to determine whether pressures are likely to have a significant effect on a site's features.

Pr-value – fishing footprint value. Defines the level of pressure for a single average day of effort for a reference vessel or fisher (land-based) within a fleet, taking into account the gear used. The value can be multiplied by the number of vessels or fishers to give the total pressure for a particular gear over a specific time period.

SAC – special area of conservation. Special areas of conservation are MPAs put in place to protect habitats and species listed in Annexes I and II of Council Directive 92/43/EEC (the Habitats Directive).

SCI – Site of community importance. Defined by the Council Directive 92/43/EEC (the Habitats Directive) as a site which contributes significantly to the maintenance or restoration at a favourable conservation status of a natural habitat type or of a species in the biogeographical region or regions to which it belongs.

Site integrity – The integrity of a site is the coherence of its ecological structure and function, across its whole area, which enables it to sustain the habitat, complex of habitats and/or the levels of populations of the species for which it was designated¹⁰.

Sensitivity assessment – Assessment of sensitivity of a species or habitat which takes into account ability to resist impacts, and rate of recovery after an impact.

⁹ <https://www.legislation.gov.uk/ukxi/2017/1013/contents/made>

¹⁰ <https://www.gov.uk/guidance/appropriate-assessment>

SNCB - statutory nature conservation body. A collective term for Natural Resources Wales (NRW), Joint Nature Conservation Committee (JNCC), Natural England (NE), Northern Ireland's Council for Nature Conservation and the Countryside (which generally works through the Northern Ireland Environment Agency) and NatureScot. These organisations have a statutory responsibility to provide conservation advice for MPAs and report on the condition of protected features.

SPA – special protection area. Special protection areas are MPAs put into place to protect threatened bird species, designated under the Wild Birds Directive.

SPIRIT - SPatial InfoRmation Toolkit. SPIRIT is the MMO Geographic Information System used for mapping environmental and other data.

SOCI – species of conservation importance. Species that are threatened, rare, or declining. More information can be found in the Ecological Network Guidance (Marine Conservation Zone Project) section 4.2.3¹¹

Target - This defines the desired condition of an attribute, taking into account fluctuations due to natural change.

VMS – vessel monitoring system. All commercial fishing vessels over 12 metres in length in UK waters must report their position via VMS when at sea. VMS devices on the vessels send regular reports of position and vector.

¹¹ <https://hub.jncc.gov.uk/assets/94f961af-0bfc-4787-92d7-0c3bcf0fd083>

1. Summary

Table 1 provides a summary of the outcomes of this assessment regarding the impact of fishing activities on site features.

Table 1: Assessment Summary.

Feature	Activity/gear	Part A outcome	Part B outcome	Part C outcome: In-combination assessment
H1110 1110 Sandbanks which are slightly covered by sea water all the time	Beam trawl (pulse/wing)	Likely to have a significant effect	May result in adverse effect on site integrity	N/A
	Mussels, clams, oysters dredges			
	Pump scoop dredges (cockles, clams)	Not likely to have a significant effect	N/A	N/A
	Suction dredges (cockles)			
	Hand working (access from vessel)			
	Longlines (pelagic)			
	Longlines (demersal)			
	Pots/creels (crustacea/gastropods)	Likely to have a significant effect	Will not result in adverse effect on site integrity	Will not result in adverse effect on site integrity
	Cuttle pots	Not likely to have a significant effect	N/A	N/A
	Fish traps			
	Drift nets (pelagic)			
	Drift nets (demersal)			
	Crab tiling			
	Digging with forks			
	Purse seine			
	Mid-water trawl (single) (pelagic)			
	Mid-water trawl (pair) (pelagic)			
	Gill nets	Likely to have a significant effect	Will not result in adverse effect on site integrity	Will not result in adverse effect on site integrity
	Trammels nets			
Entangling nets	May result in adverse effect on site integrity		N/A	
Beam trawl (whitefish)				
Beam trawl (shrimp)				
Heavy otter trawl	Not likely to have a significant effect	N/A	N/A	
Multi-rig trawls				

	Light otter trawl	Likely to have a significant effect	May result in adverse effect on site integrity			
	Pair trawl					
	Anchor seine					
	Scottish/fly seine					
	Scallop dredges					
H1170 Reef <i>Sabellaria spinulosa</i> reefs	Beam trawl (pulse/wing)	Likely to have a significant effect	N/A	N/A		
	Mussels, clams, oysters dredges					
	Pump scoop dredges (cockles, clams)					
	Suction dredges (cockles)	Not likely to have a significant effect				
	Hand working (access from vessel)					
	Longlines (pelagic)					
	Drift nets (pelagic)					
	Crab tiling					
	Digging with forks					
	Purse seine					
	Mid-water trawl (single) (pelagic)					
	Mid-water trawl (pair) (pelagic)					
	Pots/creels (crustacea/gastropods)				Likely to have a significant effect	May result in adverse effect on site integrity
	Cuttle pots	Not likely to have a significant effect			N/A	
	Fish traps					
	Longlines (demersal)	Likely to have a significant effect			May result in adverse effect on site integrity	N/A
	Drift nets (demersal)	Not likely to have a significant effect			N/A	
	Gill nets	Likely to have a significant effect			May result in adverse effect on site integrity	
	Trammels					
	Entangling					
	Beam trawl (whitefish)	Likely to have a significant effect			N/A	N/A
	Beam trawl (shrimp)					
	Heavy otter trawl					
	Multi-rig trawls					
	Light otter trawl					
	Pair trawl					
	Anchor seine					
	Scottish/fly seine					
	Scallop dredges					

2. Introduction

Table 2 shows the name and legal status of the site. Inner Dowsing, Race Bank and North Ridge Special Area of Conservation (SAC) is located off the south Lincolnshire coast, to the east of Skegness and extends eastwards and north from Burnham Flats on the North Norfolk coast.

Table 2: Site details.

Name of site	Legal status
Inner Dowsing, Race Bank and North Ridge	Special Area of Conservation (SAC)

The SAC subject to this assessment is within International Council for the Exploration of the Sea (ICES) rectangle 35F0 and 35F1. The site is situated on the approaches to The Wash, it crosses the 6 nautical mile (nm) boundary and therefore falls under two different administrative areas: the District of the Eastern Inshore Fisheries and Conservation Authority (Eastern IFCA; 0 – 6 nm) and the Marine Management Organisation (MMO; beyond 6 nm). These different administrative areas will be denoted as inshore (Eastern IFC District) and offshore (MMO area) throughout the assessment. This assessment assesses the impacts of fishing in the part of the site within the MMO area (i.e. offshore of 6 nm).

The site is designated for the Annex I features (of the Habitats Regulations) 'sandbanks which are slightly covered by sea water all the time' and 'reef' (*Sabellaria spinulosa* reefs, Table 3). The sandbanks are important headland-associated offshore systems. Water depths are generally shallow and mostly less than 30 meters deep.

Inner Dowsing, Race Bank and North Ridge SAC crosses the 12 nm boundary and is covered by both Natural England and Joint Nature Conservation Committee (JNCC) who have provided joint advice on this conservation advice package¹². In 2019, the site condition was reassessed and found to be in unfavourable condition and the conservation objective is to restore both features to favourable condition (Table 3).

The conservation objectives for the SAC are to ensure that, subject to natural change, the integrity of the site is maintained or restored as appropriate, and that the site contributes to achieving the Favourable Conservation Status of its qualifying features, by maintaining or restoring:

- the extent and distribution of qualifying natural habitats and habitats of the qualifying species
- the structure and function (including typical species) of qualifying natural habitats

¹² The Natural England and JNCC [conservation advice package](#)

- the structure and function of the habitats of the qualifying species
- the supporting processes on which qualifying natural habitats and the habitats of qualifying species rely
- the populations of each of the qualifying species
- the distribution of qualifying species within the site

Abundant *Sabellaria spinulosa* (*S. spinulosa*) agglomerations have consistently been recorded within the boundary of the site, the patches of reef have been found within 6 nm, in the 6 - 12 nm portion of the site and outside of 12 nm, the reef patches also straddle the 12 nm boundary. Survey data indicate that reef structures are concentrated in certain areas of the site, with a patchy distribution of crust-forming aggregations across the site. Reef distribution is determined by a range of factors including the hydrodynamics, bathymetry and suspended sediment loads within the site, as well as disturbance from anthropogenic activities.

Figure 1 shows the current understanding of the distribution of designated features within the Inner Dowsing, Race Bank and North Ridge SAC. Feature data were provided to MMO by Natural England and JNCC on 29 July 2020.

Table 3: Designated features.

Feature	Sub-feature	Conservation objective
Sandbanks which are slightly covered by sea water all the time	Subtidal sand, subtidal coarse sediment, subtidal mixed sediments	Restore feature to favourable condition
Reef	Subtidal biogenic reefs: <i>Sabellaria spinulosa</i> .	Restore feature to favourable condition

2.1 Sandbanks

Two areas of sandbanks which are slightly covered by water at all times feature in this site, the Inner Dowsing sandbank occurs beyond 6 nm and the Race Bank sandbank which extends mostly beyond 12 nm (Figure 1).

The Natural England and JNCC conservation advice packages for this site note that sandbanks feature consists of three sub-features:

- subtidal coarse sediment;
- subtidal mixed sediment; and
- subtidal sand.

The crests and flanks of the sandbanks are characterised by low diversity communities of polychaete worms and amphipods. The troughs which separate the sandbanks contain a diverse mosaic of biotopes on mixed and gravelly sands.

The sandbanks provide an ideal spawning and nursery ground for commercially important fish such as the sandeel (*Ammodytes sp.*) and Atlantic herring (*Clupea harengus*) while also providing important feeding grounds for lemon sole (*Microstomus kitt*) and European plaice (*Pleuronectes platessa*)¹³.

Within the distribution of designated features (Figure 1), Natural England and JNCC have applied a 500 metre margin to the known area of sandbank as a result of uncertainty in relation to feature delineation, as well as potential for broad-scale migration of the sandbank feature as well as finer scale oscillation due to hydrological processes¹⁴. Natural England and JNCC have advised that the entire area (including the margin) should be managed as sandbank in order to incorporate these processes.

2.2 Reefs

Sabellaria spinulosa reefs are an ephemeral sub-feature which is subject to constant creation and destruction across its possible range, it can also exhibit relative stability and be continuously present in named areas for prolonged periods (Foster-Smith and Hendrick, 2003). Identified areas of *S. spinulosa* reef are therefore subject to a higher level of uncertainty than many other habitat features. *S. spinulosa* reefs provide vital attachment points for both infauna and epifauna, as well as stabilising sediments. They support a variety of bryozoans, hydroids, sponges and anemones as well as the common lobster, *Homarus gammarus*, and the commercially exploitable pink shrimp, *Pandalus montagui*. Despite the widespread occurrence of the species *S. spinulosa*, there are few known areas of well developed biogenic reef formed by *S. spinulosa* in UK waters (and very few in other European waters) (JNCC and Natural England, 2010).

Biogenic reef created by *S. spinulosa* has consistently been recorded within the site and the aggregations provide additional hard substrate for the development of rich epifaunal communities. Representative communities equate to a number of biotopes including SS.SSa.IFiSa.NcirBat (*Nephtys cirrosa* and *Bathyporeia* spp. in infralittoral sand) and SS.SSa.IFiSa.IMoSa (infralittoral mobile clean sand with sparse fauna (Entec, 2008). In more gravelly areas a diverse attached epifauna is present, including bryozoans, sponges, hydroids (*Hydrallmania falcata*, *Tubularia indivisa*) and tube building worms (*Pomatoceros* sp). The tube building amphipod *Ampelisca diadema* is also abundant in some areas. Mobile epifauna include a variety of brittle stars and small crabs as well as pink shrimp (*Pandalus montagui*) and lobster (*Homarus gammarus*). In a number of places mussel (*Mytilus edulis*) density is reasonably high (JNCC and Natural England, 2010).

¹³ The Natural England and JNCC [conservation advice package](#)

¹⁴ NE and JNCC Formal Advice to MMO, 11th September 2015

As the SAC is primarily composed of soft sediments, where reef is present it provides an important habitat for both sessile and mobile species. *S. spinulosa* reefs can enhance both biomass and biodiversity. They support ascidians such as *Dendrodoa grossularia* and the encrusting worm *Pomatoceros lamarckii*. Several mobile epifauna species are associated with the reefs, including the queen scallop (*Aequipecten opercularis*), squat lobster (*Galathea* spp.) and *P. montagui*, which feeds on *S. spinulosa*. *S. spinulosa* reefs are commonly associated with crabs such as *Liocarcinus* sp. and *Pisidia longicornis*, the burrowing anemone *Cerianthus lloydii* and brittlestars (*Ophiura* spp.)¹³.

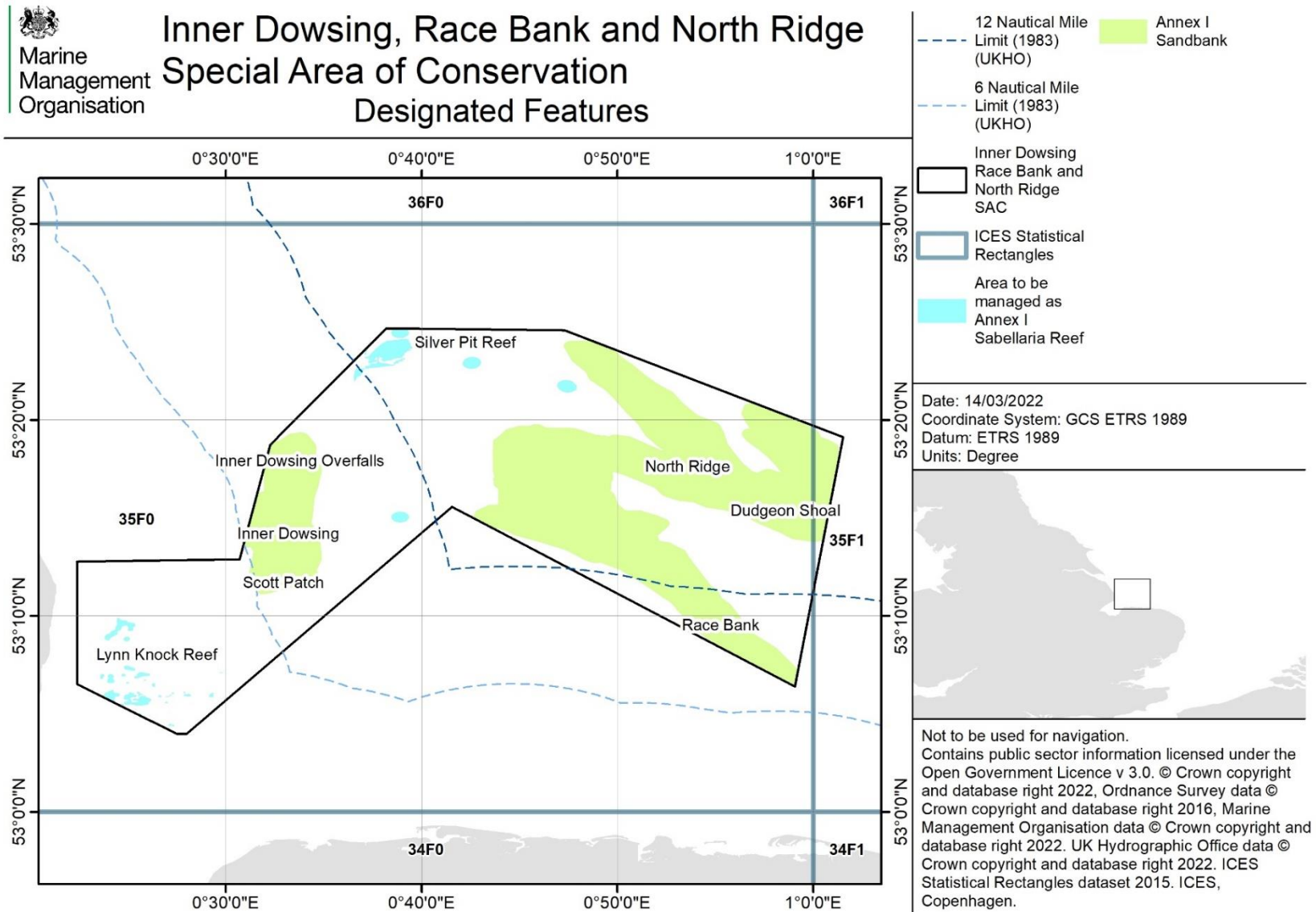
Natural England and JNCC have applied a 500 metre margin around reef polyline and point data, shown in Figure 1, in order to account for uncertainty in reef extent due to the inability of the ground truthing data used to provide information on reef extent¹⁴.

Some areas of *S. spinulosa* reef in the site are currently subject to management through the [Inner Dowsing, Race Bank and North Ridge European Marine Site \(Specified Areas\) Bottom Towed Fishing Gear Byelaw](#), an MMO byelaw prohibiting bottom towed gears over these areas. One of these areas lies between the 6 - 12 nm boundaries and the remaining two areas lie inshore of 6 nm¹⁵. There is an Eastern IFCA byelaw¹⁶ protecting the 0 - 6 nm section of the site, including a patch of *S. spinulosa* reef. Further information can be found in Section 2.3 with Figure 1 detailing the site features and nautical boundaries.

¹⁵ The MMO Inner Dowsing, Race Bank and North Ridge European Marine Site (Specified Areas) Bottom Towed Fishing Gear Byelaw. Available at: www.gov.uk/government/publications/inner-dowsing-race-bank-and-north-ridge-european-marine-site-specified-areas-bottom-towed-fishing-gear-byelaw

¹⁶ <https://www.eastern-ifca.gov.uk/wp-content/uploads/2020/05/2018-MPA-Byelaw-Guidance.pdf>

Figure 1: Inner Dowsing, Race Bank and North Ridge SAC features



2.3 Scope of this assessment – fishing activities assessed

The geographic scope of this assessment covers the site outside 6 nm (hereafter the ‘MMO portion’), this includes both designated features (Figure 1). Eastern IFCA are responsible for managing fishing in the 0 nm – 6 nm area and are currently developing management measures to protect *S. spinulosa* reef in the Lynn Knock area which falls under IFCA jurisdiction. Eastern IFCA provided relevant advice and information on local fishing activities during the preparation of this assessment.

All commercial fishing activities, excluding those specified below, have been included for assessment (Table 4):

- Pelagic fishing – the designated features of this site are seabed features and therefore will not be affected by pelagic fishing; and
- Shore based fishing – this assessment covers the portion of the site offshore of 6 nm and therefore will not be affected by shore based fishing (see Table 7 for examples of shore-based fishing activities).

A revised approach to the management of commercial fisheries in European marine sites¹⁷ was established in 2012 (see Annex 1 - MMO methodology for further details). A matrix was developed to aid regulators in assessing whether management measures should be introduced in marine protected areas¹⁸. Table 4 displays the matrix interactions for the aggregated method fishing activities and designated features.

Interactions are considered a ‘red risk’ where it is clear that the conservation objectives for a feature (or sub-feature) will not be achieved because of its sensitivity to a type of fishing - irrespective of feature condition, level of pressure, or background environmental conditions in all EMSs where that feature occurs¹⁹.

Table 4: Aggregated method fishing activities with amber or red interactions.

Feature/Fishing gear type	Sandbank	Reef
Towed (demersal)	Amber	Red
Dredges (towed)	Amber	Red
Static (pots/traps)	Amber	Amber
Static (anchored nets/lines)	Amber	Amber

Interactions are considered an ‘amber risk’ where there is risk as to whether conservation objectives for a feature will be achieved because of its sensitivity to a type of fishing. Interactions classified as amber are subject to full assessment to

¹⁷ www.gov.uk/government/publications/revised-approach-to-the-management-of-commercial-fisheries-in-european-marine-sites-overarching-policy-and-delivery

¹⁸ https://assets.publishing.service.gov.uk/government/uploads/system/uploads/attachment_data/file/310814/cefás_matrix_review.pdf

¹⁹ https://assets.publishing.service.gov.uk/government/uploads/system/uploads/attachment_data/file/310822/matrixbackground.pdf

determine whether management of activity is required to further the site's conservation objectives.

As demersal trawls and dredging all have red interactions with *S. spinulosa* reef, a site level assessment is not required, and the interaction is automatically addressed through a management measure. Evidence supporting the 'red risk' categorisation for these gear/feature interactions is set out in a separate document²⁰. Management measures will therefore be introduced to prohibit bottom-towed gear across the *S. spinulosa* reef areas.

Table 5 shows the fishing activities classified as having amber interactions with features of this site beyond 6 nm. The 'Matrix gear type' column shows the categories used in the Matrix. These are matched to the 'aggregated method' categories used in NE and JNCC conservation advice packages.

Commercial sea fishing has the potential to vary in nature and intensity over time. This assessment considers a particular range of recent and likely future activity based on activity levels and type as identified in section 4.

Fishing activity will be monitored at the site and, should future effort significantly increase from the range assessed, this assessment may be reviewed and updated in future. See Section 8 for information on ongoing monitoring and control plans at this site.

Fisheries assessments use the best available evidence to fully consider all impacts against the conservation objectives. If the assessment cannot conclude that the use of certain fishing gear types is compatible with the site's conservation objectives, appropriate management measures will be introduced. Although management measures implemented may cause displacement of fishing activity to areas outside of the SAC, or between areas of feature within the SAC, it is not possible to accurately predict the location (and thus the associated environmental costs) of displaced fishing activity. Furthermore, this potential displacement of fishing activity does not remove the requirement to introduce management to protect the designated features. MMO closely monitors fishing activity in every marine protected area (MPA) for which MMO has statutory responsibility. MMO regularly reviews and updates fisheries assessments to reflect any circumstantial changes, including displacement of fishing activity

The scope of this assessment covers fishing activities alone, and other plans or projects in combination with fishing. It does not cover other activities defined by MMO as marine non-licensable activities. Marine non-licensable activities include

²⁰

https://assets.publishing.service.gov.uk/government/uploads/system/uploads/attachment_data/file/310819/sabellaria.pdf

activities such as sailing and powerboating, as well as associated activities like anchoring. More information about MMO's management of marine non-licensable activities can be found online²¹.

Table 5: Fishing activities with amber interactions assessed for feature.

Feature	Matrix Gear Type	Gear Code	SNCB aggregated gear method
Sandbanks	Beam trawl (whitefish)	TBB	Towed (demersal/pelagic)
	Beam trawl (shrimp)		
	Beam trawl (pulse/wing)		
	Heavy otter trawl	OTB	
	Multi-rig trawls	TX	
	Light otter trawl	OTB	
	Pair trawl	PTB	
	Anchor seine	SDN	
	Scottish/fly seine	SSC	
	Towed gear (demersal/pelagic)	-	
	Scallop dredge	DRB	Dredges
	Mussel/clam/oyster dredge	DRB/HMD	
	Pump scoop (cockles, clams)	HMP/HMD	
	Suction dredge (cockles)	HMD	
<i>S. spinulosa</i> reefs and Sandbanks	Pots/creels (crustacea/gastropods)	FPO	Traps
	Cuttle pots		
	Fish traps		
	Gill nets	GNS	Anchored nets/lines
	Trammels	GTR	
	Entangling	GN	
	Drift nets (demersal)	GND	
<i>S. spinulosa</i> reefs	Long lines (demersal)	LLS	Diving
	Commercial diving	-	

²¹ <https://www.gov.uk/government/publications/managing-marine-non-licensable-activities-in-marine-protected-areas>

3. Part A Assessment

Part A of this assessment was carried out in a manner that is consistent with the likely significant effect (LSE) test required by Article 63 of the Habitats Regulations and Article 28 of the Offshore Habitats Regulations.

For each fishing activity, a series of questions were asked²²:

1. Does the activity take place, or is it likely to take place in the future?
2. What are the potential pressures exerted by the activity on the feature?
3. Are the effects/impacts of the pressures likely to be significant?

For each activity assessed in Part A, there were two possible outcomes for each identified pressure-feature interaction:

1. The pressure-feature interactions were not included for assessment in Part B if:
 - i. the feature is not exposed to the pressure, and is not likely to be in the future; or
 - ii. the effect/impact of the pressure is not likely to be significant.
2. The pressure-feature interactions were included for assessment in Part B if:
 - i. the feature is exposed to the pressure, or is likely to be in the future; and
 - ii. the potential scale or magnitude of any effect is likely to be significant; or
 - iii. it is not possible to determine whether the magnitude of any effect is likely to be significant.

The conservation advice package used to inform this assessment is provided in Table 6.

Table 6: Conservation advice package used for assessment.

Feature	Package	Link
Sandbank and Reef	Natural England and JNCC Conservation Advice for Inner Dowsing, Race Bank and North Ridge SAC (UK0030370)	https://designatedsites.naturalengland.org.uk/Marine/MarineSiteDetail.aspx?SiteCode=UK0030370&SiteName=inner%20dows&countyCode=&responsiblePerson=&SeaArea=&IFCAArea=&HasCA=1&NumMarineSeasonality=0&SiteNameDisplay=Inner%20Dowsing,%20Race%20Bank%20and%20North%20Ridge%20SAC

²² The test for likely significant effect under article 63 of the Habitats Regulations is not required for activities which are directly connected to or necessary to site management, this includes fishing activities unless otherwise indicated.

To support the conservation advice shown in Table 6, MMO has also used the Pressures Activities Database (PAD) published by JNCC²³. In particular, the PAD has been used to assess impacts to mobile species. Mobile species are not a designated feature of Inner Dowsing, Race Bank and North Ridge SAC, however, the conservation advice package may include species (including mobile species) as a component part of a feature, and impacts on certain species may influence a target attribute for a site feature. Where fishing impacts (for example, the removal of target and non-target species) has the potential to impact a sites' conservation objectives, MMO has used the best available evidence to assess this, including the PAD.

3.1 Activities not taking place

Table 7 shows activities which are excluded from further assessment as they do not take place and are not likely to take place in the future.

Table 7: Activities not taking place and not likely to take place in the future.

Feature	Gear type	Justification
Sandbank and Reef	Heavy otter trawl	Ground conditions are unsuitable for heavy otter trawls.
	Cuttle pots	Gear codes used in gear register and logbooks do not distinguish cuttle pots from crab/lobster pots. However, expert opinion from local MMO and IFC officers state with high confidence that cuttle pots are not used at this site.
	Fish traps	Fish traps are used in rivers and estuaries therefore this site is not a suitable location for this gear.
	Drift nets (demersal)	Does not occur at this site according to VMS ²⁴ and corroborated with local MMO and IFCA knowledge.
	Beach seine/ring nets	The SAC is at least 1 km offshore and not subject to shore-based activities.
	Shrimp push nets	
	Fyke and stake nets	
	Hand-working (vessel/land access)	
	Crab tiles	
	Digging with forks	Does not occur in the UK except in Poole Harbour.
Bait dragging		

²³ <https://jncc.gov.uk/our-work/marine-activities-and-pressures-evidence/>

²⁴ VMS – Vessel monitoring system to which, as of 1 January 2012, all vessels greater than 12 m in overall length should be fitted in accordance with Commission Regulation (EC) No 2244/2003 Article (2) 1 and Council Regulation (EC) No 1224/2009 Article 9(2).

	Pump scoop (cockles and clams)	Does not occur at this site according to VMS and corroborated with local MMO and IFCA knowledge.
	Suction (cockles)	Activity does not occur within the site (seed mussel fishery does not use this method).
	Commercial diving	Does not occur at this site according to VMS and corroborated with local MMO and IFCA knowledge.

3.2 Potential pressures exerted by the activities on the feature

For the remaining activities, potential pressures were identified using the Natural England conservation advice package identified in Table 6 and associated advice on operations tables. All pressures identified other than those categorised as 'not relevant' were included. Table 8 shows the potential pressures identified.

Table 8: Potential pressures on sandbank and *S. spinulosa* reef.

Feature	Aggregated method	Potential pressures
Sandbank	Traps and Anchored nets/lines	Abrasion/disturbance of the substrate on the surface of the seabed
		Removal of non-target species
		Removal of target species
		Barrier to species movement
		Deoxygenation
		Hydrocarbon and polycyclic aromatic hydrocarbon (PAH) contamination
		Introduction of light
		Introduction or spread of invasive or non-indigenous species
		Litter
		Organic enrichment
		Penetration and/or disturbance of the substratum below the surface of the seabed, including abrasion
		Synthetic compound contamination (including pesticides, antifoulants, pharmaceuticals)
		Transition elements and organo-metal (e.g. tributyltin, TBT) contamination ²⁵
		Underwater noise changes
		Visual disturbance
	Demersal trawls and	Abrasion/disturbance of the substrate on the surface of the seabed
		Changes in suspended solids (water clarity)
		Penetration and/or disturbance of the substratum below the surface of the seabed, including abrasion
		Removal of non-target species

²⁵ Includes those priority substances listed in Annex II of Directive 2008/105/EC.

	Demersal seines	Smothering and siltation rate changes (light)		
		Deoxygenation		
		Hydrocarbon and PAH contamination		
	and	Dredges	Introduction of light	
			Introduction or spread of invasive or non-indigenous species	
			Litter	
			Nutrient enrichment	
			Organic enrichment	
			Physical change (to another sediment type)	
			Synthetic compound contamination	
			Transition elements and organo-metal (e.g. TBT) contamination	
			Underwater noise changes	
			Visual disturbance	
			Dredges	Removal of target species
				Introduction of microbial pathogens
S. spinulosa reef	Traps	Abrasion/disturbance of the substrate on the surface of the seabed		
		Removal of non-target species		
		Barrier to species movement		
		Deoxygenation		
		Hydrocarbon and PAH contamination		
		Introduction or spread of invasive non-indigenous species		
	and	Anchored nets/lines	Litter	
			Organic enrichment	
			Penetration and/or disturbance of the substratum below the surface of the seabed, including abrasion	
			Synthetic compound contamination (including pesticides, antifoulants, pharmaceuticals)	
			Transition elements and organo-metal (e.g. TBT) contamination	

3.3 Significance of effects/impacts

To determine whether each pressure is capable of having an adverse effect on site integrity, the sensitivity assessments and risk profiling of pressures from the advice on operations section of the JNCC and NE conservation advice package²⁶ (Table 6) were used.

Table 9 and Table 10 identify the pressures from particular gears which are capable of having an adverse effect on site integrity. Where a pressure from a particular gear is identified as not likely to have a significant effect, justification is provided. To ensure the effects of fishing activities in-combination with other activities (including

²⁶ The Natural England and JNCC [conservation advice package](#)

other fishing activities) are fully assessed, the pressures from anchored nets/lines and traps (green activities) which are not likely to cause a significant effect but which do interact with the feature are considered in the in-combination aspect of the assessment (Section 5).

Table 9: Summary of pressures from specific activities on sandbank taken to Part B.

Potential pressures	Traps	Anchored nets/lines				Demersal Trawls				Demersal seines		Dredges	
	Pots	Gill nets	Set gill nets	Trammel net	Entangling nets	Bottom otter	Otter twin	Bottom pair	Beam	Scottish	Danish	Boat	Hand mechanised
Abrasion/disturbance of seabed surface substrate	LSE - surface disturbance is caused by contact between gear/anchors and the seabed. This occurs during setting of the pots/traps and/or by movement of the gear over the seabed, during rough weather or during retrieval.												
Removal of non-target species	LSE – bycatch is associated with almost all fishing activities and is related to factors such as the gear type and its design (i.e. its selectivity), the targeted species and effort.												
Removal of target species	LSE - species removed may be species forming part of the biotope (e.g. bivalves in sediment feature) or wider community composition associated with the designated feature. Incidental non-target catch including features of conservation importance may also be retained as part of targeted fisheries due to its commercial value.					No LSE – pressure not considered relevant to gear types					LSE – see anchored nets/lines and traps for reasoning		
Barrier to species movement	No LSE – feature is not sensitive to this pressure					No LSE - not relevant to this gear type.							
Deoxygenation	No LSE – anthropogenic hypoxia and anoxia are unlikely at this site as discards are not spatially concentrated and the site does not exhibit low-flow conditions. This pressure can also result from the release of deoxygenated ballast water, however all fishing vessels under 45 metres length have solid ballast ²⁷ . No vessels over 45 metres length fish at this site and therefore deoxygenation through ballast water is unlikely.												
Hydrocarbon and PAH contamination	No LSE - deliberate releases are already prohibited. Accidental discharges from fishing vessels leading to significant releases are extremely rare.												
Introduction of light	No LSE - fishing gears do not introduce light and operational / navigational lighting on vessels is unlikely to cause a significant impact due to water depths and temporary nature of light source.												

²⁷ <https://eur-lex.europa.eu/legal-content/en/ALL/?uri=CELEX:32002L0035>

Introduction or spread of invasive species	No LSE – ballast water is the principle source of invasive species in coastal freshwater and marine ecosystems (Drake and Lodge, 2004). Fishing vessels less than 45 metres length must have permanent ballast therefore this vector is not available.	
Litter	No LSE - the strong tidal currents and oceanic swells at the site make it unlikely that lost gear will persist at the site for long enough to cause a significant impact.	
Organic enrichment	No LSE – area is not subject to low-flow or exposed at significant levels to sources of organic enrichment. Any organic material would be quickly washed away as a result of wave and tidal streams.	
Penetration and/or disturbance of the substratum below the surface of the seabed	No LSE – feature is sensitive to but at low-risk from this particular method, which generally does not occur at a level of concern.	LSE – gear is designed to interact with the seabed. Tickler chains on beam trawls, trawl doors on an otter trawl can scour the seabed and scallop dredges cause substantial physical disruption to the seafloor by ploughing sediments and damaging organisms.
Synthetic compound contamination	No LSE - potential source is from vessel hull antifouling treatments. TBT has been banned on vessels under 25 m since 1987 and on all vessels since 2008. Copper wash can enter the marine environment but due to the strong tidal currents at this site, they are not likely to accumulate here.	
Transition elements and organo-metal contamination		
Underwater noise changes	No LSE – feature is not sensitive to this pressure at the benchmark.	
Visual disturbance	No LSE – feature is not sensitive to this pressure at the benchmark.	
Changes in suspended solids (water clarity)	No LSE - not relevant to this gear type.	LSE - results from physical disturbance of the sediment, along with hydrodynamic action caused by the passage of towed gear, leading to entrainment and suspension of the substrate behind and around the gear components. Subtidal mixed sediment as a sub-feature may be more sensitive.
Smothering and siltation rate changes (light)		
Physical change	No LSE – sandbank will remain sandbank despite fishing pressures.	
Nutrient enrichment	No LSE - not relevant to this gear type.	
Introduction of microbial pathogens	No LSE - not relevant as pressure relevant to shellfish production areas only.	

Table 10: Summary of pressures on *S. spinulosa* reef from specific activities taken to Part B.

Potential pressures	Traps	Anchored nets/lines				
	Pots	Gill nets	Set gill nets	Trammel nets	Entangling nets	Demersal longlines
Abrasion/disturbance of the substrate on the surface of the seabed	LSE - The pressure can result from surface disturbance caused by contact between the gear and substrate and occurs during setting of the gear and/or by movement of the gear over the seabed. Such physical disturbance can result in epifauna, especially emergent species such as erect sponges and coral, being dislodged or damaged, although there are limited studies of such effects.					
Removal of non-target species	LSE - bycatch is associated with almost all fishing activities and is related to factors such as the gear type and its design (i.e. its selectivity), the targeted species and effort. Although selective, pots/traps are associated with bycatch, including of non-target crustaceans. Anchored nets/lines can result in the entanglement and bycatch of a range of fauna.					
Barrier to species movement	No LSE - feature is sensitive to this pressure at the conservation advice package benchmark but the pressure does not occur at a level of concern, so the pressure is considered low-risk.					
Deoxygenation	No LSE – Insufficient evidence to assess sensitivity at the benchmark but the pressure does not occur at a level of concern.					
Hydrocarbon and PAH contamination	No LSE - Deliberate releases are prohibited under MARPOL. Accidental releases are infrequent. A sensitivity assessment has not been made for this pressure at the benchmark, but the pressure is low risk for this feature.					
Introduction or spread of invasive species	No LSE – this feature is not sensitive to this pressure at the benchmark.					
Litter	No LSE - the feature is low risk for this pressure as it is not thought to occur at a level of concern. The site is exposed to wave and tidal streams which may prevent litter from accumulating.					
Organic enrichment	No LSE - the feature is not sensitive to this pressure at the benchmark.					
Penetration and/or disturbance of the substratum below the surface of the seabed	No LSE - the pressure is low risk for this feature as reefs exist above the substrate therefore the assessment of penetration and/or disturbance below the surface of the seabed as this would require penetration below the reef itself and is unlikely to occur at a level of concern.					
Synthetic compound contamination	No LSE - Potential source is from vessel hull antifouling treatments. TBT has been banned on vessels under 25 m since 1987 and on all vessels since 2008. Copper wash can enter the marine environment but due to the strong tidal currents at this site, they are not likely to accumulate here.					
Transition elements and organo-metal contamination						

4. Part B Assessment

Part B of this assessment was carried out in a manner that is consistent with the appropriate assessment required by Article 63 of the Habitats Regulations and Article 28 of the Offshore Habitats Regulations.

Table 11 and Table 12 show the fishing activities and pressures included for assessment in part B. Pressures with similar potential impacts to a particular feature have been grouped to save repetition during this assessment.

Table 11: Fishing activities and pressures included for Part B for sandbank.

Aggregated Method	Fishing gear type	Pressures
Traps and Anchored nets/lines	Pots/creels	<ul style="list-style-type: none"> • Abrasion/disturbance of seabed surface substrate • Removal of target species • Removal of non-target species
	Gill nets	
	Trammel nets	
	Entangling nets	
Demersal trawl and Demersal seine	Beam trawl (whitefish)	<ul style="list-style-type: none"> • Abrasion/disturbance of seabed surface substrate • Penetration and/or disturbance of the substrate below the surface of the seabed, including abrasion • Removal of non-target species • Changes in suspended solids • Siltation rate changes
	Beam trawl (shrimp)	
	Beam trawl (pulse/wing)	
	Multi-rig trawls	
	Light otter trawl	
	Pair trawl	
	Scottish/fly seine	
Anchor seine		
Dredges	Scallop dredge	<ul style="list-style-type: none"> • Abrasion/disturbance of seabed surface substrate • Penetration and/or disturbance of the substrate below the surface of the seabed, including abrasion • Removal of target species • Removal of non-target species • Changes in suspended solids • Siltation rate changes
	Seed mussel dredge	

Table 12: Fishing activities and pressures included for Part B for *S. spinulosa* reefs.

Aggregated Method	Fishing gear type	Pressures
Traps and Anchored nets/lines	Pots/creels	<ul style="list-style-type: none"> • Abrasion/disturbance of seabed surface substrate • Removal of non-target species
	Gill nets	
	Trammel nets	
	Entangling nets	
	Longlines (demersal)	

The JNCC and Natural England supplementary conservation tables identify targets for maintaining or recovering features to favourable condition. MMO has identified ‘important’ targets which in this context means only those relating to attributes that are most likely to be impacted by the pressures exerted by the activities being assessed.

Table 13 and Table 14 show which targets were identified as important for each feature. The impacts of pressures on features were assessed against these targets to determine whether the activities causing the pressures are compatible with the site’s conservation objectives (Table 3).

Table 13: Favourable condition targets for identified pressures for sandbanks²⁸.

Feature	Attribute	Target	Relevance/justification
Sandbank	Distribution: presence and spatial distribution of biological communities	Restore the presence and spatial distribution of subtidal sandbank communities.	Pressures identified in Part A Assessment could affect the distribution of communities.
	Extent and distribution	Restore the total extent and spatial distribution of subtidal sandbanks to ensure no loss of integrity, while allowing for natural change and succession.	Pressures identified in Part A Assessment will not affect the extent of the sandbank.
	Structure and function: presence and abundance of key structural and influential species	[Maintain OR Recover OR Restore] the abundance of listed species*, to enable each of them to be a viable component of the habitat.	Key species not identified therefore cannot be assessed.
	Structure: non-native species and pathogens	Restrict the introduction and spread of non-native species and pathogens, and their impacts.	Pressures identified in Part A Assessment will not affect the introduction and spread of non-native species.
	Structure: sediment composition and distribution	Restore the distribution of sediment composition across the feature (and each of its sub-features).	Pressures identified in Part A Assessment could affect the distribution of sediment composition.
	Structure: species composition of component communities	Maintain the species composition of component communities.	Pressures identified in Part A Assessment could affect the species composition.
	Structure: topography	Maintain the presence of topographic features, while allowing for natural responses to hydrodynamic regime, by preventing erosion or deposition through human-induced activity.	Pressures identified in Part A Assessment will not affect the topographic features.
	Structure: volume	Maintain the existing (where no previous evidence exists) or best-known (where some evidence exists) volume of sediment in the sandbank, allowing for natural change.	Pressures identified in Part A Assessment will not affect the volume of sediment.

²⁸ Natural England and JNCC Conservation Advice – [supplementary advice](#)

Supporting processes: energy / exposure	Maintain the natural physical energy resulting from waves, tides and other water flows, so that the exposure does not cause alteration to the biotopes and stability, across the habitat.	Pressures identified in Part A Assessment will not affect the physical energy.
Supporting processes: physico-chemical properties	Maintain the natural physico-chemical properties of the water.	Pressures identified in Part A Assessment will not affect the physio-chemical properties.
Supporting processes: sediment contaminants	Restrict surface sediment contaminant levels to concentrations where they are not adversely impacting the infauna of the feature (and each of its sub-features).	Pressures identified in Part A Assessment will not affect the surface sediment contaminant levels.
Supporting processes: sediment movement and hydrodynamic regime	Maintain all hydrodynamic and physical conditions such that natural water flow and sediment movement are not significantly altered or prevented from responding to changes in environmental conditions.	Pressures identified in Part A Assessment will not affect the hydrodynamic and physical conditions.
Supporting processes: water quality - contaminants	Restrict aqueous contaminants to levels equating to High Status according to Annex VIII and Good Status according to Annex X of the Water Framework Directive, avoiding deterioration from existing levels.	Pressures identified in Part A Assessment will not affect the aqueous contaminants.
Supporting processes: water quality - dissolved oxygen	Maintain the dissolved oxygen (DO) concentration at levels equating to High Ecological Status (specifically ≥ 5.7 mg per litre (at 35 salinity) for 95% of the year), avoiding deterioration from existing levels.	Pressures identified in Part A Assessment will not affect the dissolved oxygen concentration.
Supporting processes: water quality - nutrients	Maintain water quality at mean winter dissolved inorganic nitrogen levels where biological indicators of eutrophication (opportunistic macroalgal and phytoplankton blooms) do not affect the integrity of the site and features, avoiding deterioration from existing levels.	Pressures identified in Part A Assessment will not affect the water quality.
Supporting processes: water quality - turbidity	Maintain natural levels of turbidity (e.g. concentrations of suspended sediment, plankton and other material) across the habitat.	Pressures identified in Part A Assessment will not affect the turbidity levels.

Table 14: Favourable condition targets for identified pressures for *S. spinulosa* reef²⁹.

Feature	Attribute	Target	Relevance/justification
Reef	Distribution: presence and spatial distribution of biological communities	Restore the presence and spatial distribution of reef communities.	Pressures identified in Part A Assessment could affect the distribution of reef communities.
	Extent and distribution	Restore the total extent, spatial distribution and types of reef (and each of its sub-features).	Pressures identified in Part A Assessment could affect the extent and distribution of reef
	Structure and function: presence and abundance of key structural and influential species	[Maintain OR Recover OR Restore] the abundance of listed species*, to enable each of them to be a viable component of the habitat.	Key species not identified therefore cannot be assessed.
	Structure: non-native species and pathogens	Restrict the introduction and spread of non-native species and pathogens, and their impacts.	Pressures identified in Part A Assessment will not affect the introduction and spread of non-native species.
	Structure: species composition of component communities	Restore the species composition of component communities.	Pressures identified in Part A Assessment could affect the composition of component communities.
	Supporting processes: energy / exposure	Restore the natural physical energy resulting from waves, tides and other water flows, so that the exposure does not cause alteration to the biotopes and stability, across the habitat.	Pressures identified in Part A Assessment will not affect the physical energy.
	Supporting processes: physico-chemical properties	Maintain the natural physico-chemical properties of the water.	Pressures identified in Part A Assessment will not affect the physico-chemical properties.
	Supporting processes: sedimentation rate	Maintain the natural rate of sediment deposition.	Pressures identified in Part A Assessment will not affect the rate of sediment deposition.

²⁹ Natural England and JNCC Conservation Advice – [supplementary advice](#)

	Supporting processes: water quality - contaminants	Restrict aqueous contaminants to levels equating to High Status according to Annex VIII and Good Status according to Annex X of the Water Framework Directive, avoiding deterioration from existing levels.	Pressures identified in Part A Assessment will not introduce aqueous contaminants.
	Supporting processes: water quality - dissolved oxygen	Maintain the dissolved oxygen (DO) concentration at levels equating to High Ecological Status (specifically ≥ 5.7 mg per litre (at 35 salinity) for 95% of the year), avoiding deterioration from existing levels.	Pressures identified in Part A Assessment will not affect the dissolved oxygen concentration.
	Supporting processes: water quality - nutrients	Maintain water quality at mean winter dissolved inorganic nitrogen levels where biological indicators of eutrophication (opportunistic macroalgal and phytoplankton blooms) do not affect the integrity of the site and features, avoiding deterioration from existing levels.	Pressures identified in Part A Assessment will not affect the nitrogen levels.
	Supporting processes: water quality - turbidity	Maintain natural levels of turbidity (e.g. concentrations of suspended sediment, plankton and other material) across the habitat.	Pressures identified in Part A Assessment will not affect the levels of turbidity.

4.1 Activity description: traps, anchored nets/lines, demersal trawls, demersal seines, and dredges

4.1.1 Fisheries access/existing management

The vast majority of vessels operating within Inner Dowsing, North Ridge and Race Bank SAC are UK vessels. However, there are a small number of French vessels known to fish within the SAC. EU vessels have access to fish in the site outside of 12 nm, and French vessels occasionally fish in this part of the site.

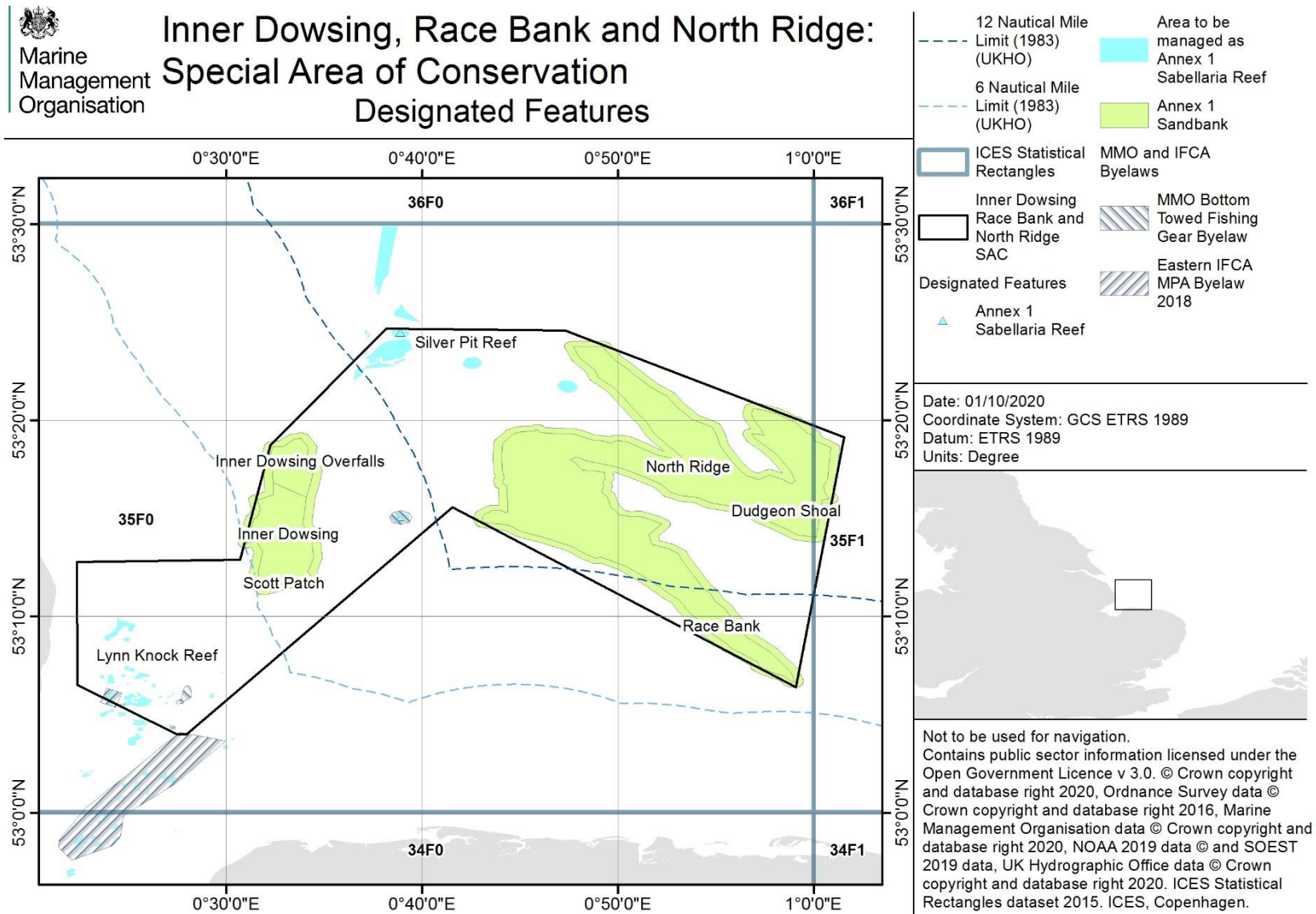
There are technical measures in operation within the Inner Dowsing, Race Bank and North Ridge SAC for stock management³⁰. However, these measures are not designed to achieve the conservation objectives for the features of this site (though they may contribute to the achievement of Good Environmental Status³¹), therefore the impacts from ongoing fishing activities still need to be assessed against the site's conservation objectives and managed where appropriate.

Eastern IFCA Marine Protect Areas Byelaw 2018¹⁶ restricts bottom towed fishing gear over areas of reef within 6 nm, including some of the Lynn Knock reef complex. Within the 0 – 12 nm zone, the use of bottom towed gear over some areas of known reef (Figure 2) have been prohibited by MMO since the introduction of the 'Inner Dowsing, Race Bank and North Ridge European Marine Site (Specified Areas) Bottom Towed Fishing Gear Byelaw¹⁵' in 2013. However, there remains large areas to be managed as reef which do not fall under the 2013 byelaw or the current Eastern IFCA byelaw prohibitions, including the majority of Silver Pit Reef in offshore waters

³⁰ https://ec.europa.eu/fisheries/cfp/fishing_rules/technical_measures_en

³¹ https://www.legislation.gov.uk/ukxi/2010/1627/pdfs/ukxi_20101627_en.pdf

Figure 2: Active byelaws within the SAC from the MMO and Eastern IFCA.



4.1.2 Evidence sources

To determine the levels of fishing activity, the following evidence sources were used:

- VMS data;
- fisheries landings data (logbooks and sales records);
- FisherMap stakeholder mapping;
- Defra commissioned project collating sightings per unit effort (SPUE) data;
- expert opinion from MMO marine officers and IFCA officers;
- fishing industry information;
- IFCA, MMO and Royal Navy sightings; and
- spatial footprint analysis using Pr-values.

Table 15 summarises the description, strengths and limitations of some of the evidence sources used.

Table 15: Summary of generic confidence associated with fishing activity evidence.

Evidence source	Confidence	Description, strengths and limitation
VMS data	High / Moderate	<ul style="list-style-type: none"> • Confidence in VMS is high for describing activity relating to larger vessels (> 12 m). But VMS information was not developed specifically for management of MPAs and does not describe activity in smaller vessels. • There are assumptions in the processing that speed of 0-6 knots is "fishing speed". This may therefore include vessels travelling at these speeds, but which are not fishing, and exclude any fishing taking place above these speeds. Therefore, this may over or under-estimate fishing activity. • VMS records the location, date, time, speed and course of the vessel. Fishing gear information has to be linked to the VMS data itself by either matching its logbook information where possible, using the fleet register which may not be up to date, or through local marine officer knowledge of the said vessel. • VMS data logs vessel movement and thus can act as a good proxy for mobile gear effort. However, it is more challenging to link VMS data to static gear effort (i.e. amount of gear, soak time etc.). • Known guard vessel data have been removed from these data. • Null gear codes are present in the data which may underrepresent fishing fleet. • Non-UK VMS is of lower resolution, presented to just three decimal degrees.
Fisheries landings data	High	<ul style="list-style-type: none"> • Annual data collated and reported to ICES statistical rectangles. • Resolution too low to directly infer landings for MPAs.
FisherMap	Low	<ul style="list-style-type: none"> • The data were collected in 2012 and are relatively dated. • A condition of the research was that only those interviewees who explicitly gave permission for their data to be shared would have their own mapping represented in the final product shared with third parties. This equated to approximately 50% of responses. • The data are self-reported estimates. • The number of skippers who allowed their data to be used represent just over one fifth of the number

		of licensed under 15 m fishing vessels registered in England.
Defra 2015 (MB0117)	Low	<ul style="list-style-type: none"> • Defra Impact Evidence Group report on the feasibility of using a spatial footprint method in appropriate assessments. • Based on recent work to describe fishing activity but is limited by raw data and other limitations highlighted in the report. • Used to inform Pr-value calculations.
Expert judgement	Moderate	<ul style="list-style-type: none"> • This depends on the area, and the knowledge of the area from MMO and IFCA staff.
Fishing industry information	Moderate	<ul style="list-style-type: none"> • Information from the fishing industry regarding intensity of fishing occurring and gear types used in the site. • Depends on the area and the amenability of the local fleet.
Sightings per unit effort data	High	<ul style="list-style-type: none"> • Taken from IFCA, Royal Navy and MMO patrols and targets inspection. • Covers all vessels, not limiting to size class. • Does not account for patrolling/inspection effort.
Pr-values	Moderate / High	<ul style="list-style-type: none"> • Spatial footprint values do not include information for non-VMS vessels. • The methodology used to calculate spatial footprints requires 'matching' of VMS data to specific gear types held on UK or EU fishing fleet registers. This relies on these registers being kept up to date. • There are assumptions in the processing that speed of 0-6 knots is "fishing speed". This may therefore include vessels travelling at these speeds, but which are not fishing, and exclude any fishing taking place above these speeds. Therefore, this may over or under-estimate fishing activity.
MMO catch recording project	Low	<ul style="list-style-type: none"> • Data from the MMO catch recording project for UK vessels under ten metres in length has been used at ICES sub-rectangle level. • This data is only preliminary information, collected from 1 January – 30 November 2020. • It does not include the full fleet and is not being used for compliance measures. There are also known issues with data quality which includes but is not limited to areas of catch, species and gears used.

4.1.3 Fishing gear types used

Fishing activity throughout the site is mostly potting, including whelk as well as crab and lobster pots. Demersal trawls, dredges and anchored nets/lines are also used in the site. In order to bridge the gaps in available data, local knowledge from MMO coastal officers and Eastern IFCA officers has been incorporated into this assessment. The following sections describe the gear types used within the site according to expert opinion. For gear type definitions, please see Annex 2 - Assumptions used to calculate spatial footprint (Pr-values).

Aggregated method: Demersal Trawls

Demersal trawling in the MMO portion of the site consists of whitefish and brown shrimp beam trawling activity. The whitefish beam trawling is limited and carried out by approximately two under 10 m vessels (MMO coastal, *pers comms*). Vessels use 20 mm nets and fish between October and July (MMO coastal, *pers comms*). Kings Lynn and Boston fisheries are largely beam trawlers targeting brown shrimp, however, in 2016 evidence suggests that this appears to be similarly limited, being conducted by seven under 10 metre beam trawlers which occasionally fish within the MMO portion of the site, we have not been able to determine whether these vessels are still operating within the site (MMO coastal 2016, *pers comms*).

VMS data indicates there is some demersal trawling occurring over the sandbank feature within the MMO portion of the site although not in large numbers, most demersal trawling activity occurring in the site is outside of the sandbank and *S. spinulosa* reef features (Figure 3 to Figure 8).

Aggregated method: Dredging

Dredging activity in the MMO portion of the site is limited and consists of shellfish dredging in the form of mussel seed prospecting and some previous hand mechanised dredge activity. The mussel seed prospecting occurs sporadically around the optimal period for relaying mussel seed in late summer (Eastern IFCA, *pers comms*). This fishery is small scale and generally will only occur one week of the year. The mussel seed beds if not fished will naturally be lost through predation or storm damage. This fishery resource has not been found inside 6 nm in this site since 2012 but is known to occur outside of 6 nm. There is some sporadic dredging, including hand mechanised dredges occurring in recent years (2015 – 2017) but this has seen a drastic decline in the area and is not known to occur over the reef or sandbank features.

Aggregated method: Traps

Within the site there are approximately ten under 10 m potting vessels which pot for crab and lobster. Up to 6 of these fish regularly in the site on and around the 6 nm limit adjacent to the Inner Dowsing sandbank. The remaining four vessels fish more occasionally (MMO coastal, *pers comms*). There are seven whelk potters which fish regularly and a further seven whom fish occasionally (MMO coastal, *pers comms*).

A few vessels larger than 12 m use pots in the site however the vast majority of effort comes from just one vessel which is responsible for approximately 77% of the potting fishing VMS reports occurring in the site. VMS data shows significant potting activity in the north of the site from 2015 onwards (Figure 4 to Figure 8).

Eastern IFCA and MMO marine offices indicated that potters do not target areas of *S. spinulosa* reefs. However, VMS data show that areas to be managed as reef are being used for potting (Figure 3 to Figure 8). This may be due to the potential for the 0-6 knot fishing speed metric to overestimate fishing activity where potting gears are concerned. Regardless of fishers' intentions to target areas containing *S. spinulosa* reef, interaction may occur either through storm movement of pots or unintentional interaction due to the ephemeral nature of the feature.

Aggregated method: Anchored nets/lines

The main netting fishery is gill netting which occurs occasionally in winter (depending on weather). In 2016, there were reports of six under 10 metre vessels fishing with long lines which work around the Inner Dowsing sandbank area in winter targeting cod, this has not been verified since (Eastern IFCA, *pers comms*).

4.1.4 VMS Data and Landings Data

VMS and landings data have been included from 2014 to the most up to date information available in order to provide at least five years of data for analysis. Currently, VMS data are available up to and including 2019, and landings data are available up to 2018 for non-UK, EU member state vessels and to 2019 for UK vessels (Table 16 to Table 23).

In 2014, the VMS data shows a substantial amount of netting occurring within the sandbanks outside 12 nm and light demersal trawls across the site (Figure 3). However, there are no observed landings from netting vessels within the site, this is corroborated with advice from Eastern IFCA who are not aware of any netting activity within the site (Eastern IFCA, *pers comms*). From 2015 the VMS data shows a shift in activity from netting to potting with increased activity occurring in the offshore portion of the site adjacent to and over areas of known reef (Figure 4 to Figure 8). The VMS data in 2015 showed the high occurrence of Danish seines near the

sandbank outside 12 nm, from two specific vessels, this is the only year this activity is recorded. There were no landings associated with these vessels, therefore it can be concluded that these vessels are working on the Race Bank Offshore Windfarm. This is supported by advice from Eastern IFCA (EIFCA) is that they are not aware of any Danish seining in the area (Eastern IFCA *pers comms*). From 2016 onwards the main activity in the area is demersal and bottom towed gears. In 2018, towards the north of the site there is a large cluster of “unknown” fishing gears, this is close to a large area of reef on the 12 nm boundary. In 2019, gear types are dominated by demersal gear on or adjacent to the sandbank feature. Potting is also recorded in 2019 on the reef and sandbank features.

The significant portion of VMS data comes from UK vessels which are prevalent in the site from 2014 – 2019 (Figure 9 to Figure 14). However, French vessels were present within Inner Dowsing, Race Bank and North Ridge SAC in 2014 using bottom towed gears near to and occasionally over areas of known *S. spinulosa* reef, although only in small numbers. From 2015 onwards French vessels are present to the north of the site along with Belgian vessels.

Spatially, the majority of VMS fishing activity is focussed in the area inside 6 nm and outside 12 nm with very little activity occurring between 6 and 12 nm. VMS data indicates a decrease in fishing activity over the offshore section of sandbanks since 2014. However, due to the presence of under 12 m vessels fishing in the MMO portion of the site, the VMS data is unlikely to represent the majority of fishing effort occurring at this site.

Table 16 shows landings derived directly from the UK VMS data within the site. A significant proportion of the UK VMS fishing records within the site did not have gear codes or landings data attached, and therefore gear codes were manually assigned to the VMS records. However, assigning gear codes to landing records was not possible and therefore data presented in Table 16 likely represents an underestimate of actual landings from within the site.

The landings data presented in Table 18 and Table 19 were calculated using the proportion of UK VMS fishing pings within ICES rectangles 35F0 and 35F1 that intersect Inner Dowsing, Race Bank and North Ridge SAC (see Table 17 for the proportion of pings from each rectangle that intersect the site). Since there was a large amount of null gear codes within the VMS data at ICES rectangle level, percentages for each year were applied across all gears present within the landings data for over 12 m vessels for each rectangle. When compared with the data derived from the UK VMS fishing records within the site, the landings calculated using this method do not appear to correlate closely. This is likely a result of the UK VMS datasets lacking gear codes and landings, which causes discrepancies when different methods to improve the resolution of this data are employed.

In order to estimate landings by UK vessels without VMS, the proportion of the area of ICES rectangles 35F0 and 35F1 that is occupied by Inner Dowsing, Race Bank and North Ridge SAC was calculated. The sea area of ICES 35F0 is 2710.35 km², of which 831.93 km² is occupied by Inner Dowsing, Race Bank and North Ridge SAC. The sea area of ICES 35F1 is 3714 km², of which 13.41 km² is occupied by Inner Dowsing, Race Bank and North Ridge SAC. Therefore, Inner Dowsing, Race Bank and North Ridge SAC accounts for 30.69% of ICES Rectangle 35F0 and 0.36% of ICES Rectangle 35F1 respectively. These percentages were used to calculate the proportion of the total landings from the rectangles that can be attributed to Inner Dowsing, Race Bank and North Ridge SAC (Table 20 and Table 21). Potting landings appear to be relatively stable across the five-year period, whilst landings from demersal trawls appear to be increasing up to 2018, with a sharp decline in 2019. There are no clear patterns for landings from other gear types. It should be noted that a significant proportion of landings within the data for ICES rectangle 35F0 are attributed to the miscellaneous gear code, which makes it difficult to draw accurate conclusions from the data presented.

Table 22 displays Scientific, Technical and Economic Committee for Fisheries (STECF) non-UK landings for ICES rectangles 35F0 and 35F1 for over and under 12 m vessels. Non-UK VMS landings have been estimated using the proportion of VMS records inside the site compared with the ICES rectangle in which they fall (Table 23). France is the only EU member state to have registered VMS reports using bottom otter trawls within the site, and these are limited in numbers. There were no non-UK VMS fishing reports in the portion of the site that sits within ICES rectangle 35F1. Therefore, a very small proportion of landings from the ICES rectangles can be attributed to non-UK vessels.

Figure 3: 2014 VMS fishing activity by gear type in Inner Dowsing, Race Bank and North Ridge SAC.

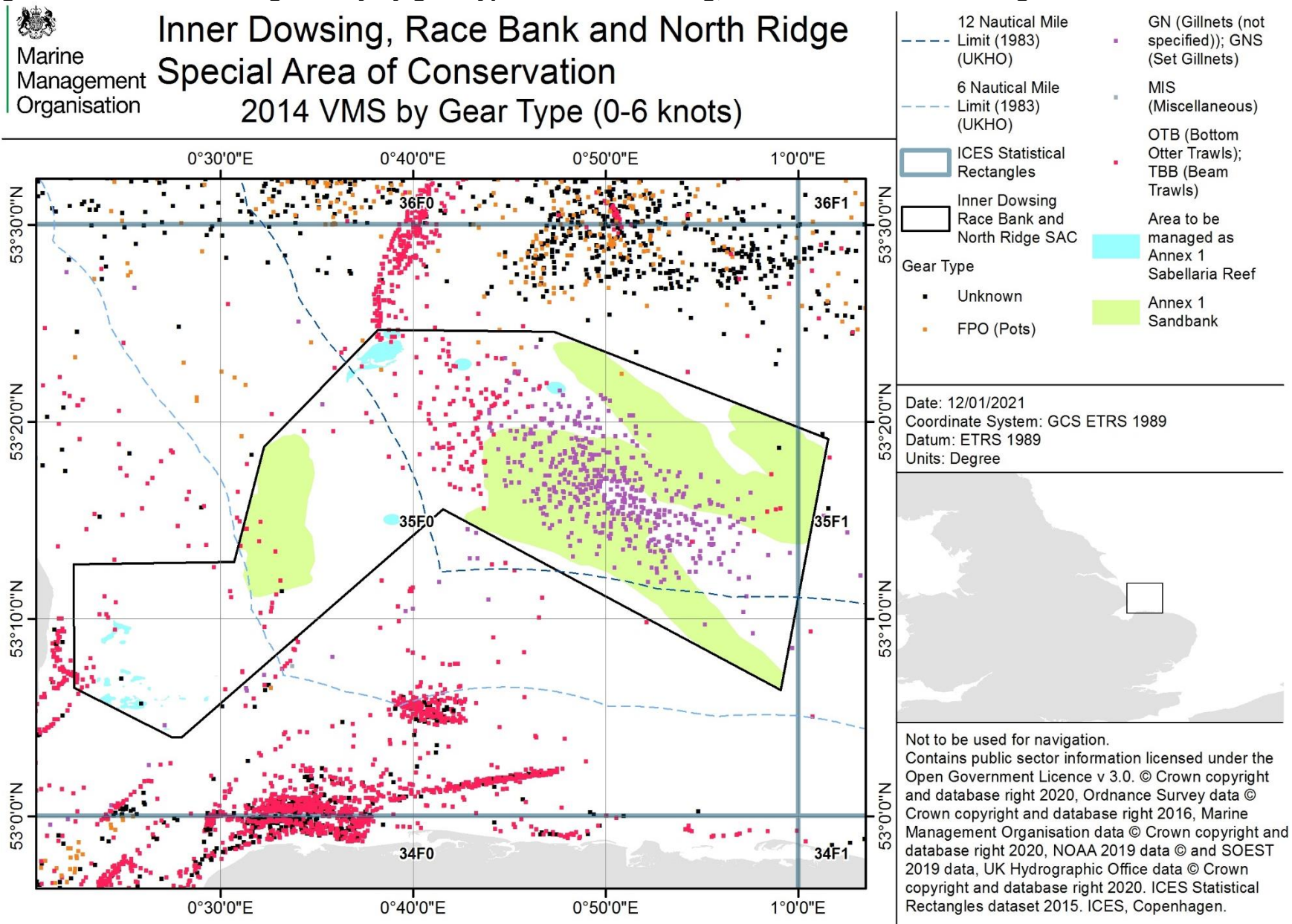


Figure 4: 2015 VMS fishing activity by gear type in Inner Dowsing, Race Bank and North Ridge SAC.

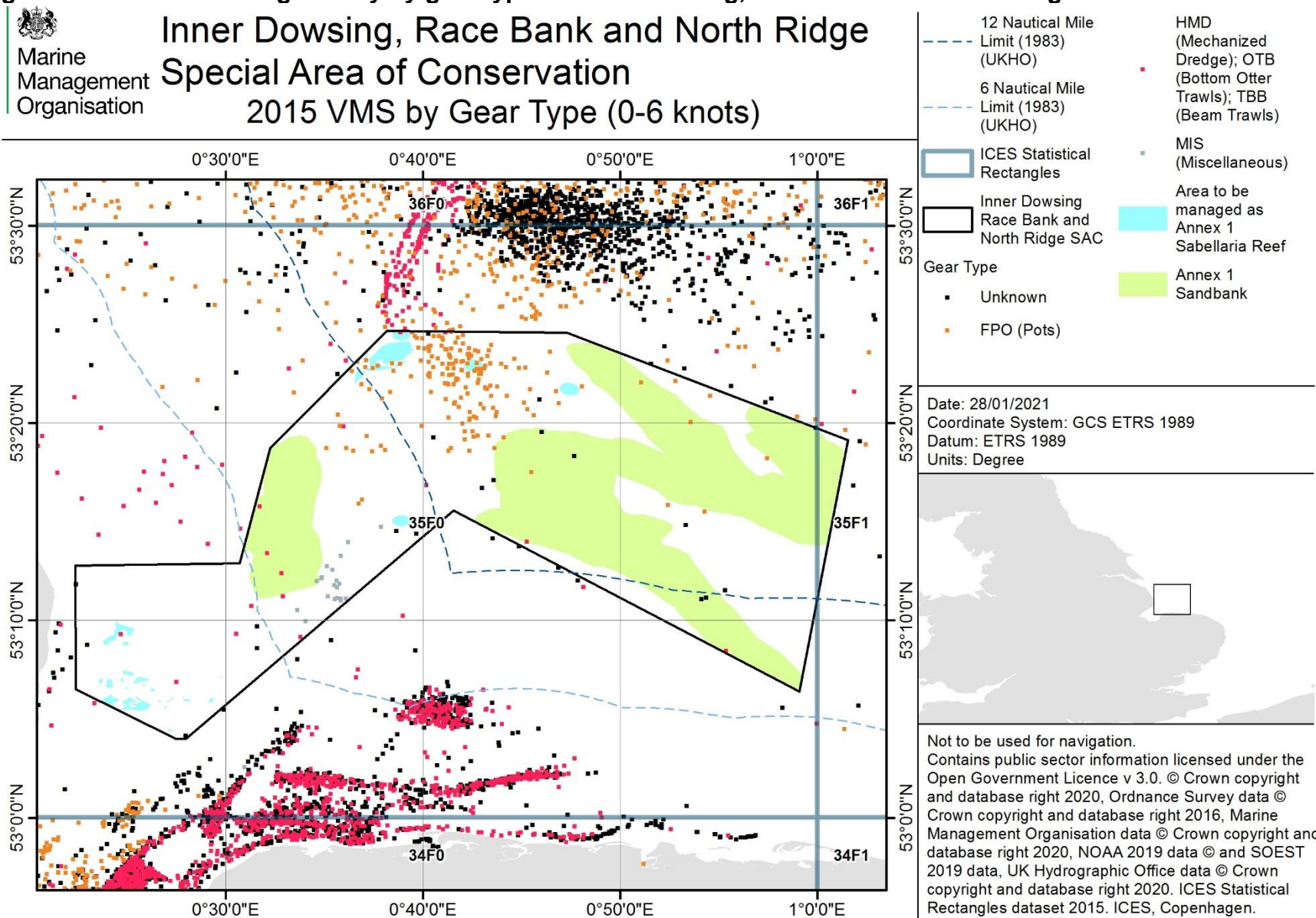


Figure 5: 2016 VMS fishing activity by gear type in Inner Dowsing, Race Bank and North Ridge SAC.

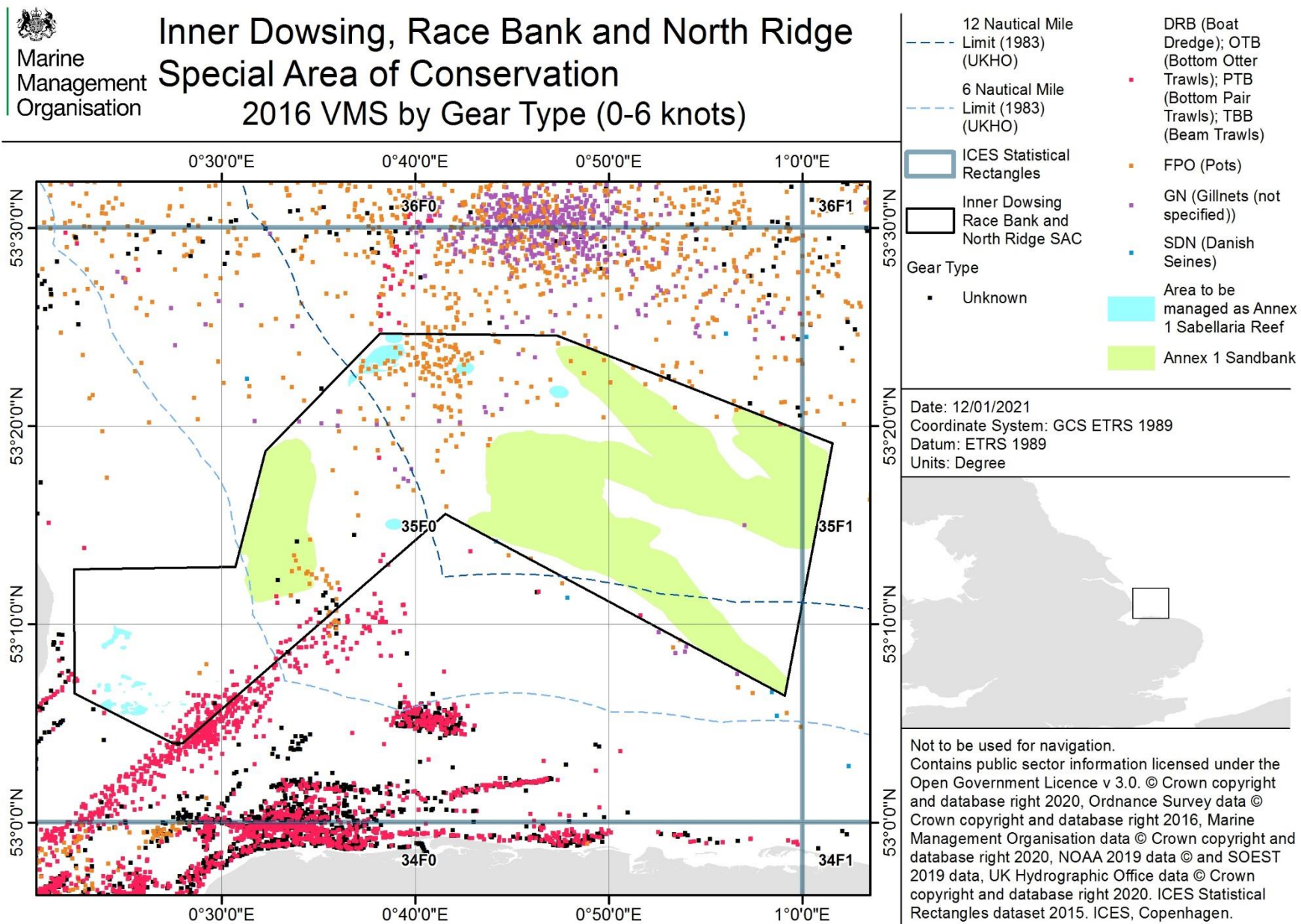


Figure 6: 2017 VMS fishing activity by gear type in Inner Dowsing, Race Bank and North Ridge SAC.

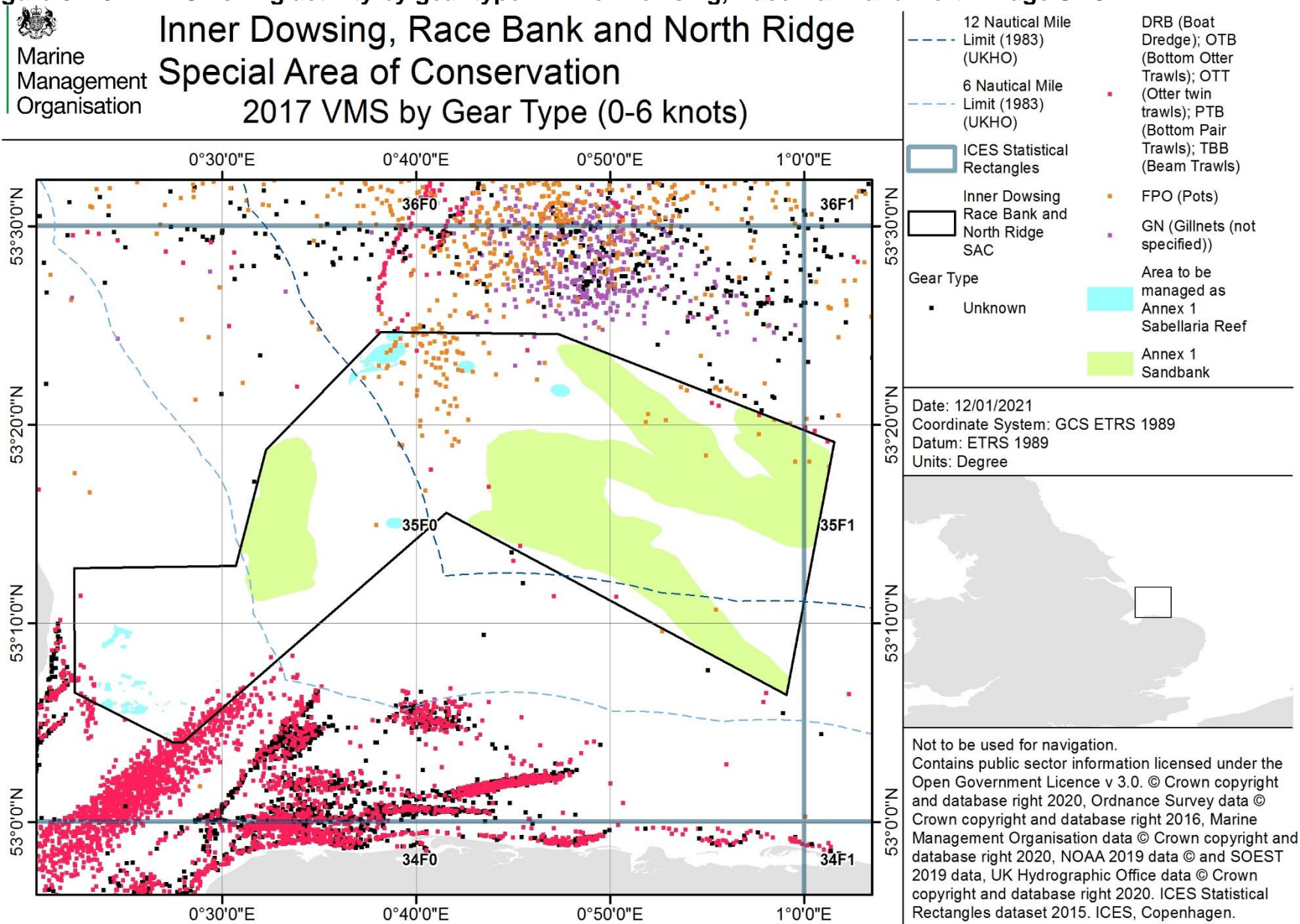


Figure 7: 2018 VMS fishing activity by gear type in Inner Dowsing, Race Bank and North Ridge SAC.

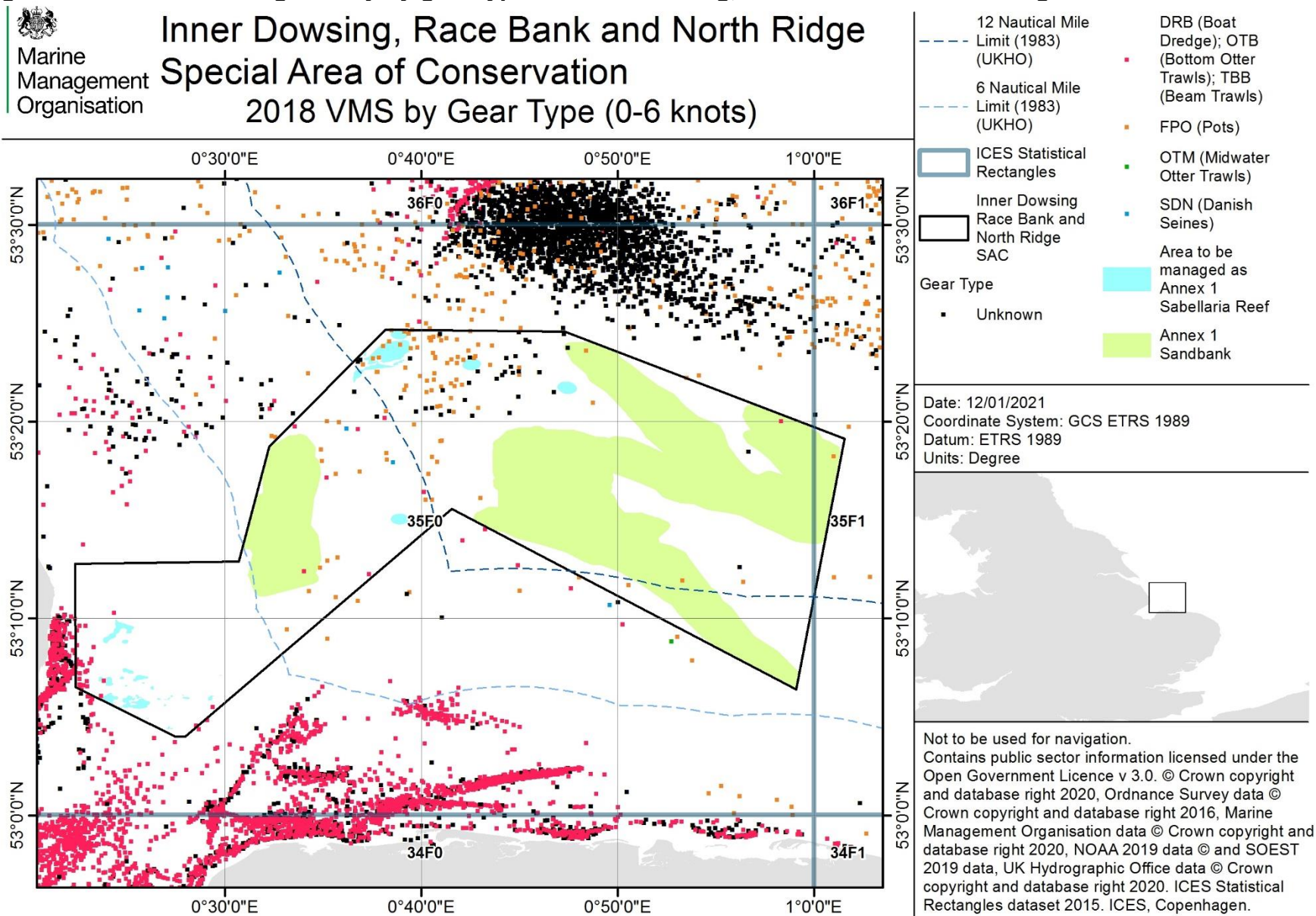


Figure 8: 2019 VMS fishing activity by gear type in Inner Dowsing, Race Bank and North Ridge SAC.

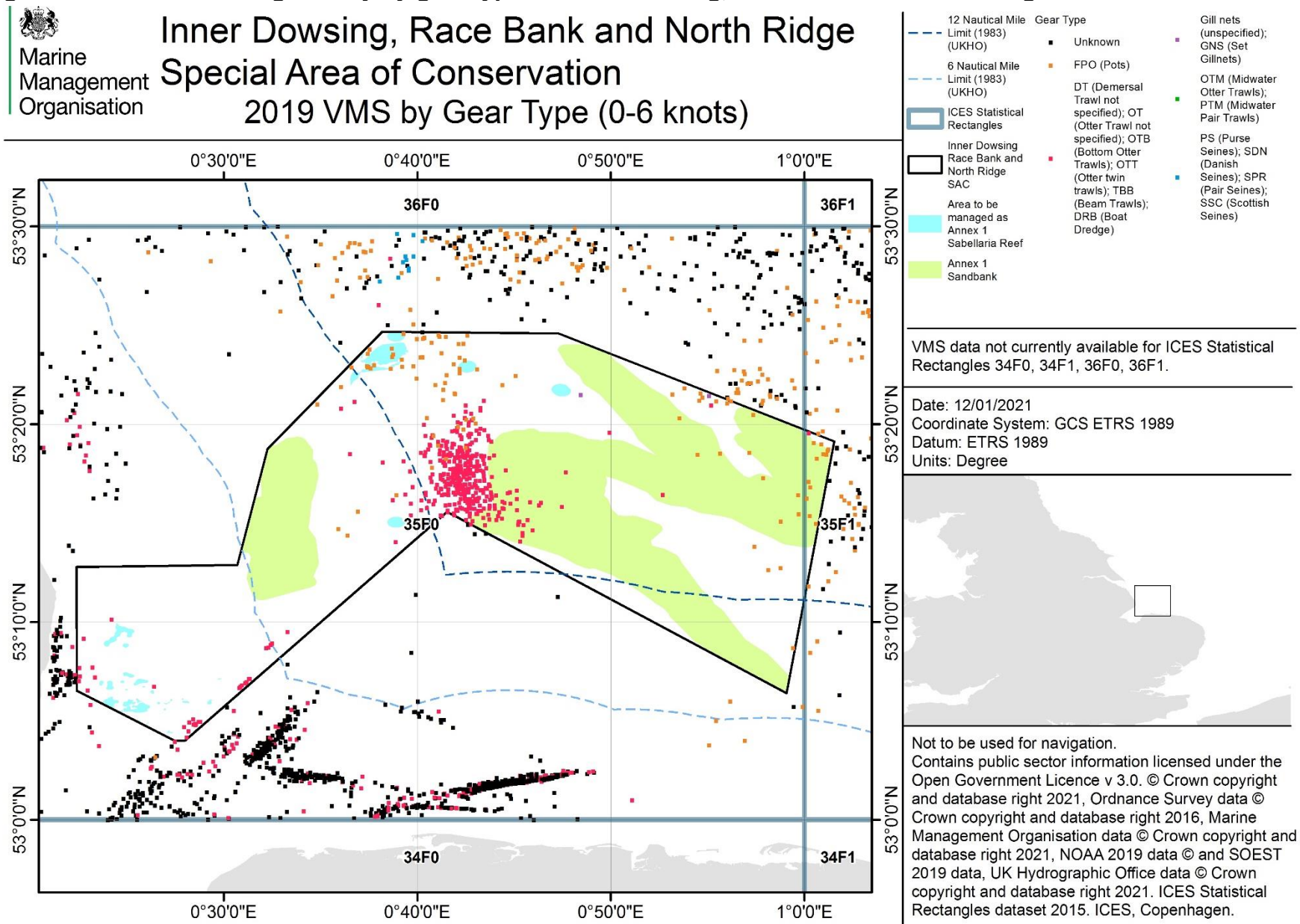


Figure 9: 2014 VMS fishing activity by nationality in Inner Dowsing, Race Bank and North Ridge SAC.

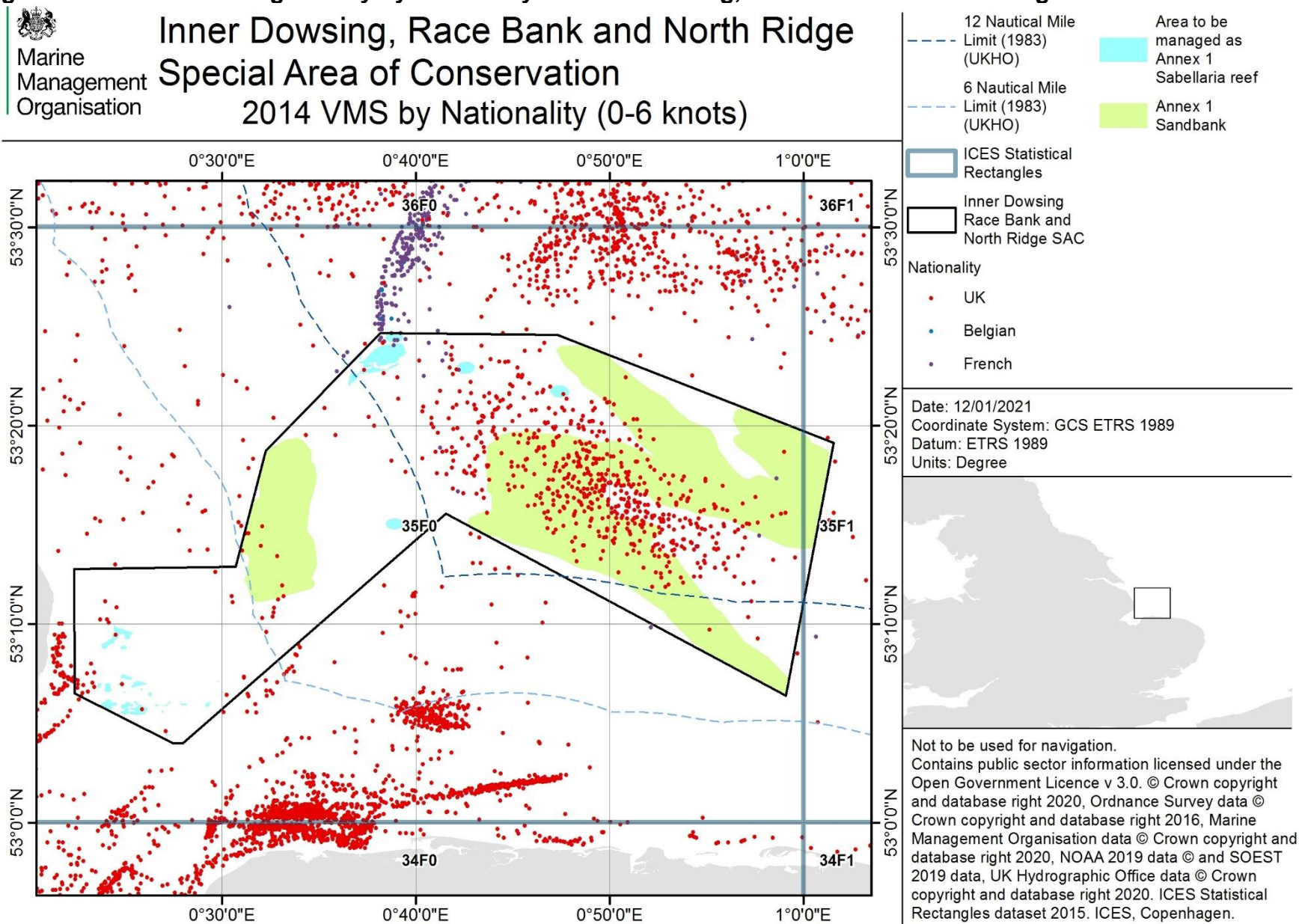


Figure 10: 2015 VMS fishing activity by nationality in Inner Dowsing, Race Bank and North Ridge SAC.

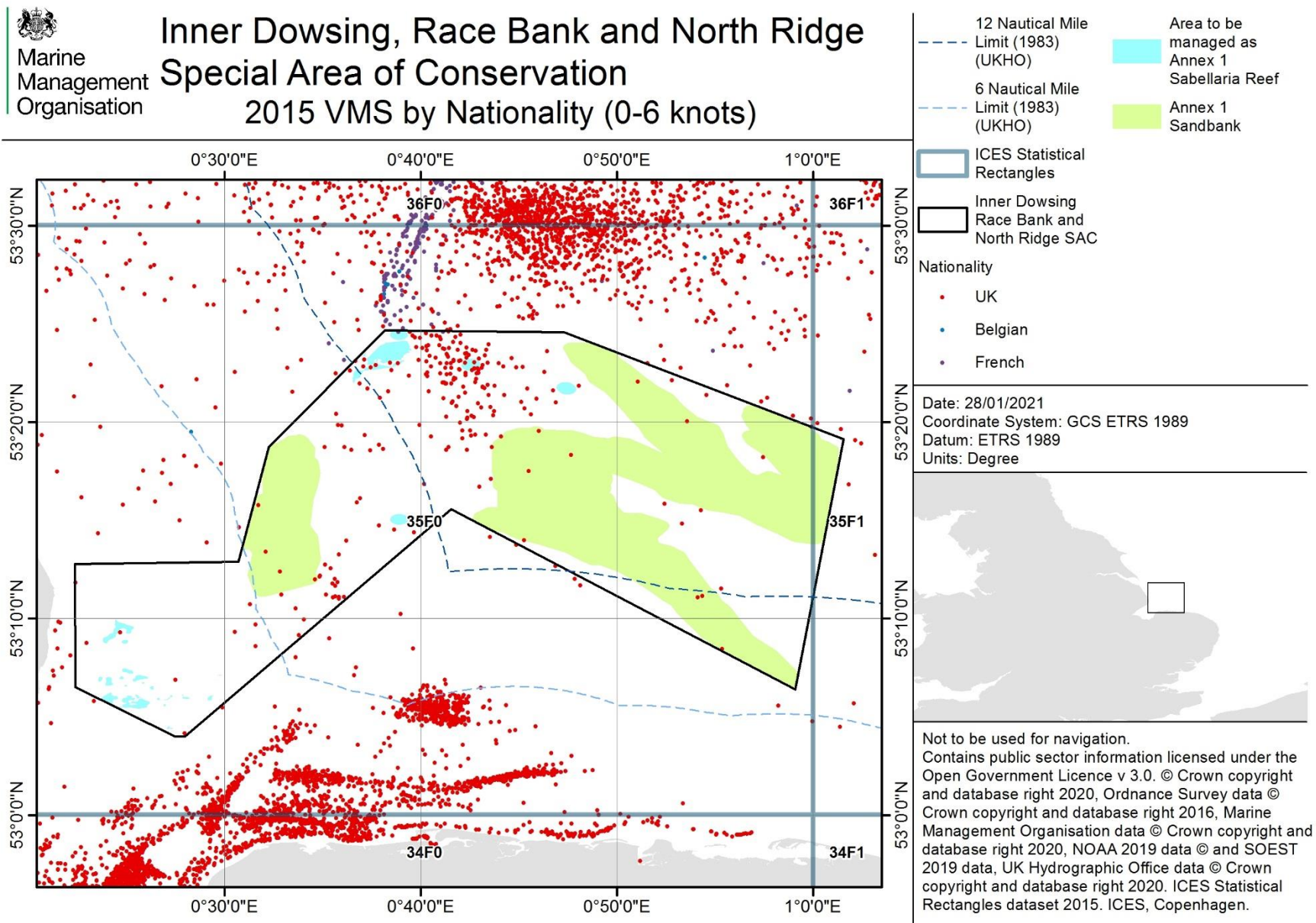


Figure 11: 2016 VMS fishing activity by nationality in Inner Dowsing, Race Bank and North Ridge SAC.

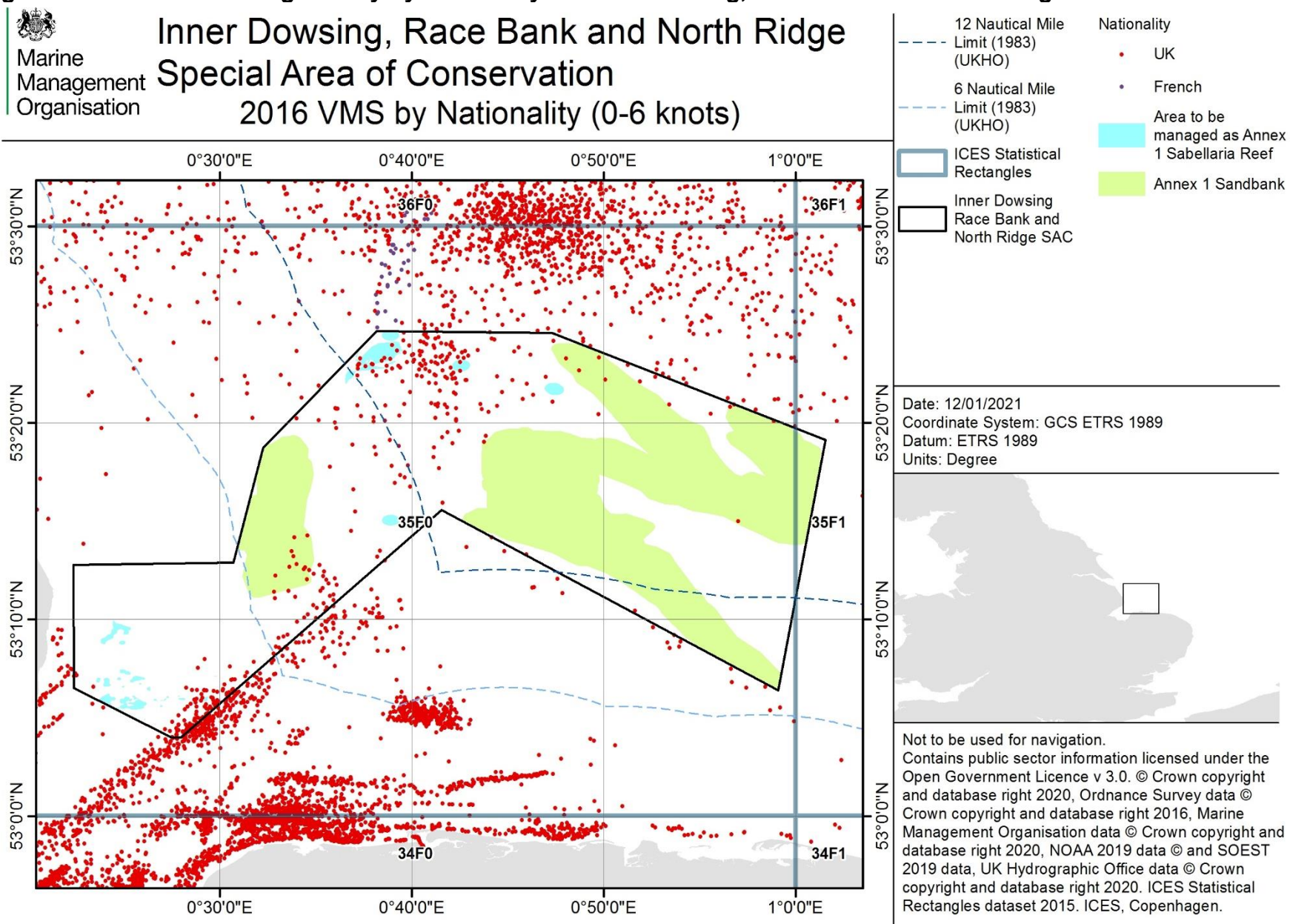


Figure 12: 2017 VMS fishing activity by nationality in Inner Dowsing, Race Bank and North Ridge SAC.

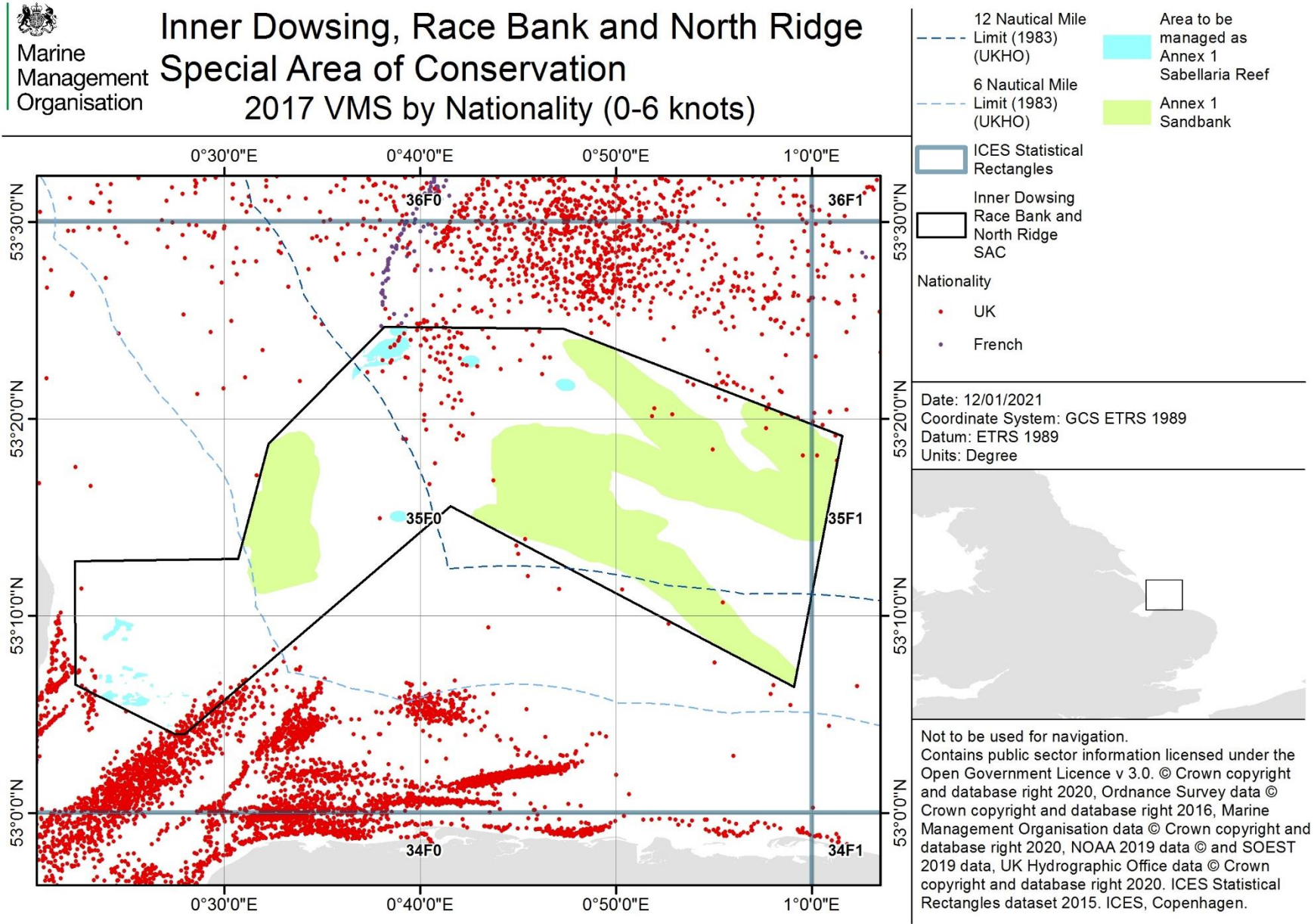


Figure 13: 2018 VMS fishing activity by nationality in Inner Dowsing, Race Bank and North Ridge SAC.

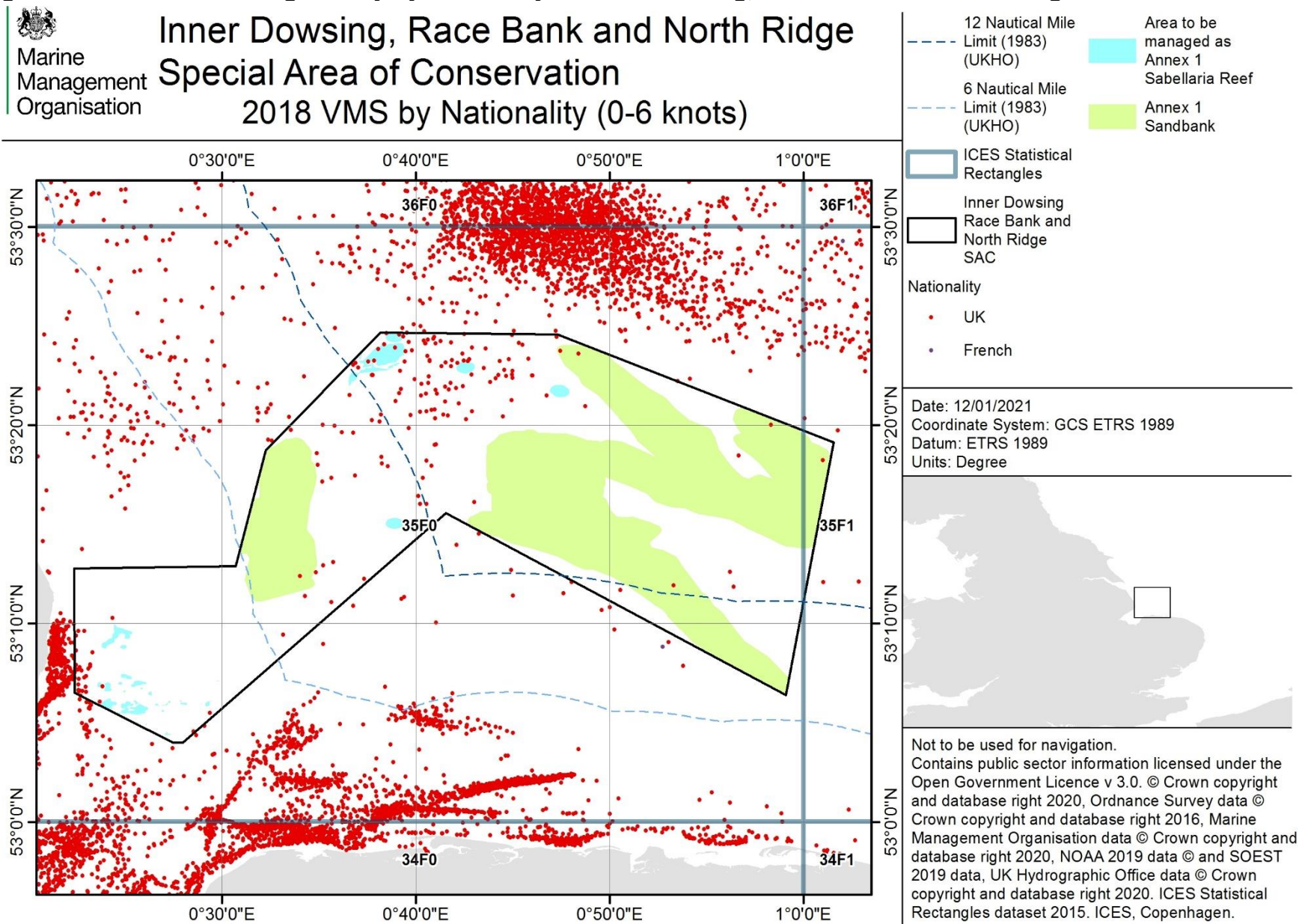


Figure 14: 2019 VMS fishing activity by nationality in Inner Dowsing, Race Bank and North Ridge SAC.

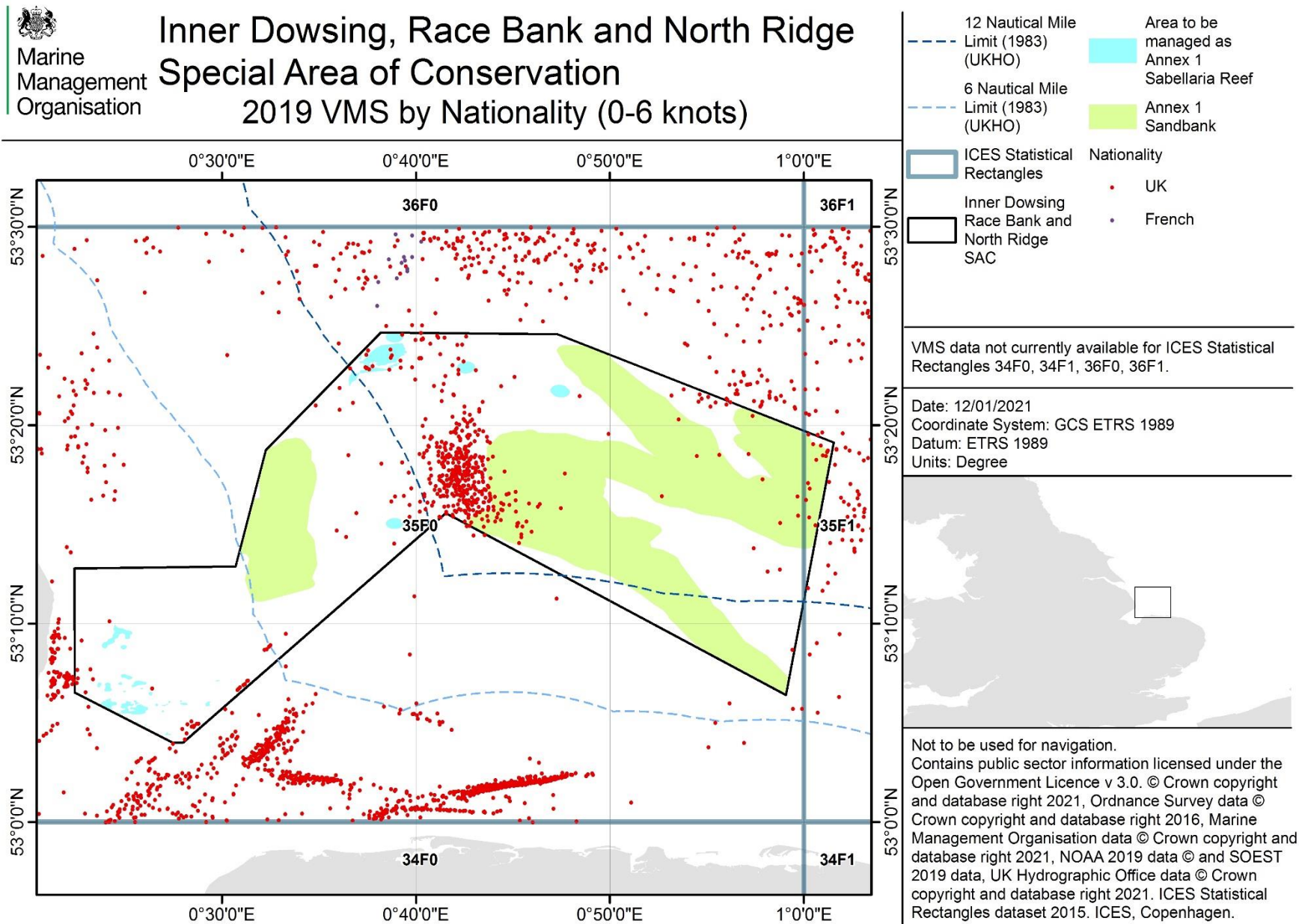


Table 16: Inner Dowsing, Race Bank and North Ridge SAC 2014-2019 demersal gear landings (tonnes) from UK vessels (derived from UK VMS).

Year	Species group	Landings by gear (t)			Total landings (t)
		FPO	OTB	TBB	
2014	All	2.74	-	0.21	2.95
	Crustacea	0.80	-	0.21	1.01
	Mollusc	1.94	-	-	1.94
	Demersal fish	-	-	-	-
2015	All	16.44	-	0.11	16.55
	Crustacea	14.22	-	0.11	14.33
	Mollusc	2.22	-	-	2.22
	Demersal fish	-	-	-	-
2016	All	23.25	0.06	-	23.30
	Crustacea	13.11	0.06	-	13.17
	Mollusc	10.14	-	-	10.14
	Demersal fish	-	-	-	-
2017	All	8.86	0.22	0.36	9.44
	Crustacea	8.58	0.22	0.36	9.16
	Mollusc	0.28	-	-	0.28
	Demersal fish	-	-	-	-
2018	All	11.90	0.47	3.59	15.96
	Crustacea	11.57	0.47	3.59	15.63
	Mollusc	0.33	-	-	0.33
	Demersal fish	-	-	-	-
2019	All	29.15	-	0.57	29.72
	Crustacea	4.35	-	0.56	4.91
	Mollusc	24.80	-	-	24.80
	Demersal fish	-	-	0.01	0.01
2014-2019	All	92.34	0.75	4.84	97.93
	Crustacea	52.63	0.75	4.83	58.21
	Mollusc	39.71	-	-	39.71
	Demersal fish	-	-	0.01	0.01

Table 17: Proportion of VMS reports that intersect Inner Dowsing, Race Bank and North Ridge SAC from those within ICES rectangles 35F0 and 35F1.

	UK VMS fishing pings in ICES 35F0	UK VMS fishing pings in the portion of the site within 35F0	Percentage (%)
2014	2925	668	22.84
2015	2828	299	10.57
2016	3124	329	10.53
2017	4875	181	3.71
2018	5221	169	3.24
2019	2196	497	22.63
	UK VMS fishing pings in ICES 35F1	UK VMS fishing pings in the portion of the site within 35F1	Percentage (%)
2014	738	0	0
2015	815	0	0
2016	3379	0	0
2017	3320	3	0.09
2018	3093	1	0.03
2019	2048	13	0.63

Table 18: Estimated UK VMS Landings for Inner Dowsing, Race Bank and North Ridge SAC based on the proportion of UK VMS fishing pings from ICES Rectangle 35F0 that intersect the site.

Landings by live weight (t) by >12 m vessels from ICES 35F0 and the section of Inner Dowsing, Race Bank and North Ridge SAC within it							
Gear		2014	2015	2016	2017	2018	2019
DRB	35F0	0.00	45.32	0.00	10.01	1.80	0.99
	Site	0.00	4.79	0.00	0.37	0.06	0.22
FPO	35F0	226.02	117.22	59.19	99.56	64.63	107.81
	Site	51.62	12.39	6.23	3.69	2.09	24.40
MIS	35F0	1,181.67	1,455.87	0.00	0.00	0.00	0
	Site	269.89	153.89	0.00	0.00	0.00	0
OTB	35F0	0.28	0.00	4.94	3.50	23.86	4.63
	Site	0.06	0.00	0.52	0.13	0.77	1.05
PTB	35F0	0.00	0.00	0.44	0.80	0.00	0
	Site	0.00	0.00	0.05	0.03	0.00	0
TBB	35F0	170.61	89.87	75.57	109.61	152.55	36.13
	Site	38.97	9.50	7.96	4.07	4.94	7.16

Table 19: Estimated UK VMS Landings for Inner Dowsing, Race Bank and North Ridge SAC based on the proportion of UK VMS fishing pings from ICES Rectangle 35F1 that intersect the site.

Landings by live weight (t) by >12 m vessels from ICES 35F1 and the section of Inner Dowsing, Race Bank and North Ridge SAC within it.							
Gear		2014	2015	2016	2017	2018	2019
DRB	35F1	12.00	5.72	0.19	0.00	0.00	0.72
	Site	0.00	0.00	0.00	0.00	0.00	0.005
FPO	35F1	622.44	245.05	455.97	626.12	794.63	777.56
	Site	0.00	0.00	0.00	0.56	0.24	4.90
MIS	35F1	4.01	2.03	0.00	0.00	0.00	0
	Site	0.00	0.00	0.00	0.00	0.00	0
TBB	35F1	25.94	8.08	1.61	0.00	0.01	0
	Site	0.00	0.00	0.00	0.00	0.00	0

Table 20: Estimated UK Non-VMS Landings for Inner Dowsing, Race Bank and North Ridge SAC based on the proportion of the area (km²) of ICES Rectangle 35F0 that the site occupies.

Landings by live weight (t) in ICES Rectangle 35F0 and the area of Inner Dowsing, Race Bank and North Ridge SAC within it.						
	2014	2015	2016	2017	2018	2019
35F0	1.70	0.00	0.00	73.81	6.00	16.86
Site	0.52	0.00	0.00	22.65	1.84	5.17
35F0	804.88	95.56	201.49	0.00	0.00	0.00
Site	247.02	29.33	61.84	0.00	0.00	0.00
35F0	462.60	385.28	513.64	398.03	473.97	388.72
Site	141.97	118.24	157.64	122.15	145.46	119.30
35F0	0.00	0.18	0.00	0.00	0.00	0.00
Site	0.00	0.05	0.00	0.00	0.00	0.00
35F0	0.19	1.67	0.12	0.00	0.00	0.00
Site	0.06	0.51	0.04	0.00	0.00	0.00
35F0	0.04	0.00	0.00	0.00	0.00	0.00
Site	0.01	0.00	0.00	0.00	0.00	0.00
35F0	7.04	7.62	1.07	0.18	0.42	0.00
Site	2.16	2.34	0.33	0.05	0.13	0.00
35F0	647.14	2174.23	0.00	1773.38	964.05	517.45
Site	198.61	667.27	0.00	544.25	295.87	158.81
35F0	4.10	0.11	0.92	0.00	0.00	0.00
Site	1.26	0.03	0.28	0.00	0.00	0.00
35F0	0.57	0.67	0.68	0.84	2.02	0.58
Site	0.17	0.21	0.21	0.26	0.62	0.18
35F0	0.00	0.13	0.00	0.00	0.00	0.00
Site	0.00	0.04	0.00	0.00	0.00	0.00
35F0	58.29	23.42	74.23	46.72	131.24	5.13
Site	17.89	7.19	22.78	14.34	40.28	1.57

Table 21: Estimated UK Non-VMS Landings for Inner Dowsing, Race Bank and North Ridge SAC based on the proportion of the area (km²) of ICES Rectangle 35F1 that the site occupies.

Landings by live weight (t) in ICES Rectangle 35F1 and the area of Inner Dowsing, Race Bank and North Ridge SAC within it.							
Gear		2014	2015	2016	2017	2018	2019
FPO	35F1	455.59	649.50	659.06	540.41	483.16	331.43
	Site	1.64	2.34	2.37	1.95	1.74	1.19
GN	35F1	0.00	0.00	0.00	0.46	0.18	0.14
	Site	0.00	0.00	0.00	0.002	0.001	0.00
GND	35F1	0.15	0.00	0.00	0.41	0.63	0.32
	Site	0.001	0.00	0.00	0.001	0.002	0.00
LL	35F1	0.26	0.00	1.43	0.02	0.11	0.03
	Site	0.001	0.00	0.01	0.0001	0.0004	0.00
LLS	35F1	0.00	0.00	0.00	0.00	0.21	0.00
	Site	0.00	0.00	0.00	0.00	0.001	0.00
MIS	35F1	1.88	25.46	0.00	0.00	0.00	0.00
	Site	0.01	0.09	0.00	0.00	0.00	0.00
OT	35F1	0.14	0.00	0.00	0.00	0.00	0.00
	Site	0.00	0.00	0.00	0.00	0.00	0.00
PTB	35F1	0.03	0.00	0.00	0.00	0.00	0.00
	Site	0.0001	0.00	0.00	0.00	0.00	0.00
TBB	35F1	0.00	0.00	0.23	0.00	0.14	0.34
	Site	0.00	0.00	0.001	0.00	0.001	0.00

Table 22: Non-UK ICES rectangle landings (t) from STECF data for ICES Rectangles 35F0 and 35F1.

Gear type	ICES rectangle	2014		2015		2016		2017		2018	
		<12	>12	<12	>12	<12	>12	<12	>12	<12	>12
DRB	Total	0	9290.7	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
	35F0	0	0	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
	35F1	0	0	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
DRH	Total	0	4817.67	95.56	0.00	201.49	0.00	0.00	0.00	0.00	0.00
	35F0	0	0	95.56	0.00	201.49	0.00	0.00	0.00	0.00	0.00
	35F1	0	0	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
FPO	Total	0	2522.5	1034.77	362.27	1172.70	515.16	938.44	725.68	957.13	859.26
	35F0	0	0	385.28	117.22	513.64	59.19	398.03	99.56	473.97	64.63
	35F1	0	0	649.50	245.05	659.06	455.97	540.41	626.12	483.16	794.63
GN	Total	0	1261.25	-0.01	0.00	0.00	0.00	0.00	0.00	0.00	0.00
	35F0	0	0	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
	35F1	0	0	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
LLS	Total	0	689.28	7.58	0.00	2.47	2.00	0.20	4.00	0.75	6.00
	35F0	0	0	7.58	0.00	1.04	0.00	0.18	0.00	0.42	0.00
	35F1	0	0	0.00	0.00	1.43	2.00	0.02	4.00	0.32	6.00
NK	Total	0	344.64	2282.76	1374.83	0.00	0.00	1773.38	0.00	964.05	0.00
	35F0	0	0	2257.30	1372.80	0.00	0.00	1773.38	0.00	964.05	0.00
	35F1	0	0	25.46	2.03	0.00	0.00	0.00	0.00	0.00	0.00
OTB	Total	0	172.32	0.78	43.67	1.61	10.87	0.84	7.98	2.02	23.86
	35F0	0	55.01	0.78	43.67	1.61	10.87	0.84	7.98	2.02	23.86
				0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
0.00				0.42	0.00	0.20	0.00	44.93	0.00	0.00	
0.00				0.42	0.00	0.00	0.00	0.00	0.00	0.00	
0.00				0.00	0.00	0.20	0.00	44.93	0.00	0.00	
OTT	0	117.31	0.13	0.00	0.00	0.00	0.00	0.00	0.00	0.00	
			0.13	0.00	0.00	0.00	0.00	0.00	0.00	0.00	
PTT	35F1	0	117.31	0.00	0.00	0.00	0.44	0.00	0.80	0.00	0.00

				0.00	0.00	0.00	0.44	0.00	0.80	0.00	0.00
				0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
TBB				23.42	171.28	74.23	151.34	46.72	181.75	131.38	215.66
				23.42	89.87	74.23	75.57	46.72	109.61	131.24	152.55
				0.00	81.41	0.00	75.77	0.00	72.14	0.14	63.11

Table 23: Estimated Non UK landings (t) from within Inner Dowsing, Race Bank and North Ridge SAC based on the proportion of VMS pings within the site and STECF data. There are no non-UK VMS pings in the portion of the site within 35F1.

Year	Nat	Gear	Pings in SAC	Pings in 35F0	Proportion of pings in SAC	35F0 landings (t)	Estimate of landings from SAC (t)
2014	FRA	OTB	17	177	9.60%	172.32	16.54
2015	FRA	OTB	2	113	1.76%	43.67	0.77
2016	Null	-	0	-	0	10.87	0
2017	FRA	OTB	1	27	3.70%	7.98	0.30
2018	Null	-	0	-	0	23.86	0

4.1.5. Spatial footprint analysis using Pr-values

The spatial footprint analysis used in this assessment is based on a report commissioned by Defra's Impact Evidence Group on the feasibility of using a spatial footprint method in appropriate assessments³² (report reference: MMO1108). Pr-values are derived from VMS data, and therefore only capture vessels with VMS. This analysis is therefore likely to be an underrepresentation of the spatial footprint of fishing gears from all vessels in the SAC. The best available evidence for non-VMS vessels can be found in the subsequent sections.

Analysis was undertaken of the total spatial footprint of fishing gears used each year. The total spatial footprint of a particular gear group was then compared to the total area of the feature, producing a ratio (Pr). A Pr-value of less than one means that the total spatial footprint of the gear in a given year was smaller than the total area of the feature. A Pr-value of more than one means that the total spatial footprint of the gear in a given year was greater than the total area of the feature. Estimates of the Pr-values for the different fishing gears in Inner Dowsing, Race Bank and North Ridge SAC are displayed in Table 24 and Table 25. The assumptions used when calculating footprints are displayed in Annex 2 - Assumptions used to calculate spatial footprint (Pr-values).

The total VMS report area calculates the sum of unique cell areas (0.2025 km²) where VMS reports occur. Over the sandbank feature, this peaked in 2014 for demersal trawls and gillnetting activity while dredging and potting activity peaked in 2016 and 2016 - 17 respectively. The only seine netting activity occurring over sandbanks occurred in 2015.

Due to the relatively small footprint of pots and anchored nets on the seabed and the little fishing activity occurring within the site, the total gear footprint for the sandbank feature, which is the total area impacted by fishing gear, is 0.000002 – 0.000037 km² for pots and 0.0004 – 0.087 km² for nets between 2014 and 2019 (Table 24). The Pr-values, which is the total extent of the sandbank feature (302 km²) impacted by the gears are <0.0000001 for pots and 0.000001 – 0.0003 for nets (Table 24).

Demersal trawls have a larger footprint on the seabed which is reflected in the larger figures in Table 24 compared with potting and netting despite less activity occurring. Demersal trawls combined, result in a total gear footprint for the sandbank feature of 0.12 – 1.29 km² and Pr-values of 0.0004 – 0.004 between 2014 and 2019 (Table 24).

Similarly, low levels of VMS activity from dredges result in a Pr-value of 0.0029-0.001 between 2014 and 2019. Seine gears have a Pr-values over the sandbank feature of 0.0007 in 2015 (Table 24).

³²http://randd.defra.gov.uk/Document.aspx?Document=12955_MMO1108SpatialFootprintAnalysisReport-FINAL.pdf, MARG Ltd in association with Envision Mapping Ltd, 2015

Only bottom otter trawl (2014 - 2018), pots (2015 - 2019) and gillnets (2014 only) VMS activity has occurred over the reef feature. As noted previously, the small footprint of gillnets and pots means that despite reasonable levels of activity over the reef feature, only small areas are impacted, with Pr-values not exceeding 0.000002 for pots and 0.00003 for gillnets.

Regarding the *S. spinulosa* reef feature, the highest impacts to reef are from demersal trawling activity, specifically bottom otter trawls. Total gear footprint and Pr-values over the reef feature peaked in 2014 (0.14 km² and 0.012 respectively) then decreased to 0.0023 km² and 0.002 in 2018 (Table 25).

It is likely that certain parts of the site are subject to more frequent levels of fishing (Figure 15 and Figure 16). It also should be noted that this only represents the activity of vessels with VMS which are likely to represent a small proportion of the fishing activity at the site. Pr-values must also be treated with a high degree of caution as they rely on numerous assumptions about size and behaviour of gear, and frequency of use.

Table 24: Spatial footprint (km²) values for VMS vessels on sandbank.

Year	2014	2015	2016	2017	2018	2019
OTB						
Total gear footprint	0.422	0.328	0.117	0.235	0.023	0.962
Pr-value	0.001	0.001	3.89E-04	0.001	7.78E-05	0.003
Pr value %	0.14	0.109	0.039	0.078	0.008	0.319
OTT						
Total gear footprint	0	0	0	0.046	0	0
Pr-value	0	0	0	1.51E-04	0	0
Pr value %	0	0	0	0.015	0	0
TBB						
Total gear footprint	0.867	0.217	0	0	0.433	0
Pr-value	0.003	0.001	0	0	0.001	0
Pr value %	0.287	0.072	0	0	0.144	0
DRB						
Total gear footprint	0	0	0.177	0.089	0.354	0
Pr-value	0	0	0.001	2.93E-04	0.001	0
Pr value %	0	0	0.059	0.029	0.117	0
BTG Total						
Total gear footprint	1.289	0.545	0.294	0.37	0.81	0.962
Pr-value	4.00E-03	2.00E-03	1.39E-03	1.44E-03	2.08E-03	0.003
Pr value %	0.427	0.181	0.098	0.122	0.269	0.319
FPO						
Total gear footprint	2.03E-06	1.82E-05	3.04E-05	3.24E-05	1.62E-05	3.65E-05
Pr-value	6.71E-09	6.04E-08	1.01E-07	1.07E-07	5.37E-08	1.21E-07
Pr value %	6.71E-07	6.04E-06	1.01E-05	1.07E-05	5.37E-06	1.21E-05
GN						
Total gear footprint	0.087	0	2.00E-03	3.66E-04	0	0
Pr-value	2.88E-04	0	7.28E-06	1.21E-06	0	0
Pr value %	0.029	0	1.00E-03	1.21E-04	0	0
GNS						
Total gear footprint	3.66E-04	0	0	0	0	0
Pr-value	1.21E-06	0	0	0	0	0
Pr value %	1.21E-04	0	0	0	0	0
SDN						
Total gear footprint	0	0.022	0	0	0	0
Pr-value	0	7.18E-05	0	0	0	0
Pr value %	0	7.00E-03	0	0	0	0
Static total						
Total gear footprint	0.087	0.022	2.03E-03	3.98E-04	1.62E-05	3.65E-05
Pr-value	2.89E-04	7.19E-05	7.38E-06	1.32E-06	5.37E-08	1.21E-07
Pr value %	0.029	7.01E-03	1.01E-03	1.32E-04	5.37E-06	1.21E-05
Static and BTG total						
Total gear footprint	1.376	0.567	0.296	0.370	0.810	0.962
Pr-value	0.004	0.002	0.001	0.001	0.002	0.003
Pr value %	0.456	0.188	0.099	0.122	0.269	0.319

Table 25: Spatial footprint (km²) values for VMS vessels on *S. spinulosa* reef.

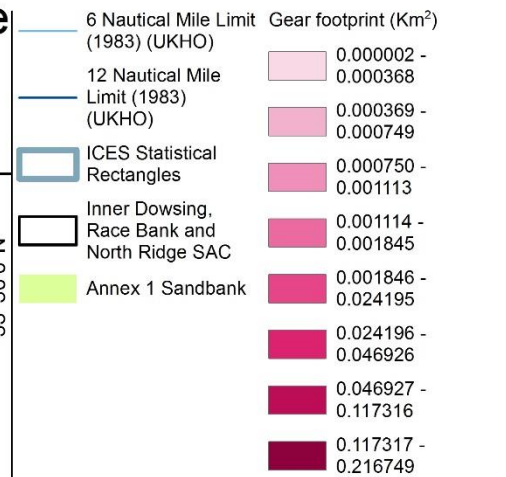
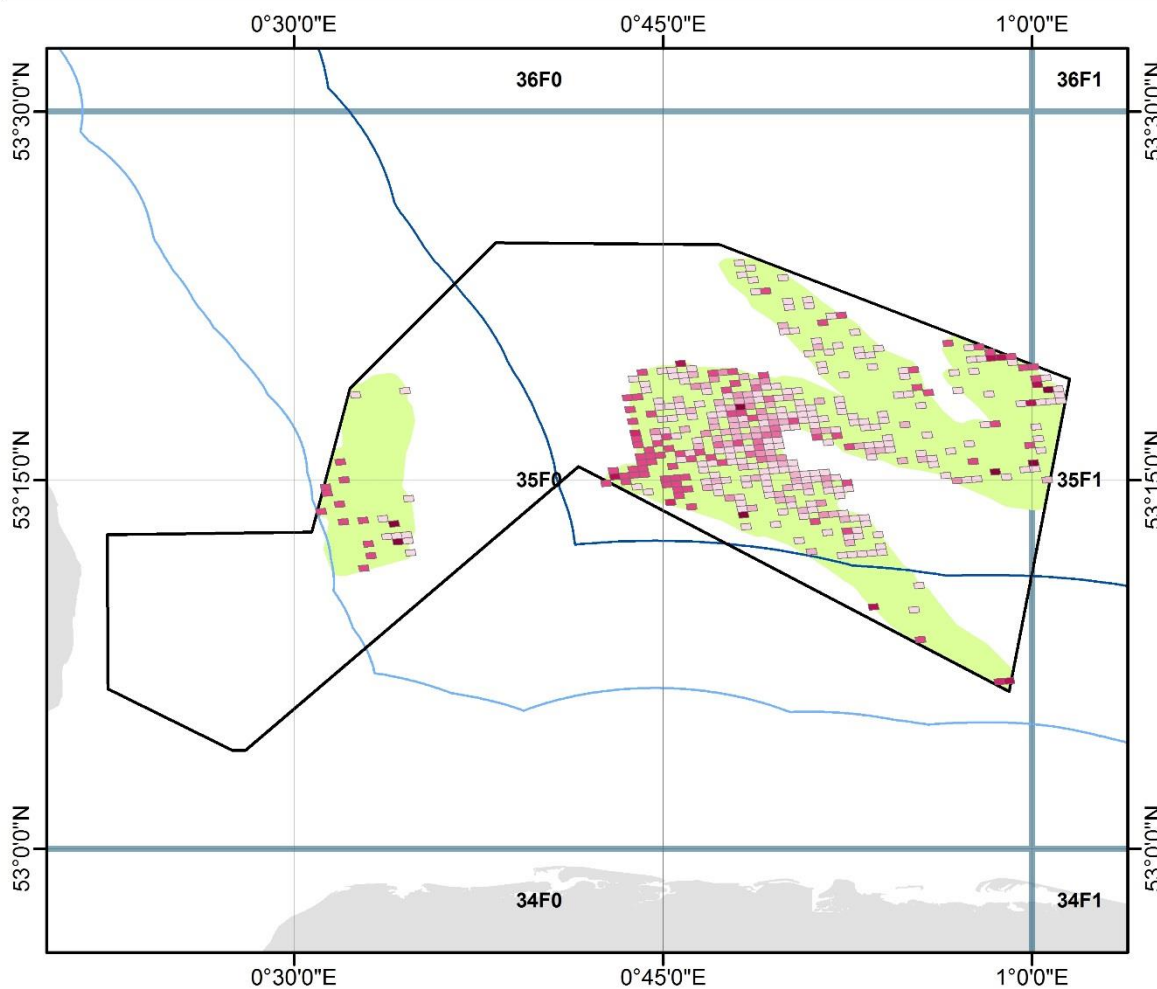
Year	2014	2015	2016	2017	2018	2019
OTB						
Total gear footprint	0.141	0	0.023	0.047	0.023	0
Pr-value	0.012	0	0.002	0.004	0.002	0
Pr value %	1.165	0	0.194	0.388	0.194	0
Total BTG						
Total gear footprint	0.141	0	0.023	0.047	0.023	0
Pr-value	0.012	0	0.002	0.004	0.002	0
Pr value %	1.165	0	0.194	0.388	0.194	0
FPO						
Total gear footprint	0	1.82E-05	1.82E-05	2.03E-05	1.01E-05	1.42E-05
Pr-value	0	1.51E-06	1.51E-06	1.68E-06	8.38E-07	1.17E-06
Pr value %	0	1.51E-04	1.51E-04	1.68E-04	8.38E-05	1.17E-04
GN						
Total gear footprint	3.66E-04	0	0	0	0	0
Pr-value	3.03E-05	0	0	0	0	0
Pr value %	0.003	0	0	0	0	0
Total static gear						
Total gear footprint	3.66E-04	1.82E-05	1.82E-05	2.03E-05	1.01E-05	1.42E-05
Pr-value	3.03E-05	1.51E-06	1.51E-06	1.68E-06	8.38E-07	1.17E-06
Pr value %	0.003	1.51E-04	1.51E-04	1.68E-04	8.38E-05	1.17E-04
Total BTG and static gear						
Total gear footprint	0.141	1.82E-05	0.023	0.047	0.023	1.42E-05
Pr-value	0.012	1.51E-06	0.002	0.004	0.002	1.17E-06
Pr value %	1.168	1.51E-04	0.194	0.388	0.194	1.17E-04

Figure 15: Spatial footprint analysis of all fishing gears over sandbank within Inner Dowsing, Race Bank and North Ridge SAC.



Inner Dowsing, Race Bank and North Ridge Special Area of Conservation

All Gear Sandbank Footprint (2014 to 2019)

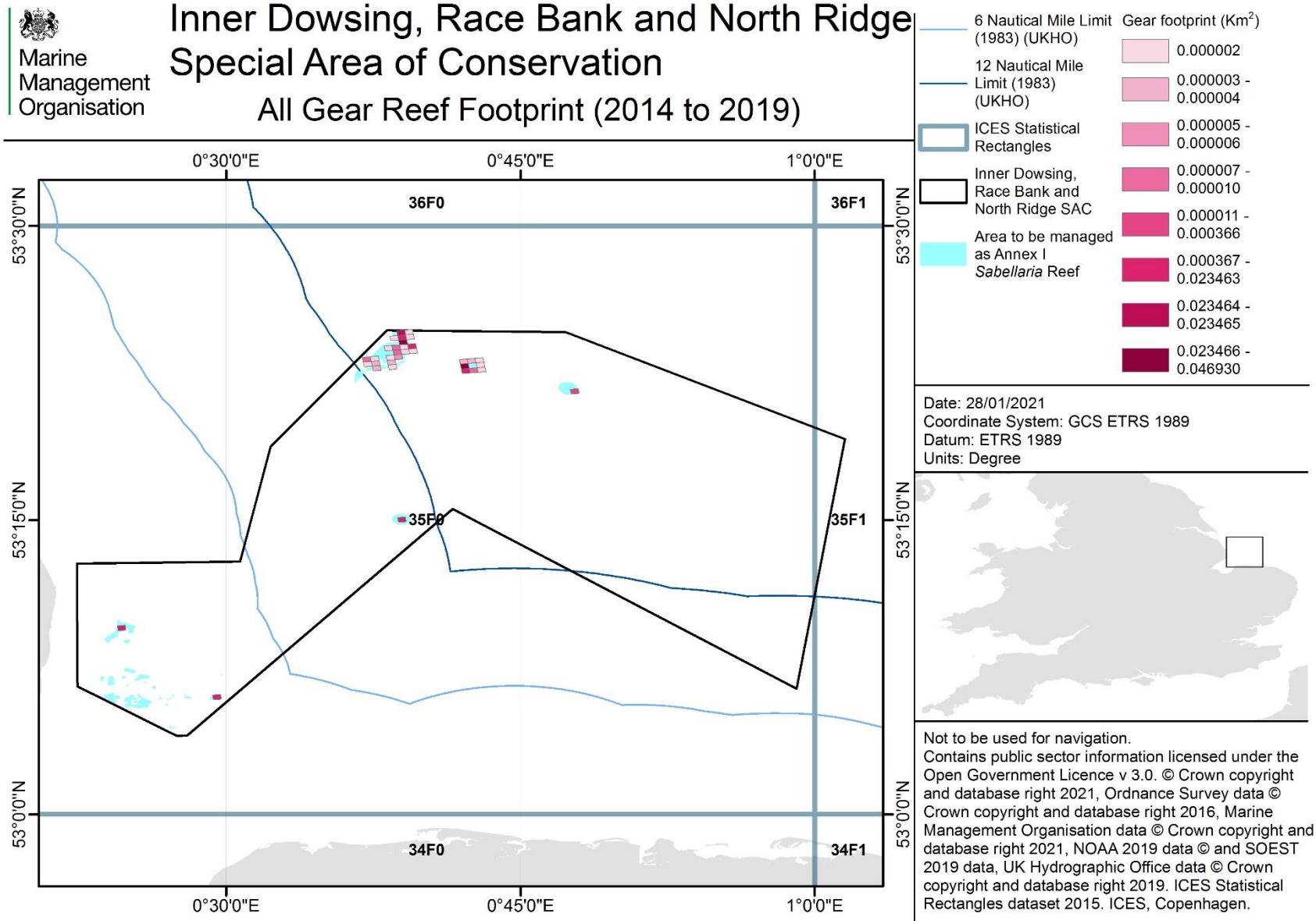


Date: 28/01/2021
 Coordinate System: GCS ETRS 1989
 Datum: ETRS 1989
 Units: Degree



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Figure 16: Spatial footprint analysis of all fishing gears over reef within Inner Dowsing, Race Bank and North Ridge SAC.



4.1.6 FisherMap

In 2012, the FisherMap project aimed to map the activities of the commercial fishing fleet. Interviews were conducted with approximately 1000 skippers of the under 15 m fishing fleet. Of those interviewed, 594 gave their permission for their data to be shared with third parties.

FisherMap data represents the number of fishers that indicated they fish within the site boundary over a year's fishing activity (collected from a series of monthly totals of vessel numbers per grid cell) using a particular gear type (des Clers *et al.*, 2008; des Clers, 2010). Inner Dowsing, Race Bank, North Ridge boundary has been overlaid with this data (Figure 17 to Figure 21).

FisherMap data indicated that the majority of the demersal towed gear fishing activity within the SAC by the under 15 m fleet took place inshore of 6 nm limit and an area to the southeast of the Inner Dowsing sandbank (see Figure 1 for labels of named sandbanks and reefs).

There was some demersal towed gear activity on the sandbanks between the 6 nm to 12 nm; the number of demersal towed gear fishing vessel visits for the various sandbank areas is shown in Table 26. Table 27 shows the number of static gear visits over the areas of *S. spinulosa* reef.

Table 26: Number of under 15 metre fishing vessel visits per year over each of the sandbank areas by gear type³³.

Sandbank	Number of fishing vessel visits per year by gear type				
	Bottom towed gear	Dredges	Pots	Lines	Nets
Inner Dowsing Overfalls	21-60	11-30	41-80	11-60	-
Inner Dowsing	41-150	31-100	41-70	11-60	-
Scott Patch	41-100	31-100	31-70	11-40	-
Race Bank	51-60	-	61-150	11-30	-
North Ridge	1-30	1-10	41-80	11-20	-
Dudgeon Shoal	11-40	-	51-70	11-20	-

Table 27: Number of under 15 metre fishing vessel visits per year over each of the reef areas by gear type¹³.

<i>S. spinulosa</i> reef	Number of fishing vessel visits per year by gear type		
	Pots	Lines	Nets
Silver Pit Reef	21-80	11-20	-

³³ As these areas fall over a number of grid cells the highest number of visits has been recorded.

Figure 17: Inner Dowsing, Race Bank and North Ridge SAC FisherMap (2012) - Bottom Towed Gear.

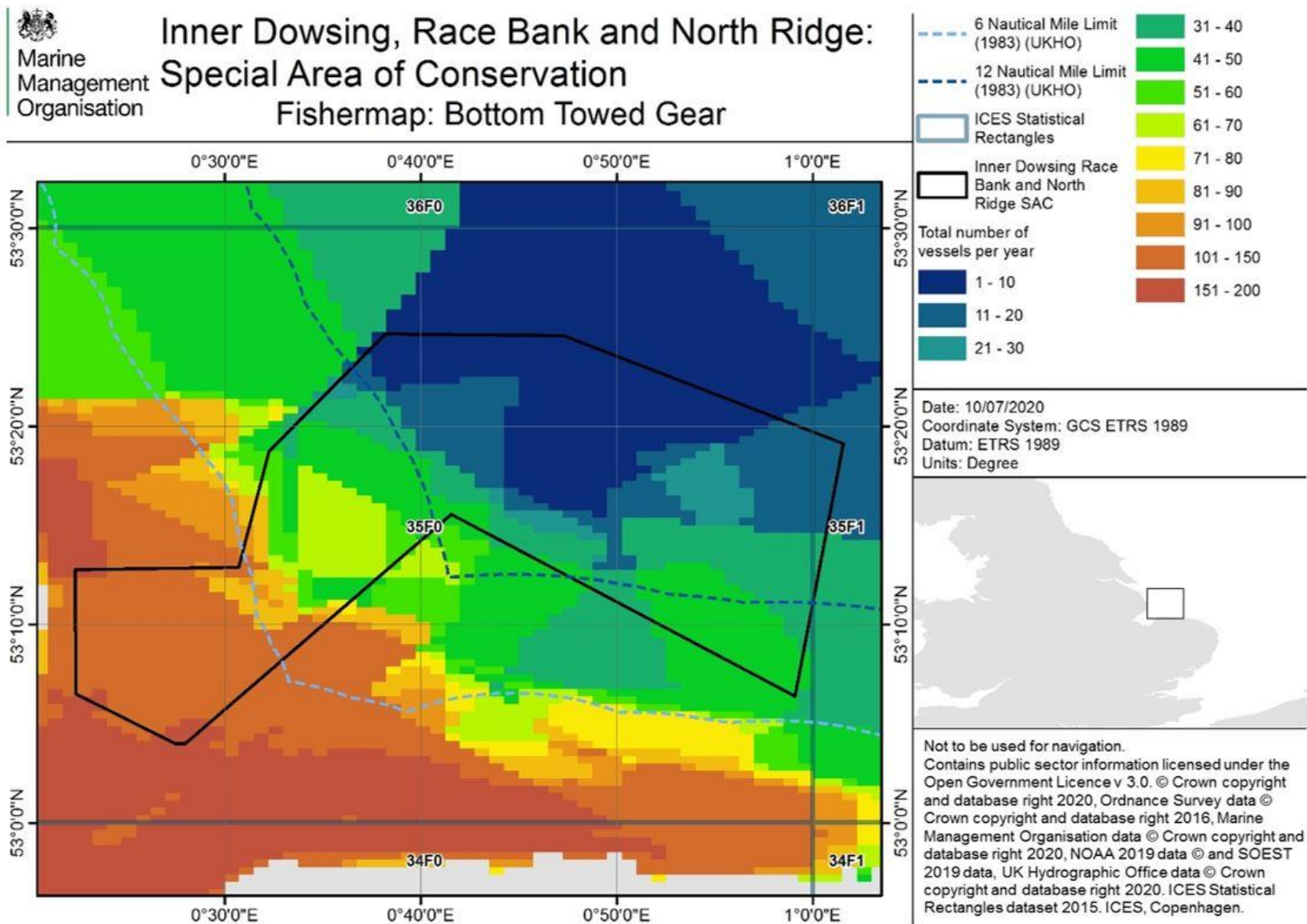


Figure 18: Inner Dowsing, Race Bank and North Ridge SAC FisherMap (2012) – Dredges.

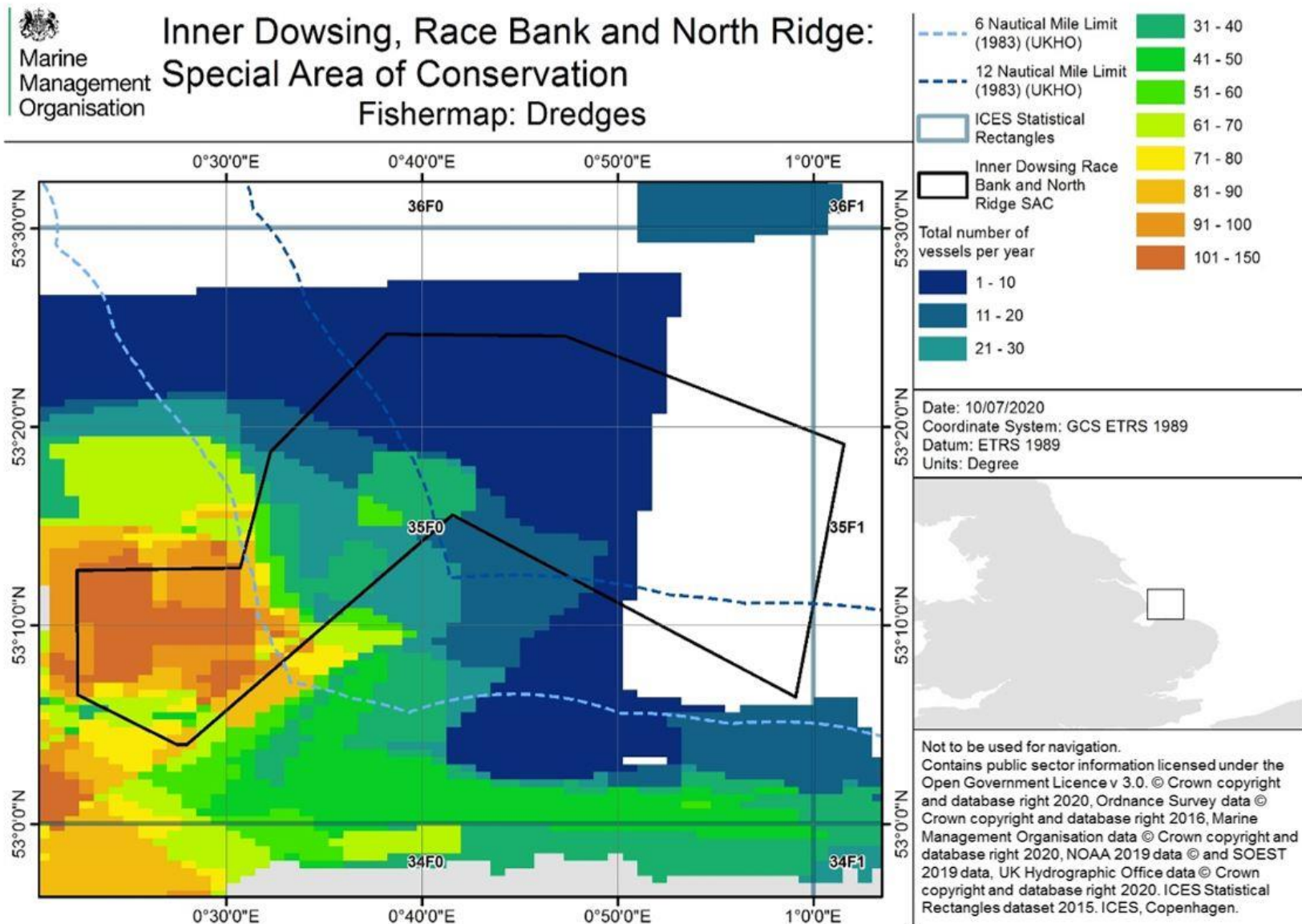


Figure 19: Inner Dowsing, Race Bank and North Ridge SAC FisherMap (2012) – Pots.

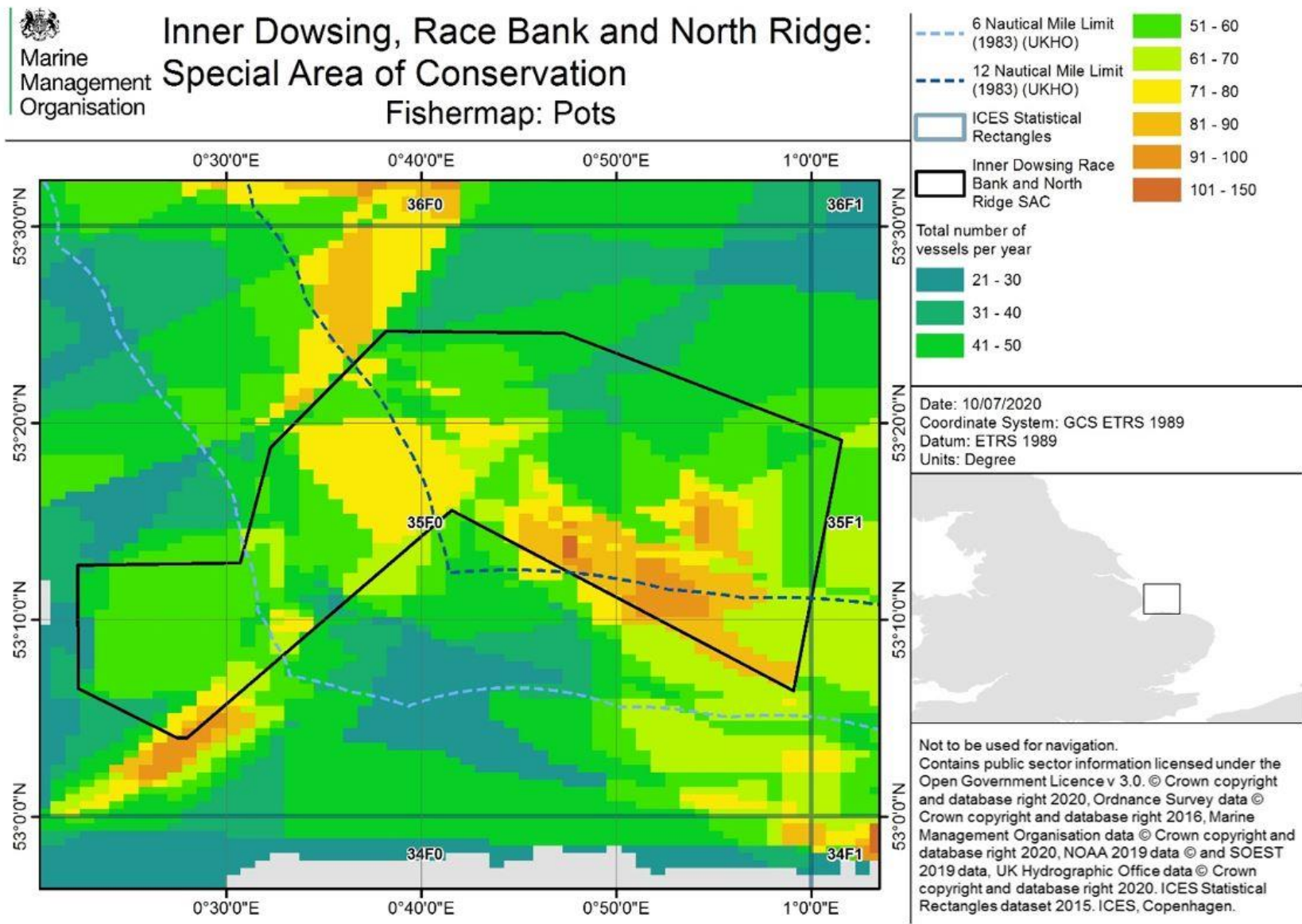


Figure 20: Inner Dowsing, Race Bank and North Ridge SAC FisherMap (2012) – Nets.

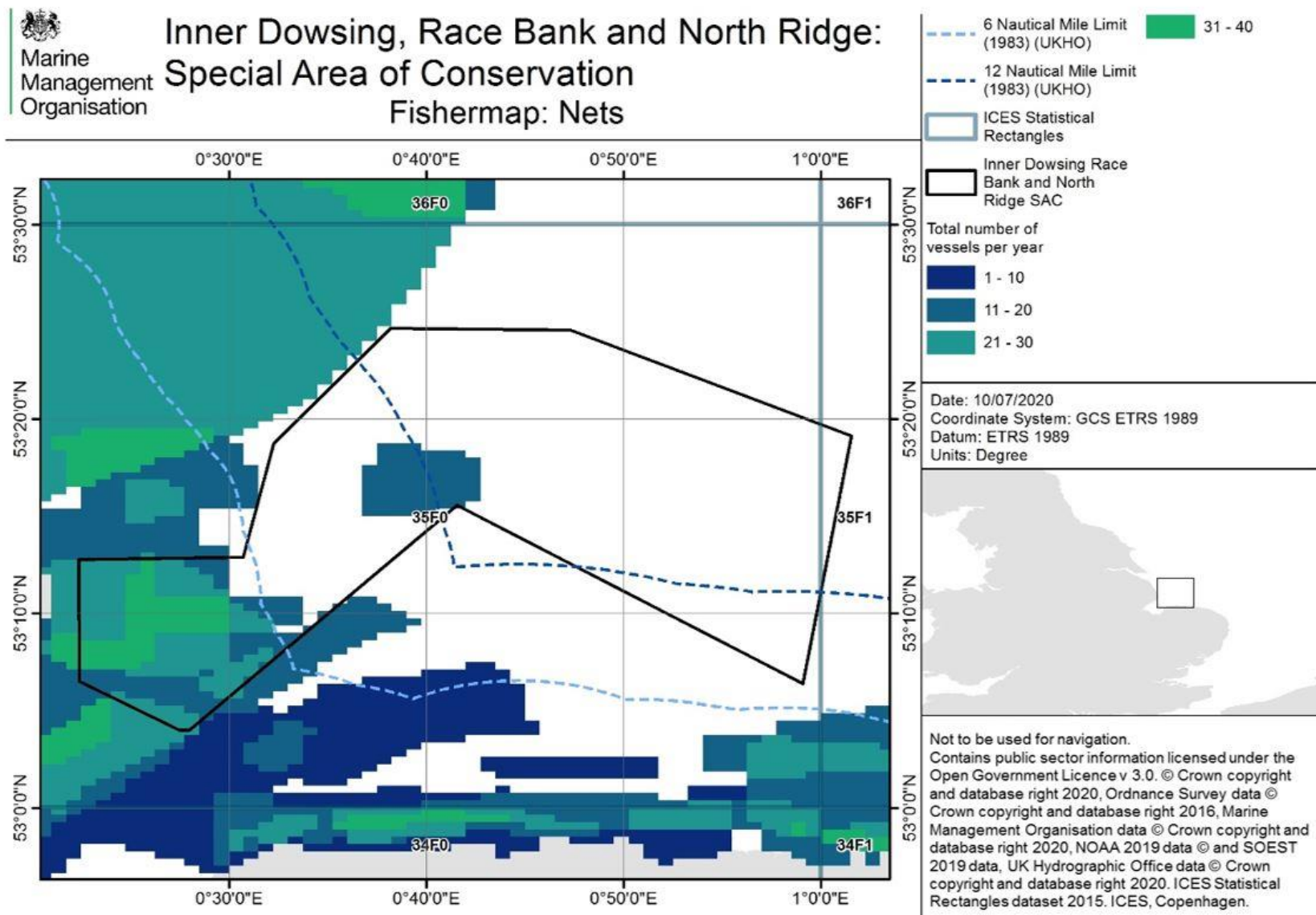
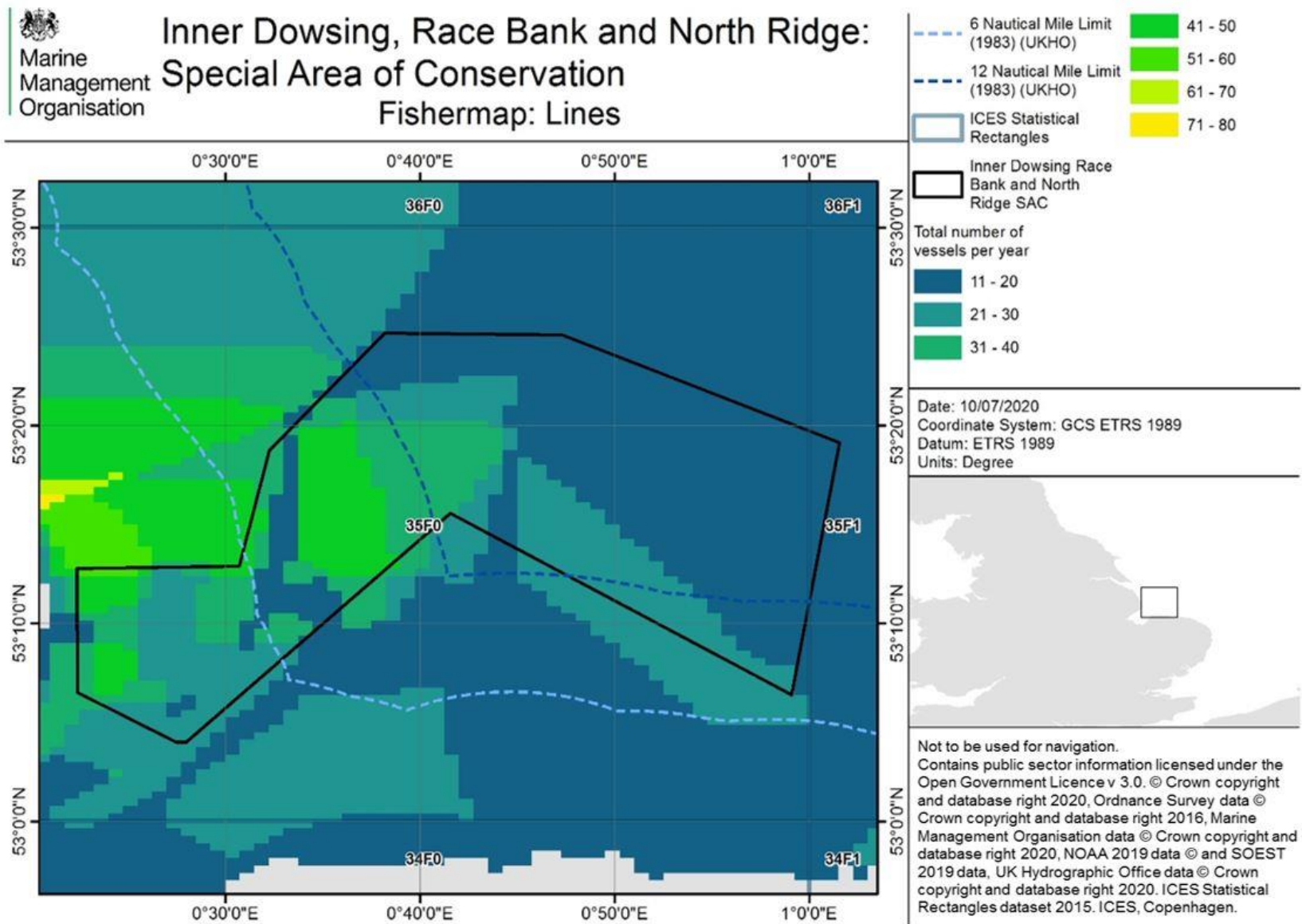


Figure 21: Inner Dowsing, Race Bank and North Ridge SAC FisherMap (2012) – Anchored Lines.



4.1.7 Sightings per unit effort

In 2014, a Defra commissioned project collated sightings data from MMO, IFCA and Navy surveillance from 2010-2012 inclusive and create a gridded geographic data layer of sightings per unit effort (SPUE = number of sightings / surveillance effort) (Vanstaen and Breen, 2014).

Figure 22 to Figure 24 have been included below for gear types where activity was observed (mobile gear - trawling, and static gear - potting). No activity was observed for dredging, netting, angling, lining, or “other” gear types.

4.1.8 MMO Catch recording app

For under 10 m vessels, data from the MMO catch recording project for vessels under 10 m in length has been used at ICES sub-rectangle level. This data is only preliminary information which has been tested from 1 January to 30 November 2020 and does not include the full fleet and is not being used currently for compliance measures. There are also known issues with data quality which include but are not limited to catch quantity, target species, and fishing gear.

Table 28 shows there was relatively limited fishing activity occurring using bottom towed gear across the ICES sub-rectangles, with no fishing taking place in three of the sub-rectangles which covers SAC (35F07, 35F08, 35F09). There were higher levels of fishing using static gear across the rectangles, however the estimated proportion of fishing taking place within the SAC is minimal due to small proportion of designated features which covers the ICES sub-rectangles.

Figure 22: Sightings per unit effort – mobile gear.

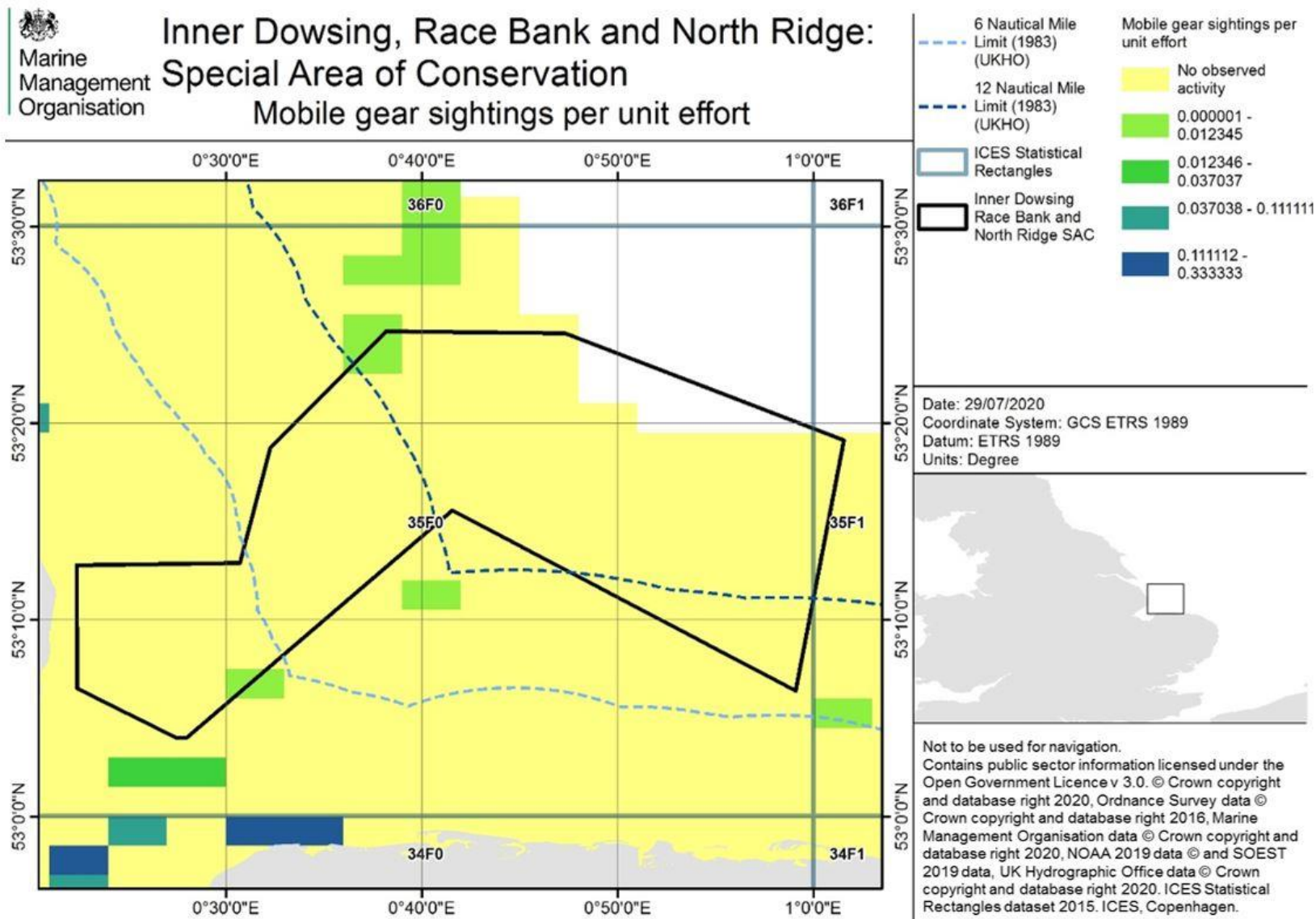


Figure 23: Sightings per unit effort – trawling.

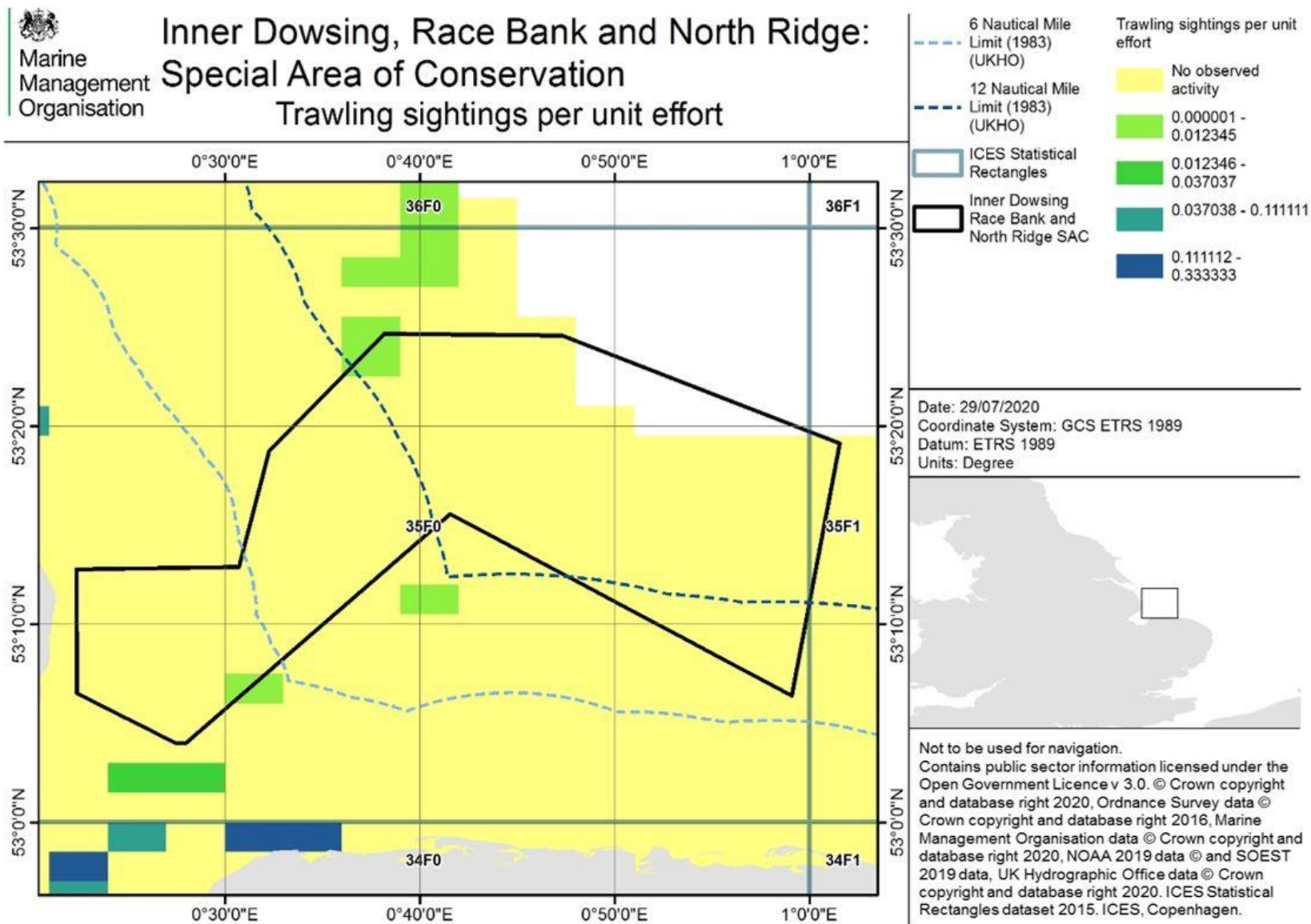


Figure 24: Sightings per unit effort: static gear – potting.

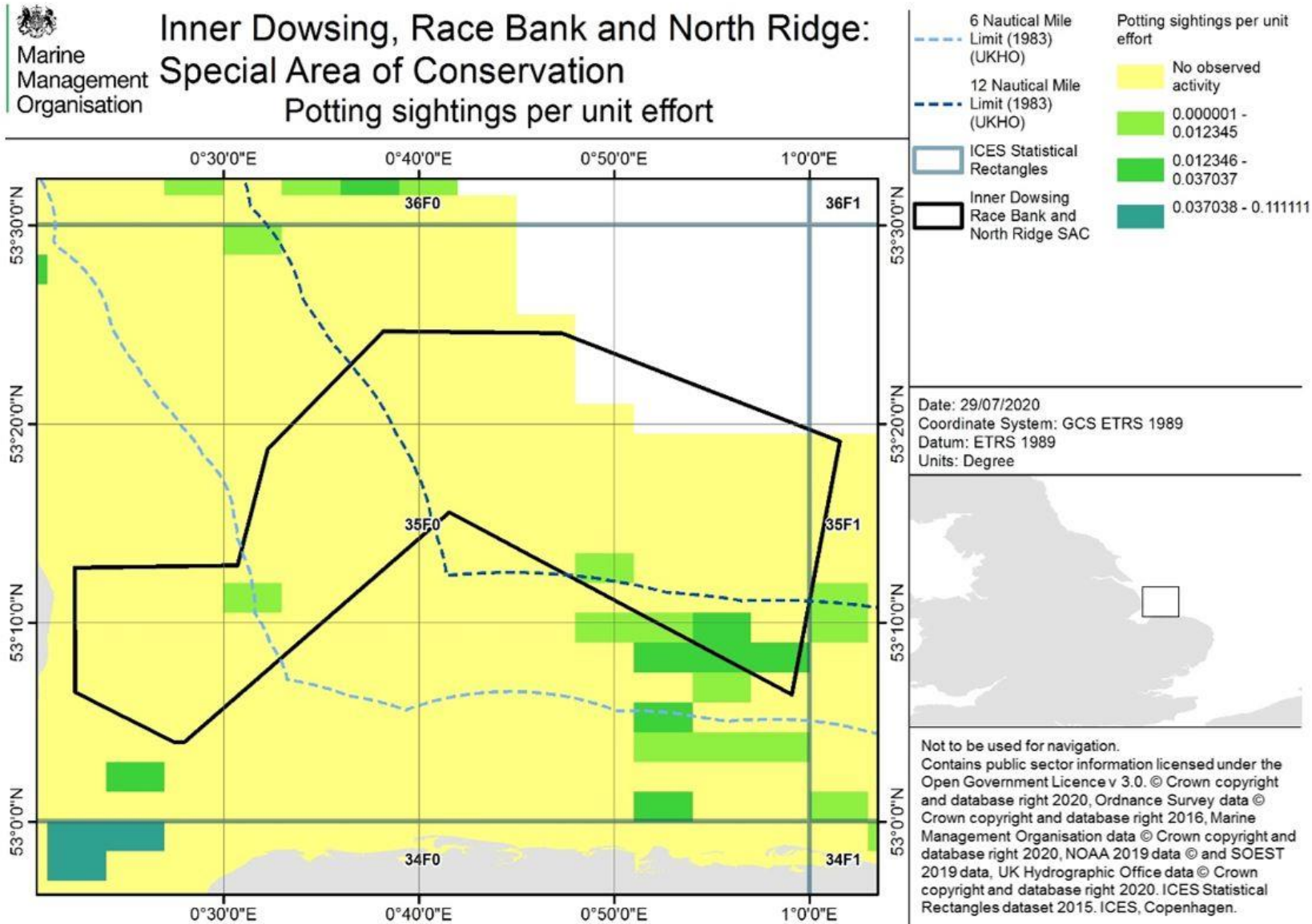


Table 28: The number of fishing trips and catch weight (kg) based on gear type derived from the MMO catch app of under 10 m vessels. Bottom towed gear includes beam trawl and shrimp trawl, static gear includes hand lines, pole lines and pots. No other gear was recorded in the 35F0 sub-rectangles which the PMA covers. There are no vessels operating the specified gears in ICES sub-rectangle 35F12, the only sub rectangle of 35F1 which contains feature. An estimate for the number of trips and catch weight (kg) was also calculated using an area-based proportion method.

ICES sub-rectangles	Bottom towed gear		Static gear		% PMA/sub rectangle		SB PMA		Reef PMA	
	No. of fishing trips	Weight (kg)	No. of fishing trips	Weight (kg)	Reef PMA	SB PMA	No. of fishing trips	Weight (kg)	No. of fishing trips	Weight (kg)
35F04	1	1,095	18	10,465	0.30	NA	N/A	NA	0.06	34.46
35F05	11	13,350	181	54,593	0.03	1.54	0.17	209.81	0.06	21.40
35F07	0	0	90	34,548	0.07	0.87	0	0	0.07	25.39
35F08	0	0	144	155,575	N/A	5.78	0	0	N/A	N/A
35F09	0	0	260	213,475	N/A	0.54	0	0	N/A	N/A

4.1.9 Summary

The MMO portion of Inner Dowsing, Race Bank and North Ridge SAC is an important area for UK fishing vessels using potting gears. While gillnetting and demersal longlining, demersal trawls, dredging and seine netting activities have occurred in the years analysed this has been with little apparent regularity or intensity.

The location of the site and its straddling of the inshore 6 nm IFC District and the offshore (beyond 6 nm) MMO area results in a fishing fleet comprising of both small (under 12 m) and large (over 12 m) vessels. Within the MMO portion of the site the available evidence suggests that the majority of activity from larger (VMS) vessels occurring over the sandbank and reef features between 2014 and 2018 is from potting, gillnetting and seining, however levels of gillnetting appears to have reduced considerably from 2015 onwards. This is likely to be an under representation of the potting activity occurring within the site as it is often undertaken by smaller vessels (Eastern IFCA *pers comms.*). For larger vessels, potting appears to be the only gear to be interacting with the features regularly. Spatial footprint analysis is similar for all gear types analysed between 2014 and 2018 showing small gear footprints and Pr-values.

Understanding the activity of the under 12 m fleet is more complex but all evidence suggests activity is low with relatively few vessels engaged in fishing activity, particularly in the MMO portion of the site which is more difficult to reach by smaller vessels due to the distance from shore. Demersal longline activities appear to be undertaken by the smaller inshore fleet with no records of longlining activity occurring in the VMS data however this too appears to be of low intensity (MMO coastal *pers comms.*) and more commonly occurring over the less sensitive sandbank features than reef (Figure 21).

Individually, the fisheries evidence sources used in this assessment each have limitations (for example in terms of fleet coverage, data confidence, and age). However, when sources are considered together they provide an indication of consistently low fishing effort around the areas of sandbanks and *S. spinulosa* reef within the SAC.

Despite the relatively low levels of activity, it is clear there is interaction, or the potential for interaction, between the fishing activities occurring and the protected Annex I sandbank and reef features of the Inner Dowsing, Race Bank and North Ridge SAC. The sections below examine the pressure that each fishing type exerts on the features of the site.

4.2 Abrasion/disturbance of seabed surface substrate and penetration of the substrate on and below the surface of the seabed

These pressures are relevant to traps, anchored nets/lines, demersal trawls, demersal seines and dredges for the sandbank feature. For the *S. spinulosa* reef feature only the abrasion/disturbance of seabed surface substrate pressure from traps and anchored nets/line gears is assessed here. Consideration is also given to the pressures from the impacts of small weights on *S. spinulosa* reef.

The pressures 'abrasion/disturbance of seabed surface substrate' and 'penetration and/or disturbance of the substrate below the surface of the seabed, including abrasion' are considered together here due to the similarities in the nature of the pressures, and the evidence available to assess impacts.

Abrasion/disturbance and penetration of seabed surface substrate can damage sedimentary habitats by direct damage to infauna and epifauna, particularly in more stable sediments where organisms tend to be more adapted to a lower energy environment and therefore longer-lived, less resilient to sediment movement and direct abrasion and have lower rates of recoverability (Tilin *et al.*, 2010).

4.2.1 Sandbank

The Annex I sandbanks at Inner Dowsing Race Bank and North Ridge SAC contain a range of physical habitats (such as subtidal sand, coarse and mixed sediments) with correspondingly different biological communities. The fauna associated with the crests of sandbanks is predominantly low diversity communities typical of disturbed, mobile sediment environments, although higher diversity assemblages can occur when cobbles or pebbles provide a firmer attachment surface. Along the flanks of the banks, and towards the troughs between the banks the sediments tend to be slightly more stable. In these regions, infaunal and epifaunal communities are more diverse (Institute of Estuarine and Coastal Studies (IECS), 1995; IECS, 1999; Foster-Smith and Sotheran, 1999; Centrica Energy, 2009; Cefas, 2013).

4.2.1.1 Impact of traps and anchored nets/lines

Abrasion from static gears (traps and anchored nets/lines) is possible through the interaction between the seabed and the gear itself (i.e. pots and nets) and associated lines and anchors. This is more likely to occur during hauling of gear or when the gear is subject to strong tides, currents or storm activity. Evidence suggests that static gears have a relatively low impact on benthic communities in comparison to towed gears, as a result of the small footprint of the seabed affected (Roberts *et al.*, 2010). In accordance with this, Hall *et al.* (2008) concluded that assuming they are set correctly, demersal static gears are not considered to have a significant impact on subtidal sand features.

There is limited direct evidence of the impacts of static gears on subtidal sediments. However, Hall *et al.* (2008) reported that all static gears are not considered to be a 'major concern' for subtidal sediments and estimated no or low sensitivity to all but heavy³⁴ levels of fishing intensity on stable species rich sediments or sand and gravel with long-lived bivalves. Hall *et al.* (2008) categorised heavy levels of potting intensity as five pots lifted per hectare per day. In Inner Dowsing, Race Bank and North Ridge SAC 30 pots are estimated to be laid per day per vessel (see Annex 2 - Assumptions used to calculate spatial footprint (Pr-values)). Using the area of the sandbank feature and the number of VMS potting records over said feature a rough estimate of pots per hectare per day in the site is 0.00017. While this only accounts for larger vessels with VMS it is unlikely the inclusion of smaller vessels will increase this to anywhere near the heavy levels described by Hall *et al.* (2008).

As noted previously, evidence from VMS suggests for larger vessels, gill netting is no longer occurring with any kind of regularity in the site and landings evidence for smaller inshore vessels suggests their gillnetting activity levels are similarly low. While potting activity has increased since 2014, in particular for larger vessels with VMS, the majority of activity does not appear to occur over the sandbank feature. The exact location of the potting activities of smaller vessels is not known but they are likely to target similar areas to larger vessels albeit perhaps closer inshore. FisherMap data contradicts that of VMS and suggests the highest levels of potting activity occur over the Race Bank sandbank in the offshore portion of the site, this could be due to the construction of the Race Bank windfarm since the FisherMap data was collected. It is also assumed that this activity can be most likely attributed to larger vessels given the distance from shore and as these larger vessels are included within VMS data, the FisherMap data is not considered to represent the most up to date location of potting activities.

Given the levels of sensitivity of subtidal sandbanks (subtidal mixed sediment is classed as medium sensitivity and subtidal sand is classed as not sensitive to medium sensitivity) at this site to abrasion and disturbance, and the low levels of static gear fishing activity on the sandbank features, **MMO has concluded that abrasion/disturbance and penetration of the substrate on and below the surface of the seabed pressures associated with traps and anchored nets/lines are compatible with the conservation objectives of the site and will not result in an adverse effect on site integrity.**

³⁴ Quantitative fishing intensity levels used are published in Hall *et al.* 2008. Heavy potting intensity was defined as 'more than 5 pots lifted per hectare per day'.

4.2.1.2 Impact of demersal trawls

As it is pulled across the seabed, various parts of demersal trawl can cause penetration, abrasion or disturbance of the seabed surface substrate.

For otter trawls, the otter boards/doors can penetrate the sediment and the depth of penetration depends on the width of gear, the door weight and the hardness of the sediment (Løkkeborg, 2005; Grieve *et al.*, 2011; Eigaard *et al.*, 2016a). The footrope, ground rope and bridles may also come into contact with the seabed (Grieve *et al.*, 2011). Furrows and berms are created through physical impact of trawl doors on the sediment, thus creating irregular features on the seabed (Løkkeborg, 2005). Otter trawling has been shown to create berms and furrows on sandy substrates, with repetitive trawling causing increased surface relief or roughness (Schwinghamer *et al.*, 1998). Trawl doors can penetrate up to 10 cm into sand, gravel and mixed substrates, with associated chains penetrating up to 8 cm (Eigaard *et al.*, 2016a; Humborstad *et al.*, 2004). Otter trawls have been shown to create visible paths and furrows on substrates dominated by pebbles less than 6.5 cm in diameter (Freese *et al.*, 1999). Eigaard *et al.* (2016 a, b) estimated that the subsurface ratio (proportion of the gear footprint where gear components penetrate the seafloor by 2 cm) for otter trawls ranges from 0.078 to 0.304, depending on target species. Otter trawls are unlikely to significantly impact the large-scale topography or sediment composition of the sandbank feature, however, impacts to the biological structure are likely are discussed below.

For beam trawlers, the main effects are from the 'shoes' or 'sleds', which on the largest vessels, can penetrate the seabed up to 6 cm, but if rockhoppers (wheels attached to the front of the trawl to help it bounce over obstacles) or tickler chains (chains which flush organisms out of the sediment into the trawl) are used, these can also impact the seabed, penetrating up to 2.2 cm (Tilin *et al.*, 2010, Grieve *et al.*, 2011). The chains of a beam trawl cover the whole width of the gear and are designed to penetrate the upper few centimetres of the sediment, ranging from a few centimetres to at least 8 cm (Løkkeborg, 2005). Beam trawls have been described to cause a flattening of bottom features such as ripples and irregular topography (Kaiser *et al.*, 1996). Beam trawl shoes and tickler chains penetrate up to 10 cm into sandy, coarse and mixed sediments (Eigaard *et al.*, 2016a). Side scan observations have indicated that beam trawling creates clear marks in fine and medium sand habitats, with seabed roughness decreasing and hardness increasing directly after the trawls (Fonteyne, 2000, Løkkeborg, 2005;). Tickler chains may also turn, displace and even remove larger pebbles and boulders in areas with mixed sediments (Eigaard *et al.*, 2016a; JNCC, *pers. comm.*). Despite this, seabed characteristics of sandy substrates have been shown to return to their original levels in 15 hours following beam trawling (Løkkeborg, 2005). Eigaard *et al.* (2016 a, b) estimated that the subsurface ratio (proportion of the gear footprint where gear components penetrate the seafloor by 2 cm) for beam trawls ranges from 0.522 to

1.000, depending on target species. As above with otter trawls, beam trawls are unlikely to significantly impact the large-scale topography or sediment composition of the sandbank feature. The likelihood of impacts to the biological structure are discussed below.

The effects of demersal trawling on sedimentary habitats can vary depending on site conditions (e.g. wave/tidal energy) with low mobility sediments being more sensitive due to the more developed epifauna and infauna (Hall *et al.*, 2008, Lambert *et al.*, 2014). Evidence of the impacts of towed gears varies depending on the gear type, particularly gear penetration depth (Sciberras *et al.*, 2018). Demersal trawling in the site is predominantly beam trawling for shrimp, which uses lighter gear than whitefish beam trawls and does not use tickler chains. Out of the different bottom trawl fisheries in the North Sea, beam trawling for brown shrimp can have the lowest impacts on the seafloor status (Rijnsdorp *et al.*, 2020).

Hall *et al.* (2008) determined sensitivities of different sediment types to types of demersal towed gears at different levels of activity. Available evidence suggests demersal towed gear activity in the MMO portion of the site is low with few VMS records and few landings deriving from these gears. What little activity does occur does not appear to be concentrated over the sandbank feature. As such demersal trawling activity would fall into the 'low' category as described by Hall *et al.* (2008).

Given beam trawling for shrimp uses lighter gear without tickler chains, and the low level of this activity in the site, both unstable coarse sediments with robust fauna and species rich mixed sediments had low levels of sensitivity to shrimp trawling. Unstable coarse sediments with robust fauna were also reported to have low sensitivity to other kinds of demersal towed gear, however species rich mixed sediment was reported to have medium sensitivity to other demersal towed gear. Given that the more sensitive sandbank sub-feature at this site is classed as having moderate diversity, it is likely that its sensitivity falls between these two.

Kaiser *et al.* (2006) undertook a meta-analysis of 101 fishing impact manipulations and reported that beam trawling resulted in an immediate 70% reduction on benthic fauna of subtidal sand (based on 2 studies providing 53 data points). However, by two to seven days after the fishing event, no change was detectable (based on one study with eight data points). Similarly, muddy sand was found to experience a 35% reduction in benthic fauna immediately following beam trawling (based on two studies providing 61 data points), although this effect had disappeared after one week (based on one study providing 2 data points). An immediate reduction in benthic fauna in gravel sediments of around 40% was also reported (two studies providing 28 data points), although this was not statistically significant. There were no longer-term studies available for gravel sediments.

Kaiser *et al.* (1998) reported that beam trawling on high mobility sand in greater than 30 metres depth was found to have no detectable effect on benthic infauna 24 hours after fishing or when the feature was assessed 6 months later.

The levels of demersal trawling in the site are relatively low, and predominantly consists of beam trawling for brown shrimp. Parts of the sandbank feature are also considered to have a relatively low sensitivity to the effects of demersal trawling activity. However, there is clear evidence that demersal trawling impacts sandbank communities, particularly longer-lived and fragile species which may form part of sub-features across the site irrespective of their overall assessed sensitivity. In addition, the sandbank feature has been assessed by Natural England as being in unfavourable condition and has the target to restore the presence and spatial distribution of subtidal sandbank communities. **MMO therefore conclude that abrasion, disturbance and penetration of the substrate on and below the surface of the seabed pressures may not be compatible with the conservation objectives of the site and an adverse effect on site integrity due to demersal trawls cannot be ruled out.**

4.2.1.3 Impact of dredges

The potential for abrasion and disturbance from dredging is due to occasional scallop dredging and mussel prospecting for seed mussel (*Mytilus edulis*). Mussel seed beds are ephemeral therefore mussel prospecting can occur sporadically around the optimal period for relaying in late summer. This fishery is small scale and generally will only occur one week of the year.

The ground gear of dredges used for catching molluscs is mostly homogenous across the entire width of the dredge, with the exception of scallop dredges that have teeth protruding into the sediment (Eigaard *et al.*, 2016a). Scallop dredges therefore produce a more uneven sediment furrow (Eigaard *et al.*, 2016a, O'Neill *et al.*, 2013). Scallop dredging can cause a flattening of irregular bottom topography by eliminating natural features such as ripples, bioturbation mounds and faunal tubes (Løkkeborg, 2005). The ground gear of dredges can penetrate up to 15 cm into sandy substrates (Eigaard *et al.*, 2016a). A study by Lambert *et al.* (2015) and Murray *et al.* (2015) demonstrated how tracks from scallop dredges persisted for up to ten months in coarse sediment, whereas dredge tracks were not found to be visible in sand. This impact on the physical structure of the sandbank is not compatible with the restore structure and function target for the site.

The epifauna and infaunal assemblages of both stable and dynamic fine sands are known to be susceptible to direct physical disturbance from dredges which penetrate and disturb the sediment (Roberts *et al.*, 2010). A meta-analysis by Kaiser *et al.* (2006) indicated that both deposit- and suspension-feeders were consistently vulnerable to scallop dredging across gravel, sand and mud habitats. Slow-growing

species, such as soft corals took much longer to recover (up to 8 year) from scallop dredging than biota with shorter lifespans such as polychaetes (<1 year) (Kaiser *et al.*, 2006). Therefore, surface and sub-surface abrasion and penetration by demersal dredges may impact the biological communities found in the sandbank feature. As described for demersal trawls in section 4.2.1.2, dredges may adversely impact infauna and epifauna found on the sandbank feature through direct physical impacts. This impact is not compatible with the restore extent and distribution and structure and function targets for the site with regards to the biological communities.

Mussel prospecting is a non-licensable activity and will only require a consent to land the fishery or to relay the mussel seed on to prospective aquaculture beds. The potential for this fishery to occur is limited to late summer as this is the optimal time for successful transportation of seed. It is not envisaged that the current scale or magnitude of this potential fishery will have a significant effect on the site however; this potential activity will be monitored in the site. Mussel seed prospecting depends on the recruitment of mussel seed and tends to be less of an impact than scallop dredging (Hall *et al.*, 2008).

VMS data, FisherMap and SPUE data suggests activity levels in the site are very low but interaction with the sandbank feature does occur. Given the unfavourable condition of the sandbank feature in the site and the target to restore the presence and spatial distribution of subtidal sandbank communities, **MMO conclude that abrasion, disturbance and penetration of the substrate on and below the surface of the seabed pressures may not be compatible with the conservation objectives of the site and an adverse effect on site integrity due to dredges cannot be ruled out.**

4.2.1.4 Impact of demersal seines

Demersal seines have been identified as gear types which may have an impact via surface and sub-surface abrasion and penetration on the sandbank feature. Biotopes containing attached or sessile epifauna are considered sensitive to abrasion due to the removal of these non-target species (MBIEG, 2020).

Demersal seine hauls can impact the seabed either via contact of the seine rope or ground gear, with the largest impact by area coming from the seine rope when they are pulled together in the first phase of fishing operation (Eigaard *et al.*, 2016a, Rijnsdorp, 2013).

Given the absence of otter boards and lighter groundgear, seines tend to be considered as less damaging to seabed habitats via abrasion and penetration compared to other demersal gear types (Polet & Depestele, 2010). Eigaard *et al.* (2016 a, b) estimated the sub-surface ratio to be < 0.001 for Danish seines. In comparison, predicted sub-surface ratios for otter trawls ranged from 0.078 to 0.304

and from 0.522 to 1.000 for beam trawls, depending on target species (Eigaard *et al.*, 2016 a, b). The physical structure of the sandbank feature is therefore unlikely to be impacted by demersal seines. Thus, demersal seining activity is considered compatible with the target for the sandbank to restore the distribution of sediment composition across the feature.

The predictions of Eigaard *et al.* (2016 a, b) are in line with the conclusions of MBIEG (2020) which suggest that demersal seines alone may not have a significant impact on benthic communities via surface abrasion and subsurface penetration where sessile or attached epifauna are absent. However, the sandbanks of this site are home to a wide range of sessile and attached epifauna including bryozoans, sponges, hydroids and tube building worms. These species may be sensitive to the impact of abrasion through damage and removal as by-catch (Waardenburg, 2017). Despite the lower sub-surface ratios for demersal seines compared to beam and otter trawls, the surface footprint of Danish seines (1 km²), defined as the surface area covered during one hour of fishing, is relatively high compared to the otter trawl (0.3 – 1.2 km²) and beam trawl (0.2 km²) (Eigaard *et al.*, 2016a, Rijnsdorp, 2015). As a result, demersal seining may affect the distribution and structure of the benthic communities. This impact would not be compatible with the favourable condition target of the site to restore the presence and spatial distribution of subtidal sandbank communities. The abrasion/penetration pressure through removal of non-target species is explored further in section 4.4.

VMS data indicated that only UK vessels appear to use seine gears in the site and only Danish/anchor seines, this has since been determined to be vessels working on the adjacent wind farms as there are no landings associated with any of these vessels and has subsequently been removed from the VMS analysis.

Given the unfavourable condition of the sandbank feature in the site and the target to restore the presence and spatial distribution of subtidal sandbank communities, **MMO conclude that impacts of surface abrasion on the sandbank feature from demersal seines are not compatible with the conservation objectives of the site and will result in an adverse effect on site integrity.**

4.2.2 *Sabellaria spinulosa* reef

Abrasion/disturbance of the seabed can impact *S. spinulosa* reefs in several ways. Physical abrasion can break off or damage parts of the reef, reducing its extent and reducing growth rates. This pressure includes unintentional removal of *S. spinulosa* reef by fishing gears targeting other species.

Abrasion/disturbance of seabed surface substrate may also result in higher sediment loads, which could affect reef formation. However high suspended sediment loads would be unlikely to affect *S. spinulosa* reef as they are evolved to exist in, and are

dependent on, such conditions to promote reef growth. Therefore, the reef is not considered to be sensitive to changes (increases) in suspended sediment loads (JNCC and Natural England 2013).

4.2.2.1 Impact of traps and anchored nets/lines

Static gears can damage *S. spinulosa* reef through gear (demersal longlines, pots, nets and their associated anchors or lines) striking or becoming entangled with the reef. This is most likely to occur upon deployment, through movement of gear on the benthos due to tide, current and storm activity, and as the gear is dragged along the seafloor on retrieval (Coleman *et al.*, 2013, Grieve *et al.*, 2014). Although focussed on the impact of potting on rocky reef, Gall *et al.* (2020) highlights that the impact of potting is more destructive on reef habitats than previously thought, impacting 14 of the 18 taxa identified in the study including all the indicator species. As the first study to quantify the true impact of potting, considering the impact of a string of pots during the deployment, soaking and hauling of the gear in concluded that although the impacted area would not be the entire pot haul path, it is more damaging than previously considered, indicating a minimum recovery time of key rocky reef associated species of 6 to 36 months, with longer term impacts unknown, providing further disturbance does not occur (Stephenson *et al.*, 2017, Gall *et al.*, 2020). Whilst understanding that there are substantial differences in recovery of rocky reef and biogenic reef, specifically with recovery rates, we cannot rule out that there is evidence available highlighting that potting and static gears over reef are more damaging than originally thought.

Walmsley *et al.* (2015) noted that there is no primary evidence on the impact of potting on *S. spinulosa* reef. However, sensitivity assessments based on expert knowledge are available. Table 29 shows a summary of several sensitivity assessments which have considered the sensitivity of *Sabellaria spp.* to impacts from static gears.

Table 29: Summary of sensitivity assessments for potting impacts on *S. spinulosa* reef.

Reference	Summary	Notes
Eno <i>et al.</i> (2013)	Honeycomb-worm (<i>S. alveolata</i>) reefs have medium sensitivity to all levels of potting and to high levels of netting or lining. These reefs have low or no sensitivity to all other levels of netting or lining.	Sensitivity was generated by combining semi-quantitative scores for resilience and recoverability. Quantitative fishing intensity levels were not published. Intensity levels were based on fishing practices around Wales. Sensitivity was not assessed for <i>S. spinulosa</i> reef and evidence suggests <i>S. spinulosa</i> is more fragile than <i>S. alveolata</i> and would therefore show greater sensitivity to

		these gears and levels of intensity (Gibb <i>et al.</i> , 2014).
Hall <i>et al.</i> (2008)	Biogenic reef on sediment habitats have medium sensitivity to heavy levels of potting and low sensitivity to all other levels of potting.	Sensitivity was assessed in terms of various factors including degree of physical disturbance, size of area damaged, effect on fauna and community makeup. Fishing intensity levels are quantified in Appendix 3 of the report.
Roberts <i>et al.</i> (2010)	<i>S. spinulosa</i> reefs may be affected by the use of static and towed fishing gears.	Assessment was based on existing literature. Sensitivity is assessed as a combination of resistance and resilience.
Tilin <i>et al.</i> (2010)	<i>S. spinulosa</i> reefs have a low sensitivity to surface abrasion.	Sensitivity assessments were based on a combination of resistance (tolerance) and resilience (recovery). Sensitivities were assessed in terms of pressure benchmarks rather than particular activities. The pressure benchmark for surface abrasion was “damage to seabed surface features.

Gibb *et al.* (2014) reviewed the sensitivity of *S. spinulosa* reef to various pressures, including abrasion/disturbance of seabed sediment, however this pressure was not linked directly to static fishing gears. Gibb *et al.* (2014) cites studies which show *S. alveolata* reefs recovered within 23 days from trampling, walking and stamping (Cunningham *et al.*, 1984).

However, Cunningham *et al.* (1984) also reported that more severe damage caused by kicking and jumping on the reef was still not fully repaired 23 days later. *S. spinulosa* reefs are also recorded to be more fragile and less resilient than *S. alveolata* reefs, meaning the impacts of abrasion/disturbance may be greater and recovery times longer (Gibb *et al.*, 2014) than those observed in *S. alveolata* by Cunningham *et al.* (1984).

Vorberg (2000) demonstrated that large well attached aggregations of *S. alveolata* reef shows rapid recovery from the impacts of shrimp trawling, but it was highlighted in Last *et al.* (2012) that aggregations of *S. spinulosa* reef which are patchier or resting on mixed sediment could be more impacted by abrasion pressures. It is therefore suggested that surface abrasion may lead to greater damage and a lower recovery rate for *S. spinulosa* than that of *S. alveolata* and no direct observations of reef recovery through repair from abrasion were found for *S. spinulosa* (Gibb *et al.*, 2014).

VMS data indicate that for larger vessels, potting activity occurs in close proximity to, but rarely directly over, the *S. spinulosa* reef. However, given the length of a potting string and the uncertainty of the location of the string in accordance with the VMS location it is certainly possible that pots are laid within the *S. spinulosa* reef area. For smaller vessels, exact locations for use of static gear is unknown but they are likely to be in similar areas to the larger vessels and FisherMap and SPUE data suggest activity levels, while low, could potentially be occurring in areas of *S. spinulosa* reef.

The site has targets to restore the presence and spatial distribution of reef communities, restore the total extent, spatial distributions and types of reef, as well as restore the species composition of component communities. Abrasion to *S. spinulosa* from potting will not further these targets. Resistance of *S. spinulosa* is assessed as 'low' due to the likely damage to the tubes and sub-lethal and lethal damage to the worms via abrasion (Gibb *et al.* 2014).

Given the unfavourable condition of the reef feature in the site and the targets to restore the presence and spatial distribution of reef communities, the total extent, spatial distribution of reef and the species composition of component communities **MMO conclude that abrasion, disturbance and penetration of the substrate on and below the surface of the seabed pressures may not be compatible with the conservation objectives of the site and an adverse effect on site integrity due to traps, anchored lines/nets cannot be ruled out.**

4.2.2.2 Impacts of small weights from marine recreational fishing (angling)

Fishing weights (alongside leaders, glow sticks, hooks and lines) from recreational fishing can contribute to marine debris (Chiappone *et al.*, 2002; Lewin *et al.*, 2020; Schernewski *et al.*, 2018). This 'ghost gear' may in-turn cause localised habitat degradation through entanglement with fauna (Cooke & Cowx, 2006). A study of small-scale coastal fisheries in European seas also suggests that lost gears can damage sessile organisms through abrasion (Lloret *et al.*, 2018).

However, the overall physical pressures exerted from recreational fishing equipment are likely to be low. Lewin *et al.*, (2019) undertook an extensive literature review of the environmental impacts of marine recreational fishing (MRF) and found that the risks associated with MRF litter were low (the impacts occur locally and are reversible). MRF litter is therefore unlikely to substantially impact wildlife populations and, even where MRF activity is high (Lewin *et al.*, 2019). Other studies by Lewin state that destruction of benthic habitats is less significant for recreational fishing than it is for commercial marine fisheries (Lewin *et al.*, 2006).

Scientific evidence for the effects of angling on biogenic habitats is sparse (Hall *et al.*, 2008). Due to the lack of data providing evidence on interaction of abrasion from

small weights used in recreational angling, MMO has used a proxy to understand potential impacts of sea angling on *Sabellaria* reef. However, studies of other weighted objects, weighing more than the weights than those used for MRF³⁵, indicate that sea angling weights are unlikely to cause an issue for *Sabellaria* via abrasion. The evidence of abrasion of static gears on *S. spinulosa* (Section 4.2.2.1) is from heavier gears with a larger footprint than the weight used in MRF. There is some evidence of potential physical impact (abrasion) from small weights used in recreational sea angling, however the evidence indicates that the light weights used during MRF (angling) will not significantly damage *S. spinulosa* reef. **MMO conclude that abrasion from recreational fishing (angling) is not resulting in an adverse effect on site integrity.**

4.2.3 Summary of abrasion, disturbance of seabed surface substrate and penetration of the substrate on and below the surface of the seabed pressures on sandbank and reef features

4.2.3.1 Sandbank

Given the evidence above, surface abrasion and sub-surface penetration caused by anchored nets/lines alone is unlikely to hinder the restoration of the extent and distribution as well as structure and function of the sandbank feature. **MMO conclude that anchored nets/lines and traps are compatible with the conservation objectives of the site and will not result in an adverse effect on site integrity via this pressure.**

There is a risk that surface abrasion and sub-surface penetration caused by demersal trawls, demersal seines and dredges may hinder the achievement of favourable condition targets. Use of these gear types may impact the physical and biological structure of the sandbank feature via direct physical impacts from gear interacting with the seabed and species. This may impact the extent and distribution of biological assemblages. **MMO conclude that demersal trawls, demersal seines or dredges are not compatible with the conservation objectives of the site and may result in an adverse effect on site integrity via this pressure (Table 30).**

4.2.3.2 *S. spinulosa* reef

While expert opinion from Eastern IFCA indicate that potters are unlikely to directly target *S. spinulosa* reef areas other academic literature (Section 4.2.2.1) an interaction of traps and anchored nets/lines cannot be ruled out. The impact of anchors and weighted pots landing on *S. spinulosa* reef is likely to be similar to the damage exerted by Cunningham *et al.* (1984) on *S. alveolata* reefs through trampling/stamping/kicking and jumping etc. and while recovery is possible in reasonable time frames, given the increased sensitivity of *S. spinulosa* compared

³⁵ <https://britishseafishing.co.uk/terminal-tackle-2/weights/>

with *S. alveolata* (Gibb *et al.*, 2014) and the current unfavourable condition of the feature in the site and the requirement to recover the feature to favourable status **MMO conclude that traps and anchored nets/lines may not be compatible with the conservation objectives of the site and an adverse effect on site integrity of these gears cannot be ruled out (Table 30).**

Given the evidence above, surface abrasion caused by small weights associated with recreational angling alone is unlikely to hinder the restoration of the extent and distribution as well as structure and function of the reef feature. **MMO conclude that small weights associated with recreational angling are compatible with the conservation objectives of the site and will not result in an adverse effect on site integrity via this pressure.**

Table 30: Sandbank and reef features abrasion/disturbance and penetration assessment

Pressure	Feature	Favourable condition target	Activity	Compatible with the conservation objectives?
Abrasion/ disturbance of the substrate on the surface of the seabed and Penetration and/or disturbance of the substrate below the surface of the seabed, including abrasion	Sandbank	Restore the distribution of sediment composition across the feature (and each of its sub-features).	Traps	Yes
			Nets	Yes
			Demersal trawl	No
			Dredges	No
			Demersal seines	No
		Restore the presence and spatial distribution of subtidal sandbank communities.	Traps	Yes
			Nets	Yes
			Demersal trawl	No
			Dredges	No
			Demersal seines	No
Abrasion or disturbance of the substrate on the surface of the seabed	Reef (<i>S. spinulosa</i> reef)	No reduction in extent of reef, subject to natural change.	Lines	No
			Pots	No
			Nets	No
			Small weights	Yes
		Reef shows no significant decline in community with different growth phases present, subject to natural change.	Lines	No
			Pots	No
			Nets	No
			Small weights	Yes
		Maintain age/size class structure of individual species, subject to natural change.	Lines	No
			Pots	No
			Nets	No
			Small weights	Yes

4.3 Removal of target species

Fishing gears are designed to remove target species from the marine environment. Impacts of traps, anchored nets/lines and dredges have been assessed for this pressure on the sandbank feature. No likely significant effect was determined for demersal trawls and seines in Part A, so these gear types are not considered in this section.

4.3.1 Sandbank

4.3.1.1 Impacts of traps

Traps in this area target crustacea and gastropod molluscs. In terms of crustacea, lobsters and crabs are the most common target species and whelks are the predominant gastropod mollusc caught. Between 2014 and 2019 an estimated total of 76.82 tonnes were landed by over 12 m UK potting vessels within the site. VMS charts indicate that potting by UK vessels with VMS over the sandbank feature is minimal and so these vessels are unlikely to be having a significant impact on target species within the sandbank feature. An estimated 695.5 tonnes were also landed from non-VMS UK potting vessels during this time period, with an estimated 120.58 to 147.2 tonnes landed each year. This data indicates that landings from non-VMS vessels from within the site has remained consistent between 2014 and 2019. Pr-values indicate a low footprint from potting activity between 2014 and 2019, with a total gear footprint over the sandbank feature of 0.000002 – 0.000037 km². Given that traps can be altered with the appropriate use of mesh sizes in cover netting and escape gaps, juvenile target species are at low risk from traps³⁶. Consistency in effort in addition to protection of juvenile stock reduces the risk to target species.

Considering the discussion above, **MMO conclude that impacts from removal of target species by traps on the sandbank feature are compatible with the conservation objectives of the site and will not result in an adverse effect on site integrity.**

4.3.1.2 Impacts of anchored nets/lines

Anchored nets in this area target bass, cod, pollock, sole and anglerfish. There are no landings estimated from anchored nets/lines from over 12 m UK vessels within the site between 2014 and 2019. Despite this, VMS charts show that anchored nets/lines are used over the sandbank feature in 2014, however, in subsequent years activity is absent over the sandbank feature. Landings from anchored nets/lines from non-VMS UK vessels are minimal, estimated at 5.7 tonnes between 2014 and 2018. Non-UK landings from anchored nets/lines are also very low, with the only record in STECF data being in 2013 from French vessels at 0.37 tonnes. Pr-

values are consistent with these low activity levels with the total gear footprint between 2014 and 2019 being 0.0004 – 0.087 km². These low activity levels suggest that there is low risk of impacts on target species within the site.

Considering the discussion above, **MMO conclude that impacts from removal of target species by anchored nets/lines on the sandbank feature are compatible with the conservation objectives of the site and will not result in an adverse effect on site integrity.**

4.3.1.3 Impacts of dredges

Dredges in this area target seed mussels and scallops. Seed mussel dredging occurs sporadically typically occurring during one week in late summer (Eastern IFCA, *pers comms*). This fishery is considered to be small scale and therefore is unlikely to have a significant impact. Landings from dredges from UK vessels are estimated at 368.42 tonnes from 2014 to 2018. This is skewed by an estimated 247.54 tonnes being landed in 2014 with landings in subsequent years not rising above 62 tonnes. There are no landings recorded from dredges from non-UK vessels. VMS data indicates minimal dredging activity over the sandbank feature with an average Pr-value of 0.0029 - 0.001 between 2014 and 2019. This suggest that dredging is not a significant risk to target species.

Considering the discussion above, **MMO conclude that impacts from removal of target by dredges on the sandbank feature are compatible with the conservation objectives of the site and will not result in an adverse effect on site integrity.**

4.3.2 Summary of removal of target species Sandbank assessment

Given the evidence above, removal of target species within the site is unlikely to hinder the targets to restore the presence and distribution or maintain species composition of sandbank communities. **MMO conclude that traps, anchored nets/lines and dredges alone are compatible with the conservation objectives of the site and will not result in an adverse effect on site integrity via this pressure (Table 31).**

Table 31: Removal of target species assessment

Pressure	Feature	Favourable condition target	Activity	Compatible with the conservation objectives?
Removal of target species	Sandbank	Restore the presence and spatial distribution of subtidal sandbank communities.	Traps	Yes
			Anchored nets/lines	Yes
			Dredges	Yes
		Maintain the species composition of component communities.	Traps	Yes
			Anchored nets/lines	Yes
			Dredges	Yes

4.4 Removal of non-target species

Fishing may remove by-catch ('non-target') species, depending on the gear, methods used and ecological makeup of the fishery. Impacts of traps, anchored nets/lines, demersal trawls, demersal seines and dredges have been assessed for this pressure on the sandbank features. Impacts of traps and anchored nets/lines have been assessed for the *S. spinulosa* reef feature.

4.4.1 Sandbank

4.4.1.1 Impacts of traps

In terms of non-target species, by-catch from crab and lobster pots around the UK is low. A Marine Stewardship Council report found that only 1% of total catch (excluding undersize and berried individuals returned to the sea before landing) was made up of by-catch in the crab potting fishery around the Shetland Islands (Hervás *et al.*, 2012). Very little by-catch is expected from pots and traps as the design means that fish and shellfish can escape easily before the gear is hauled³⁶. Any by-catch can also be released back into the sea immediately without harm¹. Epifauna such as sea fans have been shown to be able to recover from all creel impacts, by bending to avoid the impact of dropped creels and reinserting themselves following uprooting (Eno *et al.*, 2001). Trapping activity is therefore compatible with the favourable condition target to maintain the distribution of subtidal sandbank communities and will not adversely impact species richness or species of ecological importance.

Considering the discussion above, **MMO conclude that impacts from removal of non-target species by traps on the sandbank feature are compatible with the**

³⁶ <https://seafish.org/gear-database/gear/pots-and-traps/>

conservation objectives of the site and will not result in an adverse effect on site integrity.

4.4.1.2 Impacts of anchored nets/lines

Anchored nets such as gill nets have the potential to damage and/or remove non-target species. Species that are likely to become entangled include diving seabirds, seals and cetaceans (Gislason, 1994) and erect, branching benthic species such as pink sea fans (*Eunicella verrucosa*) (Eno *et al.*, 2013). Characteristic communities within the sandbank feature are not known to include the species listed above, with epifauna mostly consisting of bryozoans, sponges and hydroids on more gravelly areas. Given that anchored nets are not towed, these species are unlikely to be removed by gill nets. Anchored net/line activity is therefore compatible with the favourable condition target to restore the presence and distribution or maintain species composition of sandbank communities.

Considering the discussion above, **MMO conclude that impacts from removal of non-target species by anchored nets/lines on the sandbank feature are compatible with the conservation objectives of the site and will not result in an adverse effect on site integrity.**

4.4.1.3 Impacts of demersal trawls

Demersal trawls interact directly with the seabed and penetrate into the sediment which means that species occupying this area may be removed by passing gear.

Demersal trawls and dredges may remove crabs, lobsters, molluscs and sessile epifauna as by-catch. The mortality of non-target species caught by demersal gear such as beam trawls varies. One study found that mortality ranges from 0% for hermit crab, whelks and starfish to 100% for shells such as *Arctica islandica* (Gislason, 1994). De Groot and Lindeboom (1994) found that high mortalities occurred for undersized fish discarded, 50% or less for most crabs and molluscs and very little mortality (<10%) for starfish. Overall findings indicated a decrease of 0-85% from initial numbers for different mollusc species (solid-shelled or very small species (De Groot & Lindeboom, 1994). Mobile epifauna, attached epifauna, polychaete worms and amphipods are characteristic of the sandbank feature of the site. These may therefore be removed or damaged by demersal trawls.

VMS data indicate that demersal trawling by over 12 m vessels over the sandbank feature has declined since 2014 with only a small number of records in recent years. The majority of activity takes place outside the boundary of the site and does not frequently take place over the sandbank feature. Expert opinion states that two under 10 m beam trawlers are active within the site. Despite the activity levels being

low, due to the restore target for presence and spatial distribution of subtidal sandbank communities, any activity happening over the feature will compromise this target.

Given the unfavourable condition of the sandbank feature in the site and the target to restore the presence and spatial distribution of subtidal sandbank communities, **MMO conclude that impacts from removal of non-target species by demersal trawls on the sandbank feature may not be compatible with the conservation objectives of the site and may result in an adverse effect on site integrity.**

4.4.1.4 Impacts of demersal seines

When the ropes of a seine net are closed up in order to herd demersal fish, there is the potential for removal of epifauna.

Biotopes containing attached or sessile epifauna are considered sensitive to abrasion due to the removal of these non-target species (MBIEG, 2020). Observations in the North Sea show that seining caught species which are characteristic of the sandbank community of Inner Dowsing Race Bank North Ridge SAC. These include brittlestars (*Ophiura sp.*), queen scallop (*Aequipecten opercularis*) and edible crab (*Cancer pagarus*) (Waardenburg, 2017). Bioengineers such as bryozoa, for example *Flustra foliacea* found in the site, are also estimated to be sensitive to removal by demersal seining (Waardenburg, 2017).

VMS data indicates that demersal seining only took place over the sandbank feature in 2015. Due to the restore target for presence and spatial distribution of subtidal sandbank communities, any level of activity happening over the feature will compromise this target.

Given the unfavourable condition of the sandbank feature in the site and the target to restore the presence and spatial distribution of subtidal sandbank communities, **MMO conclude that impacts from removal of non-target species by demersal seines on the sandbank feature may not be compatible with the conservation objectives of the site and may result in an adverse effect on site integrity.**

4.4.1.5 Impacts of dredges

Dredges can cause significant amounts of by-catch for a range of non-commercially targeted species, the majority of which is discarded damaged, dying or dead (Howarth and Stewart, 2014). Dredges penetrate into the sediment and so may remove both infauna and epifauna from the sandbank feature.

Hinz *et al.* (2012) studied the environmental impact of different types of queen scallop fishing gears, including dredges. Results showed that traditional scallop dredges contained larger amounts of non-target species such as invertebrates than other gear types such as otter trawls (Hinz *et al.*, 2012). For example, clear negative effects were found for brittlestars, *Ophiura ophiura* (Hinz *et al.*, 2012). Species such as brittlestars, as well as other benthic invertebrates, are known to be key members of the sandbank feature of Inner Dowsing Race Bank North Ridge SAC.

VMS data indicates that dredging activity by over 12 m vessels over the sand bank feature has declined since 2014 with only a small number of records in recent years. The majority of activity takes place outside the boundary of the site and does not frequently take place over the sandbank feature. Expert opinion describes dredging is described to be sporadic and not been known to occur over the reef or sandbank features. Despite the activity levels being low, due to the restore target for presence and spatial distribution of subtidal sandbank communities, any activity happening over the feature will compromise with this target.

Given the unfavourable condition of the sandbank feature in the site and the target to restore the presence and spatial distribution of subtidal sandbank communities, **MMO conclude that impacts from removal of non-target species by dredges on the sandbank feature may not be compatible with the conservation objectives of the site and may result in an adverse effect on site integrity.**

4.4.2 *Sabellaria spinulosa* reef

Removal of non-target species refers to the removal of organisms associated with, and important to, *S. spinulosa* reef, and does not include the direct removal of *S. spinulosa* reef by fishing gears. Direct removal is covered by the abrasion assessment. The assessment for traps and anchored nets/line has been combined due to similarity in impacts.

4.4.2.1 Impacts of traps and anchored nets/lines

Gibb *et al.* (2014) reported that although evidence for ecological interaction between *S. spinulosa* and other species was limited, no evidence for significant biological effects from the removal of non-target species associated with *S. spinulosa* reef was identified.

There is some evidence that the stabilisation of sediments by the sand mason worm *Lanice conchilega* may facilitate formation of *S. alveolata* reefs which may also be possible for *S. spinulosa* (Gibb *et al.*, 2014). However, *L. conchilega* is very unlikely to be removed by static gears.

Removal of non-target species may in fact be beneficial through removal of predators such as the butterflyfish *Pholis gunnelis* and dragonet *Callionymus lyra*, common starfish *Asterias rubens* or competitors such as brittlestars *Ophiothrix fragilis* (Gibb *et al.*, 2014). Dense aggregations of the brittle star, *Ophiothrix fragilis*, have been suggested to compete with *S. spinulosa* for space and food and potentially to consume the gametes inhibiting recruitment (George & Warwick 1985 cited in Gibb *et al.*, 2014). However, the removal of predatory species within an ecosystem has been known to have equally detrimental impacts on ecosystem health and stability including tropic cascades (Pinnegar *et al.*, 2000)

As static fishing gears do not appear to remove species which are important to *S. spinulosa* reef, Gibb *et al.* (2014) classified *S. spinulosa* reef as not sensitive to removal of non-target species.

Considering the discussion above, **MMO conclude that impacts from removal of non-target species by traps and anchored nets/lines on the reef feature are compatible with the conservation objectives of the site will not result in an adverse effect on site integrity.**

4.4.3 Summary of removal of non-target species assessment

Given the evidence above, removal of non-target species is unlikely to hinder the targets to restore the presence, spatial distribution and abundance of reef communities. **MMO conclude that traps and anchored nets/lines alone are compatible with the conservation objectives of the site and will not result in an adverse effect on site integrity via this pressure (Table 32).**

The removal of non-target species by demersal trawls, demersal seines and dredges within the site may hinder the target to restore the presence and distribution of subtidal sandbank communities. **MMO conclude that demersal trawls, demersal seines and dredges alone may not be compatible with the conservation objectives of the site and may result in an adverse effect on site integrity via this pressure (Table 32).**

Traps and anchored nets/lines are unlikely to hinder the conservation objectives of reef and sandbank feature. **MMO conclude that traps and anchored nets/lines alone are compatible with the conservation objectives of the site and will not result in an adverse effect on site integrity via this pressure (Table 32).**

Table 32: Removal of non-target species assessment

Pressure	Feature	Favourable condition target	Gear type	Compatible with the conservation objectives?
Removal of non-target species	S. <i>spinulosa</i> reef	Restore the presence and spatial distribution of reef communities.	Traps	Yes
			Anchored nets/lines	Yes
	Sandbank	Restore the presence and spatial distribution of subtidal sandbank communities.	Traps	Yes
			Anchored nets/lines	Yes
			Demersal Trawl	No
			Demersal seines	No
			Dredges	No
			Traps	Yes
		Maintain the species composition of component communities.	Anchored nets/lines	Yes
			Demersal Trawl	Yes
			Demersal seines	Yes
			Dredges	Yes

4.5 Siltation rate changes (low) including smothering and changes in suspended solids (water clarity) on the sandbank feature

4.5.1 Impacts of demersal trawl, seines and dredges

Demersal towed gears such as demersal trawls, seines and dredges will generate a plume of suspended sediment as the gear is pulled across the seabed. This can result in increased suspended solids and siltation rates as sediment resettles, potentially impacting sandbank communities through smothering of organisms.

The amount of material brought into suspension is dependent on the gear being used and the seabed habitat (O'Neill and Summerbell, 2011). The sandbank feature consists of subtidal coarse sediment, subtidal mixed sediment and subtidal sand. Subtidal coarse and subtidal mixed sediment tend to produce less of a plume than subtidal sand. Finer sediments, such as muddy sand, have been shown to produce a plume similar to background levels of sedimentation (O'Neill and Summerbell, 2011).

ABPmer and Ichthys Marine (2015, 2015a) modelled sedimentation levels resulting from the use of demersal towed gears over sedimentary habitats in two sites in the Southern North Sea and estimated that for sandy sediments with a 20% silt fraction, the amount of sediment mobilised by a beam trawl through hydrodynamic drag equates to a sediment depth of between 3.4 mm and 9.7 mm (average across the gear footprint).

Dynamic sand communities and gravelly muddy sand communities are relatively high energy habitats, meaning that species will be adapted to high levels of sediment resuspension from natural oceanic processes (JNCC and Natural England 2013). As the sediment plume disperses to background levels, the main impacts will be immediately behind the head of the gear (O'Neill and Summerbell, 2011). Most organisms in this area will be affected more by the abrasion and penetration caused by the fishing gear which has been discussed in section 4.2.

VMS data shows that little activity related to demersal trawls, seines and dredges occurred across the sandbank feature between 2014 and 2018 (Figure 3 to Figure 8). The area of impact and Pr-values across the sandbank feature also remains low, for example for demersal trawls the combined result in a total gear footprint for the sandbank feature is 0.1 – 2.8 km² and Pr-values are 0.005 – 0.0092 between 2014 and 2018. No activity for dredging was recorded in 2014 and 2015 and the total gear footprint for the sandbank feature is 0.18 – 0.35 km² and Pr-values are 0.0006 – 0.001 between 2016 -2018. Seining activity only took place in 2015, with total gear footprint for the sandbank feature being 0.02 km² and a Pr-value of 0.0007.

4.5.2 Summary of siltation rate and changes in suspended solids assessment

Given the low levels of fishing using demersal towed gears over the sandbank features in this site, and the small level of sediment suspended by these gears, the effects on the communities related to siltation rate changes is likely to be negligible compared to sediment resuspension from natural processes associated with sandbank feature. Therefore, **MMO conclude that demersal trawls, seines and dredges alone are compatible with the conservation objectives of the site and will not result in an adverse effect on site integrity via this pressure (Table 33).**

Table 33: Siltation rate changes (low) and water clarity assessment

Pressure	Interest feature	Favourable condition target	Activity	Compatible with the conservation objectives?
Siltation rate changes (low), including smothering (depth of vertical sediment overburden) and changes in suspended solids (water clarity)	Sandbank	Restore the presence and spatial distribution of subtidal sandbank communities.	Demersal Trawl and Seines	Yes
			Dredges	Yes
		Restore the distribution of sediment composition across the feature (and each of its sub-features).	Demersal Trawl and Seines	Yes
			Dredges	Yes
		Maintain the species composition of component communities.	Demersal Trawl and Seines	Yes
			Dredges	Yes
		Maintain all hydrodynamic and physical conditions such that natural water flow and sediment movement are not significantly altered or prevented from responding to changes in environmental conditions.	Demersal Trawl and Seines	Yes
			Dredges	Yes

4.6 Part B conclusion

4.6.1 Fishing on sandbanks

The assessment of fishing pressures on the sandbank feature within the MMO portion of Inner Dowsing, Race Bank and North Ridge SAC has revealed that an adverse effect on site integrity cannot be ruled out where demersal trawl, demersal seine and dredging activities occur, specifically with the pressure they exert from abrasion and the removal of non-target species. As such **MMO conclude that management measures are required to restrict these activities over the sandbanks features within the MMO portion of the site.** Section 7 contains further details of these measures.

With the introduction of the aforementioned management measures, **MMO conclude that, when considered alone and at current levels, the remaining fishing activities (traps and anchored nets and lines), where occurring over the sandbank feature, are compatible with the conservation objectives of the site and will not result in an adverse effect on site integrity.**

4.6.2 Fishing on *Sabellaria spinulosa* reef

The assessment of fishing pressures on the reef feature within the MMO portion of Inner Dowsing, Race Bank and North Ridge SAC has revealed that an adverse effect on site integrity cannot be ruled out where traps and anchored nets/lines activities occur specifically in relation to the abrasion pressure. As such **MMO conclude that management measures are required to restrict these activities over the reef features within the MMO portion of the site.** Section 7 contains further details of these measures.

5. Part C assessment

This section assesses the effects of activities considered as compatible with the conservation objectives of Inner Dowsing, Race Bank and North Ridge SAC in combination with other relevant activities taking place which includes the following:

- fishing activity/pressure combinations which were excluded in Part A of this assessment as having no likely significant effect (Table 8);
- fishing interactions assessed in Part B but not resulting in adverse effect;
- plans and projects.

The MMO [SPIRIT](#) (SPatial InfoRmatlon Toolkit) system was used to check relevant activities that occur within, or adjacent to, the assessed site where there could be a pathway for disturbance. SPIRIT includes information on all activities for which MMO has received a marine licence application as well as the majority of infrastructure already in place in the marine environment. To determine plans and projects to be included in this part of the assessment, a distance of 5 km was selected as suitable to capture any potential source receptor pathways which could impact the site in combination with effects of the fishing activities assessed.

Demersal trawls, seines and dredges have been identified as a red risk interaction for the reef feature and therefore assessment was not required. In Part B it was identified that demersal trawls, seines and dredges required management for the sandbank feature and traps and anchored nets/lines activities required management for the reef feature to avoid adverse effects to site integrity. Therefore, these fishing activities will not be considered in Part C. Anchored nets/lines and traps for the sandbank feature are the other fishing activities occurring within 5 km of Inner

Dowsing, Race Bank and North Ridge SAC. Therefore, in-combination effects of anchored nets/lines and traps with other project/plans for the sandbank feature will be assessed in Part C.

5.1 Pressures exerted by fishing and plans and projects

In accordance with the methodology detailed above, the SPIRIT system identified military surface/firing danger areas, offshore windfarm construction, disposal sites, pipelines and submarine cables as potential plans or projects occurring within 5 km of Inner Dowsing, Race Bank and North Ridge SAC (Table 34).

Several recreational activities were identified using SPIRIT, including Royal Yachting Association (RYA) clubs, RYA offshore routes and RYA general boating areas. Inner Dowsing, Race Bank and North Ridge SAC is at least 1 km offshore with submerged features, it is highly unlikely that there will be any contact through these activities and the designated features. No additional fishing activities to those already assessed in Part B occur within 5 km of the Inner Dowsing, Race Bank and North Ridge SAC.

To identify the specific pressures that the above activities exert on the feature of this site MMO has used the Advice on Operations (AoO) section in NE and JNCC's conservation advice package for Inner Dowsing, Race Bank and North Ridge SAC. This required identified activities to be matched against the activity categories used in the conservation advice (Table 35).

Table 34: Plans and projects considered in combination with fishing activities included in this assessment

Relevant Activity	Description
Pipelines	Five pipelines run through the site
Submarine Cables	Several submarine cables run through the site
Well Heads	Two well heads are located within the site
Disposal Sites	Two open disposal sites ID: HU126 Racebank OWF, HU123 Sherringham shoal drillings
Offshore Wind Farm: Lincs Wind Farm Limited	Active/in operation
Offshore Wind Farm: Lynn Wind Farm Ltd	Active/in operation
Offshore Wind Farm: Inner Dowsing Wind Farm Ltd	Active/in operation
Offshore Wind Farm: Race Bank Wind Farm Ltd	Active/in operation
Hanson Aggregates Ltd, Westminster Gravels Limited, Dong Energy RB (UK) Limited	Two aggregate dredging licences: <ul style="list-style-type: none"> - MLA/2013/00336/3 area 106/3 valid until 2029 - MLA/2015/00452/5 area 515/1 valid until 2037 One other dredging licence - MLA/2015/00452/5 area Racebank Offshore Wind Farm valid until 2030
Lincs Offshore Wind Farm Q&M Licence, Lynn Offshore Wind Farm Q&M Licence, Lincshore Beach Renourishment Works	Two deposit licenses within Lincs, Lynn and Inner Dowsing Offshore Farms <ul style="list-style-type: none"> - MLA/2014/00276 valid until 2038 - MLA/2014/00291 valid until 2038 One disposal of dredged material licence within Mablethorpe to Skegness - MLA/2016/00014/1

Table 35: Categories from the AoO section that have been used to inform pressures information for identified activities and Amber and Green fishing activities.

Name of Activity	NE AoO Operation	Activity
Pipelines	Oil, gas and carbon capture storage	Pipelines
Submarine Cables	Cables	Power cable: operation and maintenance
Well Heads	Oil gas and carbon capture storage	Oil and gas production
Disposal Sites	Ports and harbours (construction and maintenance)	Capital dredging disposal Maintenance dredging disposal
Offshore Wind Farms	Electricity from renewable energy sources	Offshore wind: operation and maintenance, Offshore wind: during construction
Aggregate Dredging	Aggregates extraction	Aggregate dredging
Demersal Trawl	Fishing	Demersal trawl
Dredges	Fishing	Dredges
Traps	Fishing	Traps
Anchored Nets	Fishing	Anchored nets/lines

Information in the Inner Dowsing, Race Bank and North Ridge SAC conservation advice package was used to determine which pressure-feature interaction to include in this part of the assessment.

A list of pressures has been collated from fishing activity, and it is only those pressures that have been discussed below. Equally if a multiple plans or projects give off a pressure that fishing does not contribute towards, those pressures are not within the scope of this assessment.

All pressure-feature interactions from fishing other than those identified as “Not Relevant” (the evidence base suggests that there is no interaction of concern between the pressure and the feature OR the activity and the feature could not interact) have been considered.

From these considerations, Table 36 details the pressures exerted by military firing activity; power cables: laying, burial and protection and operation and maintenance; telecommunication cables: operation and maintenance; offshore wind: during construction and operation and maintenance; disposal sites; pipelines; traps fishing activities; and anchored net/line fishing activities. Pressures highlighted green have been screened out as not requiring further consideration in this assessment as they are not exerted by the traps and anchored net/line fishing activities occurring within Inner Dowsing, Race Bank and North Ridge SAC.

Table 36 also indicates pressures which are exerted by each activity (Y – pressure exerted, N – pressure not exerted).

Table 36: Pressures exerted by fishing and non-fishing activities occurring in Inner Dowsing, Race Bank and North Ridge SAC. Non fishing pressures similarly exerted by anchored nets/lines and traps require further assessment and are highlighted in red.

Pressure	Telecommunication cable: operation & maintenance	Power cable: operation & maintenance	Power cable: laying, burial & protection	Offshore wind: during construction	Offshore wind: Operation & maintenance	Sea surface military activity	Dredge and soil disposal	Pipelines	Anchored nets/lines	Pots/Traps
Abrasion/disturbance of the substrate on the surface of the seabed	Y	Y	Y	Y	Y	Y	Y	Y	Y	Y
Changes in suspended solids (water clarity)	Y	Y	Y	Y	Y	Y	Y	N	N	N
Deoxygenation	N	Y	Y	N	N	Y	Y	Y	Y	Y
Electromagnetic changes	N	Y	N	N	N	N	N	N	N	N
Habitat structure changes - removal of substratum (extraction)	N	N	Y	Y	Y	Y	N	Y	N	N
Hydrocarbon & PAH contamination.	Y	Y	Y	Y	Y	Y	Y	Y	Y	Y
Introduction of other substances (solid, liquid or gas)	N	N	N	Y	Y	Y	Y	Y	N	N
Introduction or spread of invasive non-indigenous species (INIS)	Y	Y	Y	Y	Y	Y	Y	Y	Y	Y
Litter	Y	Y	Y	Y	Y	Y	N	Y	Y	Y
Nutrient enrichment	N	Y	Y	N	N	Y	Y	Y	N	N
Organic enrichment	N	N	N	N	N	N	N	N	Y	Y
Penetration and/or disturbance of the substrate below the surface of the seabed, including abrasion	Y	Y	Y	Y	Y	Y	N	Y	Y	Y
Physical change (to another seabed type)	Y	Y	Y	Y	Y	N	Y	Y	N	N

Physical change (to another sediment type)	N	N	N	N	N	N	Y	N	N	N
Radionuclide contamination	N	N	N	N	N	N	Y	N	N	N
Removal of non-target species	N	N	N	N	N	N	N	N	Y	Y
Removal of target species	N	N	N	N	N	N	N	N	Y	Y
Siltation rate changes (high), including smothering (depth of vertical sediment overburden)	N	N	Y	Y	Y	N	Y	N	N	N
Siltation rate changes (low), including smothering (depth of vertical sediment overburden)	Y	Y	Y	Y	Y	Y	Y	Y	N	N
Synthetic compound contamination (incl. pesticides, antifoulants, pharmaceuticals).	Y	Y	Y	Y	Y	Y	Y	Y	Y	Y
Temperature changes - local	N	Y	N	N	N	N	N	N	N	N
Transition elements & organo-metal (e.g. TBT) contamination.	Y	Y	Y	Y	Y	Y	Y	Y	Y	Y
Vibration	Y	Y	Y	Y	Y	N	N	Y	N	N
Water flow (tidal current) changes, including sediment transport considerations	Y	Y	Y	Y	Y	N	N	Y	N	N
Wave exposure changes - local	N	N	N	N	N	Y	N	N	N	N

5.2 In-combination pressure discussion for remaining pressures

5.2.1 Abrasion/disturbance of the substrate on the surface of the seabed AND Penetration and/or disturbance of the substrate below the surface of the seabed, including abrasion

This pressure is relevant for all anchored nets/lines in addition to traps and all plans and projects. Sensitivity of the Inner Dowsing, Race Bank and North Ridge SAC conservation feature to physical damage from static gears and anchored nets/lines is through surface abrasion from pots, through deployment, movement of gear on the benthos due to strong tidal current and storm activity; and as the gear is dragged along the seafloor during retrieval. However, the Inner Dowsing, Race Bank and North Ridge SAC sandbank feature is not considered sensitive to the pressure associated with sea surface military activity as it is derived from propellers and ship movements causing scour around berth pockets and channel margins which does not occur the site.

Decommissioning, burial, protection and maintenance of submarine cables as well as maintenance of offshore windfarms may have impacts due to the physical disruption of the sediment. Vessels associated with these activities will anchor or use jack-up legs which will penetrate into the sediment. Anchor handling of vessels within the anchor corridor will cause disturbance up to 1 km on each side of the cable through embedment in the substrate as well as subsequent scouring during retrieval, although the pressure exerted will be low when smaller anchors are used. Despite this, the frequency of maintenance to existing cables will be low. Additionally, this is a licensable activity, if there was a positive determination on applications for maintenance, licence conditions would be put in place to mitigate against any significant impacts to the features of the site. Therefore, it is unlikely that operation and maintenance of existing submarine cables will have a significant in combination impact with fishing and other activities via this pressure.

Pipelines are predicted to cause abrasion and penetration disturbance to a maximum of 100 m either side of the pipelines. Beyond this, disturbance may be caused through maintenance of the pipeline when anchors are used to secure vessels. There are multiple pipelines which intersect the site, these are mostly towards the northern boundary. Given that these pipelines are already in place, there are no potential in combination impacts through installation. Maintenance of pipelines is a licensable activity, therefore licence conditions would be put in place to mitigate against any significant impacts to the features of the site. Consequently, it is unlikely that pipelines will have a significant in combination impact with fishing and other activities via this pressure.

There are currently 3 aggregate dredging licences which are valid up until the years 2029, 2030 and 2037. Although the activity has the potential to cause structural

damage by removing the surface layers of the sediment, the licenses have a condition requiring applicants to ensure no extraction of materials representing Annex 1 sandbank habitat takes place.

MMO conclude that abrasion/disturbance and penetration pressures associated with anchored nets/lines and traps, in combination with the plans/projects/activities occurring in the site are compatible with the conservation objectives of the site and will not result in an adverse effect on site integrity.

5.2.2 Deoxygenation

This pressure is relevant for traps and all anchored nets/lines in combination with submarine cables, well heads and pipelines.

Discards are not spatially concentrated at this site and it is not an area of low flow so the conditions for localised hypoxia or anoxia of the seabed are not present. Given the size and dynamics of the site the combined effects of fishing and plans or projects would not reduce oxygen concentration over a prolonged period, capable of affecting the Water Framework Directive³⁷ status.

Modern equipment and techniques reduce the re-suspension of sediment during cable burial, repair and removal, however, increases in suspended sediment may occur (OSPAR, 2012). The magnitude of this depends on the silt fraction, the equipment used and background levels (OSPAR, 2012). With regards to impacts caused during maintenance of cables, the frequency of this activity will be low. Furthermore, this is a licensable activity and so licence conditions would be put in place to mitigate against any significant impacts to the features of the site. Therefore, it is unlikely that operation and maintenance of existing submarine cables will have a considerable in combination impact with fishing and other activities via these pressures.

With regards to pipelines and well heads, seabed currents and the type of sediment will affect the accumulation and scouring of sediment around these structures. Once the structures have been scoured to their equilibrium depth, there will be an absence of sediment for further scouring therefore limiting resuspension and ultimately deoxygenation.

MMO conclude that this pressure associated with anchored nets/lines and traps, in combination with the plans/projects/activities occurring in the site are compatible with the conservation objectives of the site and will not result in an adverse effect on site integrity.

³⁷ https://ec.europa.eu/environment/water/water-framework/index_en.html

5.2.3 Hydrocarbon & PAH contamination AND Transition elements & organo-metal (e.g. TBT) contamination.

The primary route of chemicals of concern is via vessel oil and fuel and therefore covered by hydrocarbon and PAH contamination. Synthetic compound contamination is not considered further as these compounds are likely to originate from terrestrial sources.

This pressure is relevant for all gears in combination with all plans or projects and includes consideration of priority substances listed in Annex II of Directive 2008/105/EC.

Polycyclic aromatic hydrocarbons (PAH) in vessel oil and fuel are of environmental concern when released into the water. Fishing vessels of all gear types may contribute to this pressure in combination with military vessels. However, deliberate releases of oil or oil/water mixture from ships are prohibited within the North West European Waters Special Area, established by the International Maritime Organisation (IMO) under MARPOL Annex I in 1999³⁸. This area includes all waters around the UK and its approaches. While Navy vessels are exempt from MARPOL, they are expected to act in a manner consistent with MARPOL as far as is reasonable and practicable³⁹. Accidental discharges may occur, however significant releases are extremely rare. Releases of significant amounts of oil are typically from large shipping vessels and tankers. Sea surface military vessels are therefore unlikely to contribute considerably to the minor, existing impact from fishing vessels via this pressure.

Hydrocarbon and PAH contamination may occur through antifouling compounds like copper wash and TBT from ship coatings. However, fishing and MOD vessels comply with IMO standards for hull coatings and so are unlikely to contribute via this pathway.

Pipelines may be a source of hydrocarbon and PAH contamination. Additionally, cuttings from drilling operations and old cutting piles may contain organic-phase drilling fluids which may be disturbed during decommissioning of the pipelines (BEIS, 2019). However, results from surveys undertaken in other areas of the North Sea demonstrate very little contamination from heavy or trace metals or hydrocarbons, with the majority of samples reporting levels similar to background levels (BEIS, 2019). Therefore, in combination effects with other activities are unlikely to mean that fishing will have a significant impact via this pressure.

³⁸ <https://www.imo.org/en/OurWork/Environment/Pages/OilPollution-Default.aspx>

³⁹ <http://www.mar.ist.utl.pt/mventura/Projecto-Navios-I/IMO-Conventions%20%28copies%29/MARPOL.pdf>

Through licensing processes all material disposed at sea would have passed Cefas testing to be below Action Level 2.

MMO conclude that this pressure associated with anchored nets/lines and traps, in combination with the plans/projects/activities occurring in the site are compatible with the conservation objectives of the site and will not result in an adverse effect on site integrity.

5.2.4 Introduction or spread of invasive non-indigenous species (INIS)

This pressure is relevant for anchored nets/lines and traps in combination with submarine cables, offshore wind farms, well heads and pipelines.

Aquatic organisms may be transferred to new locations through biofouling which takes place on all craft, even if recently cleaned or anti-fouled (IMO, 2012). Ballast water of vessels may also be a vector for transferral (OSPAR, 2009). Military vessels, and vessels associated with installation, operation or maintenance of submarine cables, offshore windfarms and pipelines may therefore transport organisms.

With regards to submarine cables, offshore wind farms and pipelines, the artificial structures themselves may encourage the spread of INIS. It has been demonstrated that new artificial substrata offer opportunities for INIS to enter an area, or if already present, allows them to expand their population size and hence strengthen their strategic position (Kerckhof *et al.*, 2011). This is particularly important for the obligate intertidal hard substrata species, for which offshore habitat is rare to non-existing (Kerckhof *et al.*, 2011). Despite this, numerous monitoring for the construction of other wind farms have shown no presence of INIS associated with infrastructure (Forewind, 2014). This pressure is unlikely to have a significant in combination impact with fishing.

For fishing vessels, ballast water is the principal vector for invasive non-indigenous species. VMS data shows that the majority of fishing vessels visiting the site are smaller than 45 m in length which means they use solid ballast. Additionally, for vessels using ballast water, the International Convention for the Control and Management of Ships' Ballast Water and Sediments⁴⁰ requires them to manage ballast water and sediments to a certain standard to prevent the spread of organisms. This means that the contribution of fishing activities to this pressure is minimal. Therefore, in combination effects with other activities are unlikely to mean that fishing will have a significant impact via this pressure.

⁴⁰ [https://www.imo.org/en/About/Conventions/Pages/International-Convention-for-the-Control-and-Management-of-Ships%27-Ballast-Water-and-Sediments-\(BWM\).aspx](https://www.imo.org/en/About/Conventions/Pages/International-Convention-for-the-Control-and-Management-of-Ships%27-Ballast-Water-and-Sediments-(BWM).aspx)

MMO conclude that this pressure associated with anchored nets/lines and traps, in combination with the plans/projects/activities occurring in the site are compatible with the conservation objectives of the site and will not result in an adverse effect on site integrity.

5.2.5 Litter

This pressure is relevant to anchored net/line and traps fishing activities and all plans/ projects with the exception of disposal sites.

For installation, operation and maintenance of submarine cables, offshore wind farms, military activities and pipelines, this pressure is relevant to the vessels associated with the activity. Vessels may release litter accidentally, due to inappropriate storage, or deliberately (Potts & Hasting, 2011; Lozano & Mouat, 2009). Litter may include pallets, strapping bands and drums or materials related to the construction of infrastructure. Similarly, military vessels may also contribute to marine litter via accidental or deliberate releases.

Litter released by fishing vessels may include galley waste, fish boxes, floats/buoys, nets, ropes, weights and microplastic particles resulting from disintegration of plastic gear (Lozano & Mouat, 2009). These may cause damage to benthic habitats through abrasion or ghost fishing.

All vessels, bar those attaining to the Navy, adhere to MARPOL requirements which prohibit the discharge of plastics. While exempt, Navy vessels are expected to act in a manner consistent with MARPOL so far as is reasonable and practicable⁴¹ and therefore releases of litter is likely to be minimal from all vessels.

The exposure of this site means that any marine litter that does occur, is unlikely to persist in the same location long enough to cause damage to the sand bank feature, for example via abrasion. Therefore, it is unlikely that this pressure will be significant when considered in combination with non-fishing activities.

MMO conclude that this pressure associated with anchored nets/lines and traps, in combination with the plans/projects/activities occurring in the site are compatible with the conservation objectives of the site and will not result in an adverse effect on site integrity.

⁴¹ <http://www.mar.ist.utl.pt/mventura/Projecto-Navios-I/IMO-Conventions%20%28copies%29/MARPOL.pdf>

5.2.6 Organic enrichment

This pressure is relevant for traps, all anchored nets/lines but is not exerted by any plans or projects. Degraded remains from these fishing gears in combination will not result in significant impacts to the site as the tidal range and water movement would not allow levels to reach the pressure benchmark.

MMO conclude that this pressure associated with anchored nets/lines and traps, in combination with the plans/projects/activities occurring in the site are compatible with the conservation objectives of the site and will not result in an adverse effect on site integrity.

5.3 In-combination conclusion

MMO conclude, taking into account the introduction of management areas for demersal trawls, seines and dredges outlined in section 7, that fishing activities in combination with other relevant activities are not adversely affecting the site integrity of Inner Dowsing, Race Bank and North Ridge SAC.

6. Assessment result

6.1 Fishing alone

MMO consider that for sandbank feature the impacts from bottom-towed gear (demersal trawl, demersal seine and dredging) are not compatible with the conservation objectives of the site and may result in adverse effect on site integrity. As a red risk interaction, the impact of bottom towed gear (demersal trawl, demersal seine and dredging) over area to be managed as reef is not compatible with the conservation objectives of the site and may result in an adverse effect of site integrity. **MMO therefore conclude that management measures are required to restrict these activities over both the sandbank and reef features within the MMO portion of the site.**

MMO consider that for the sandbank feature traps, anchored nets/lines will not result in adverse effect on site integrity.

MMO consider that for the *S. spinulosa* reef feature the impacts from traps, anchored nets/lines are not compatible with the conservation objectives of the site and may result in adverse effect on site integrity. **MMO therefore conclude that management measures are required to restrict these activities over the reef features within the MMO portion of the site.**

6.2 In-combination

For the sandbank feature, when pressures from traps and anchored nets/lines fishing activities were combined and considered alongside pressures from the potential non fishing activities taking place, none were identified which may result in adverse effect on site integrity in combination. **MMO therefore conclude that assessed fishing activities from traps and anchored nets/lines, in-combination with other known activities, are compatible with the conservation objectives of the site and are not causing an adverse effect on site integrity.**

7. Management options

Option 0: No fisheries restrictions. Introduce a monitoring and control plan within the site.

Option 1: Remove/avoid pressures (whole site prohibition). Demersal trawls, demersal seines and dredges, traps and anchored nets/lines will be prohibited in all areas of the site.

Option 2 (preferred option): Reduce/limit pressures (whole feature prohibition). Zoned management will be introduced to prohibit use of all bottom towed fishing gear and static gear over the features of the site to ensure the achievement of the conservation objectives.

Option 3: Reduce/limit pressures (zoned feature prohibition). Zoned management will be introduced to prohibit the use of bottom towed fishing gear and static gear over the most sensitive parts of the features of the site.

Option 4: Introduce a voluntary agreement.

Option 0 is not sufficient to protect Inner Dowsing, Race Bank and North Ridge SAC due to likely adverse effects to site integrity from fishing with gears that interact with the seabed. The risks to the site from damaging activities would not be addressed and that MMO duties under the Conservation of Habitats and Species Regulations 2017 and the Conservation of Offshore Marine Habitats and Species Regulations 2017 would not be met.

Option 1 would prohibit fishing gears across the whole SAC including areas in which gear feature interactions will not have an adverse effect on the site. Although Option 1 would ensure that there are no impacts from fishing over the whole site through a whole site prohibition for demersal trawls, demersal seines and dredges, traps and anchored nets/lines. This would limit the use of sustainable fishing practices over

area of the site where there is no designated feature, causing an undue burden on the fishing industry and legitimate users of the sea.

Option 2 is the most appropriate option for the site to enable sustainable use of our oceans and protect the features. An MMO byelaw will be introduced to ensure the risk of adverse effect on site integrity is removed by prohibiting bottom towed fishing gear over the sandbank and reef features and prohibiting static gears over the reef features. The boundaries of the proposed management area will include an appropriate buffer zone of 80 – 150 m to prevent direct damaging physical interactions between a fishing activity and the designated features (Figure 2).

Option 3 would prohibit bottom towed and static gear fishing over a proportion of the features of the site. Bottom towed gear would be prohibited over all reef features as these features are sensitive to impacts from these gear types (red risk interaction) but would only be prohibited over certain parts of the sandbank feature. Static gears (including pots, traps and anchored nets/lines) would similarly be prohibited over a proportion of the area to be managed as reef.

MMO concluded that it was not possible to identify areas of sandbank where an adverse effect on site integrity from bottom towed fishing could be ruled out. The extent of area to be managed as reef across the whole site totals 15 km² made up of smaller patches of reef as well as the larger Silver Pitt reef complex making effective enforcement of a zoned management approach problematic. As the levels of fishing occurring within this site are likely to be an underestimate, especially with regards to the under 12 m fleet, MMO makes decisions in line with the precautionary principle and, in this case, determine that this option is not suitable as it will not provide the level of protection required due to the high levels of uncertainty surrounding activity levels.

Option 4 would involve the development of voluntary codes of practice to protect features. MMO has considered this option in light of Better Regulation principles⁴², which require that new regulation is introduced only as a last resort. However, the government's expectation is that management measures for commercial fishing in marine protected areas (MPAs) should be implemented through statutory regulation to ensure adequate protection is achieved.

MMO strives to avoid any unnecessary costs to the fishing industry, financial or otherwise in the development of management measures, however, MMO has legal obligations in relation to SACs under the Conservation of Habitats and Species Regulations, 2017, Regulation 9, and Conservation of Offshore Marine Habitats and Species Regulations 2017 Regulation 6⁴³ to ensure compliance with the Habitats

⁴²

https://assets.publishing.service.gov.uk/government/uploads/system/uploads/attachment_data/file/317555/betterregulationassessment2014.pdf

⁴³ <https://www.legislation.gov.uk/ukxi/2017/1013/contents/made>

Directive⁴⁴. Of particular relevance to marine conservation is section 6(2): to avoid the deterioration of habitats and disturbance of designated species. The proposed management measure outlined in the assessment are deemed necessary to meet this duty.

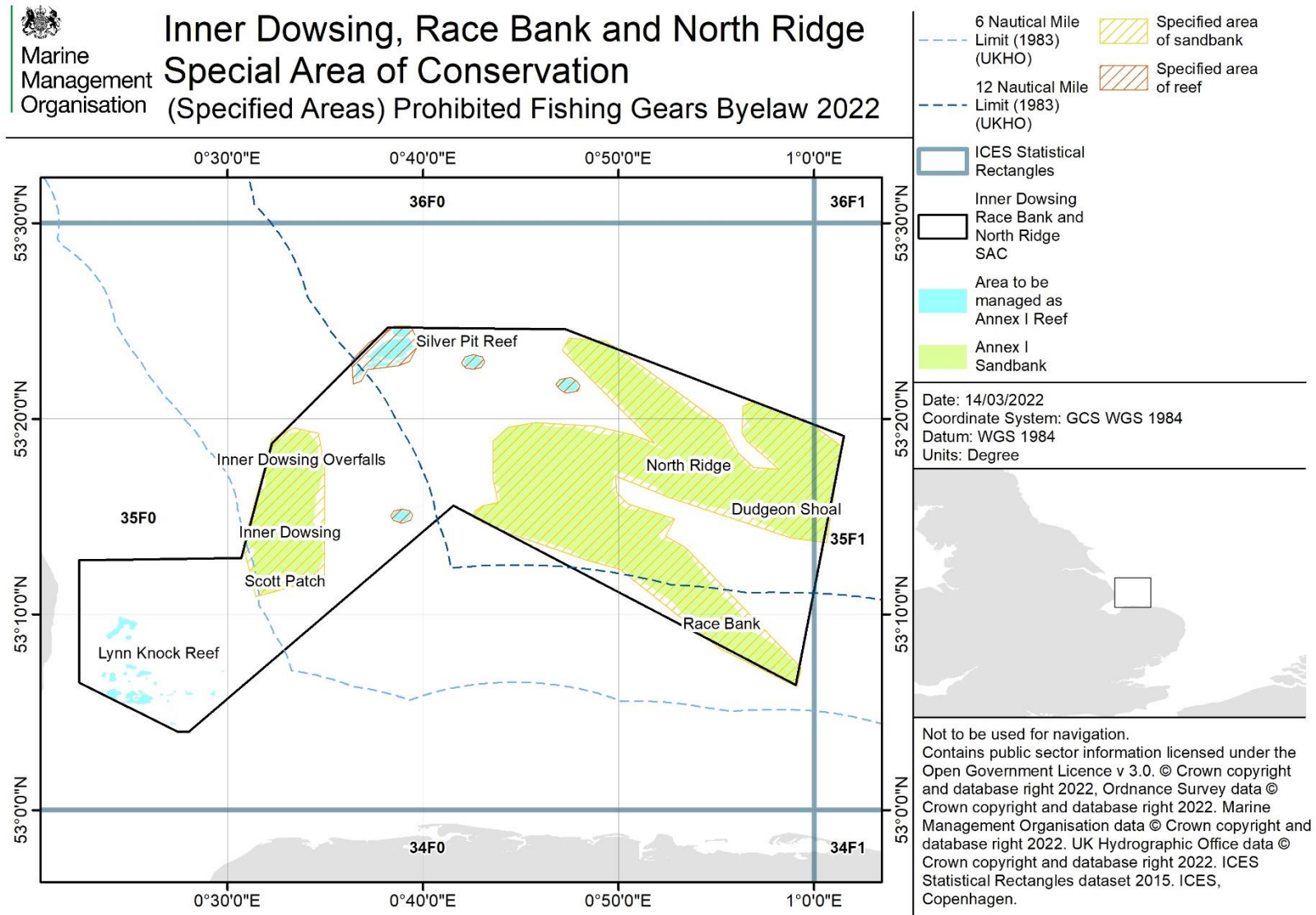
Marine Plans

Inner Dowsing, Race Bank and North Ridge SAC lies within the East Marine Plan Area. The East Marine Plan⁴⁵ was adopted in 2014. Management decisions will be compliant and made in accordance with relevant policies. Consideration of policies will be detailed in the regulatory triage assessment which will accompany the proposed management.

⁴⁴ <https://eur-lex.europa.eu/legal-content/EN/TXT/?uri=CELEX:31992L0043>

⁴⁵ <https://www.gov.uk/government/collections/east-marine-plans>

Figure 25: Management areas for the Inner Dowsing, Race Bank and North Ridge SAC.



8. Review of this assessment

MMO will review this assessment every five years, or earlier if significant new information is received. Such information could include:

- updated conservation advice;
- updated advice on the condition of the feature;
- considerable change in activity levels.

To coordinate the collection and analysis of information regarding fishing activity, and to ensure that any required management is implemented in a timely manner, a monitoring and control plan will be implemented for this site. This plan will be developed in line with the MMO Monitoring and Control Plan framework.

Monitoring of fishing activity will occur through a combination of surface surveillance and ongoing monitoring of VMS and landings data. Should activity levels, including those fishing activities not currently considered a risk to the features, increase considerably or in a manner that could affect the site features, this will trigger further investigation into the level and distribution of the activity, including consultation with Natural England and JNCC regarding current site condition. Any subsequent evidence gathered will be used to assess the need for further management measures.

Monitoring will be recorded through annual MPA reporting. Inner Dowsing, Race Bank and North Ridge SAC is categorised as Tier 2 which means an individual report is produced by MMO's Marine Conservation Team for this site annually. The report includes VMS data for fishing activity over the reporting period and a 5-year period as well as information on inspected/observed activities, intelligence and non-compliant activity (if applicable). Coastal questionnaires are completed by local MMO officers regarding any changes in activity within the site. This will act as an early warning system for potential negative impacts on the site. If the report determines that a change in fishing activity is a risk to the conservation objectives of the site, an assessment of the site will be triggered regardless of whether a review is due. An increase in fishing activity above levels considered in this assessment will initiate discussion with Natural England and JNCC following the annual MPA report.

Possible management measures include a MMO emergency byelaw, which can be implemented immediately for up to 12 months, or a (non-emergency) MMO byelaw which would be subject to public consultation before implementation.

An overview of the monitoring and control process is illustrated in Annex 3.

9. Conclusion

MMO has had regard to best available evidence and through consultation with relevant advisors and the public, concludes that, provided that appropriate management measures for the fishing activities identified above are implemented, all remaining fishing activities are compatible with the conservation objectives of this marine protected area.

MMO has determined therefore, that Option 2 is the preferred option to ensure protection of the site. An MMO byelaw has been proposed to ensure the risk of adverse effect on site integrity is removed by prohibiting bottom towed fishing gear over the sandbank and reef features and prohibiting static gears over the reef features (Figure 2).

10. References

- ABPmer and Ichthys Marine (2015). Supporting Risk-Based Fisheries Assessments for MPAs, Assessment of Beam Trawling Activity in North Norfolk Sandbanks and Saturn Reef SCI. ABPmer Report No. R.2551A. A report produced by ABPmer and Ichthys Marine Ecological Consulting Ltd. for National Federation of Fishermen's Organisations, December 2015.
- ABPmer and Ichthys Marine (2015a). Supporting Risk-Based Fisheries Assessments for MPAs, Assessment of Otter Trawling Activity in Margate and Long Sands SCI. ABPmer Report No. R.2551C. A report produced by ABPmer for National Federation of Fishermen's Organisations, December 2015.
- Blasdale, T, Duffy, M, Enever, R, Fisher, R, Lannin, FA, Marubini, F, Stevens, H, Tasker, M (2011). Advice from the Joint Nature Conservation Committee and Natural England with regard to fisheries impacts on Marine Conservation Zones.
- Bolam, SG, Coggan, RC, Eggleton, J, Diesing, M, Stephens, D (2014). Sensitivity of macrobenthic secondary production to trawling in the English sector of the Greater North Sea: A biological trait approach. *Journal of Sea Research*. Vol 85, 162–177
- Centre for Environment Fisheries and Aquaculture Sciences (Cefas). (2013). Benthic Survey of Inner Dowsing, Race Bank and North Ridge cSAC, and of Haisborough, Hammond and Winterton cSAC: Natural England (NE).
- Centrica Energy (2009). Race Bank Offshore Windfarm Environmental Statement Volume 1 Offshore: Centrica Energy.
- des Clers, S (2010). Development of the FisherMap methodology to map commercial fishing grounds and fishermen's knowledge. Seafish Report No. SR634. Available [here](#).
- des Clers, S., Lewin, S., Edwards, D., Searle, S., Lieberknecht, L. and Murphy, D. (2008). FisherMap - Mapping the Grounds: recording fishermen's use of the seas. Final Report. A report published for the Finding Sanctuary project. 58pp.
- Drake, JM, Lodge DM (2004). Global hotspots of biological invasions: evaluating options for ballast-water management. *Proceedings of the Royal Society, Series B* 271: 575-580.
- Chiappone, M., White, A., Swanson, D. W. and Miller, S. L. (2002). Occurrence and biological impacts of fishing gear and other marine debris in the Florida Keys. *Marine Pollution Bulletin*, 44(7), 597-604.
- Coleman, RA, Hoskin, MG, von Carlshausen, E, Davis, CM (2013). Using a no-take zone to assess the impacts of fishing: Sessile epifauna appear insensitive to

environmental disturbances from commercial potting. *Journal of Experimental Marine Biology and Ecology*. Vol 440, 100-107.

Collie, JS, Hall, SJ, Kaiser, MJ, Poiner, IR (2000). A quantitative analysis of fishing impacts on shelf-sea benthos. *Journal of animal ecology*. Vol 69, 785-798.

Cooke, SJ, Cowx, IG (2006). Contrasting recreational and commercial fishing: Searching for common issues to promote unified conservation of fisheries resources and aquatic environments. *Biological Conservation*. 128: 93-108.

Cunningham, P, Hawkins, S, Jones, H, Burrows, M (1984). The geographical distribution of *Sabellaria alveolata* (L.) in England, Wales and Scotland, with investigations into the community structure of, and the effects of trampling on *Sabellaria alveolata* colonies. Report to the Nature Conservancy Council from the Department of Zoology, Manchester University, Manchester.

De Groot, SJ, Lindeboom, HJ (1994). Environmental impact of bottom gears on benthic fauna in relation to natural resources management and protection of the North Sea. Netherlands Institute for Sea Research. NIOZ-Rapport 1994- 11, RIVO-DLO report CO26/94.

Department for Business, Energy and Industrial Strategy (BEIS) (2019). Record of the habitats regulations assessment undertaken under Regulation 5 of the Offshore Petroleum Activities (Conservation of Habitats) Regulations 2001 (as amended). Dogger Bank SAC Oil and Gas Decommissioning Strategic HRA.

Dernie, KM, Kaiser, MJ, Warwick, RM (2003). Recovery rates of benthic communities following physical disturbance. *Journal of Animal Ecology* Vol 72, 1043–1056.

Eigaard, OR, Bastardie, F, Breen, M, Dinesen, GE, Hintzen NT, *et al.*, (2016a). Estimating seabed pressure from demersal trawls, seines, and dredges based on gear design and dimensions. *ICES Journal of Marine Science* 73: i27-i47.

Eigaard, OR, Bastardie, F, Breen, M, Dinesen, GE, Hintzen, NT, Laffargue, P, Mortensen, L. O. *et al.*, (2016b). A correction to “Estimating seabed pressure from demersal trawls, seines and dredges based on gear design and dimensions”. *ICES Journal of Marine Science*, 73: 2420–2423.

Eno NC, Frid DLJ, Hall K, Ramsay K, Sharp RAM, Brazier DP, Hearn S, Dernie KM, Robinson KA, Paramore OAL, Robinson LA (2013). Assessing the sensitivity of habitats to fishing: from seabed maps to sensitivity maps. *Journal of Fish Biology*. doi:10.1111/jfb.12132, available online at wileyonlinelibrary.com.

Eno, NC, MacDonald, DS, Kinnear, JAM, Amos, SC, Chapman, CJ, Clark, RA, Bunker, F StP, Munro, C (2001). Effects of crustacean traps on benthic fauna. ICES Journal of Marine Science, 58, 11-20.

ENTEC UK LTD (2008) SAC selection assessment: Outer Wash Sandbanks. Report to Natural England as part of Contract FST20-18-030.

Foden, J, Rogers, SI, Jones, AP (2010). Recovery of UK seabed habitats from benthic fishing and aggregate extraction – towards a cumulative impact assessment. Marine Ecology Progress series. Vol. 411, 259–270.

Fonteyne, R (2000). "Physical impact of beam trawls on seabed sediments", in "The Effects of Fishing on Non-Target Species and Habitats: Biological, Conservation and Socio-Economic Issues", pp. 15 – 36. Ed. by Kaiser M. J. de Groot S. J. Fishing News Books, 399 pp.

Forewind (2014). Dogger Bank Teesside A & B Environmental Statement – Chapter 13 Fish and Shellfish Ecology. Forewind Document Reference F-OFC-CH-013 Issue 4.1.

Foster-Smith, RL and Hendrick, VJ (2003). *Sabellaria spinulosa* reef in the Wash and North Norfolk cSAC and its approaches: Part III, Summary of knowledge, recommended monitoring strategies and out-standing research requirements. English Nature Research Reports No.543.

Foster-Smith, RL and Sotheran, I (1999). Broad scale remote survey and mapping of sub-littoral habitats and biota of The Wash and the Lincolnshire and the North Norfolk coasts: Natural England.

Freese, L, Auster, PJ, Heifetz, J, Wing, BL (1999). Effects of trawling on seafloor habitat and associated invertebrate taxa in the Gulf of Alaska. Marine Ecology Progress Series, 182, 119–126.

Gall, SC, Rodwell, LD, Clark, S, Robbins, T, Attril, MJ, Holmes, LA, Sheehan, EV, (2020). The impact of potting for crustaceans on temperate rocky reef habitats: Implications for management. Marine Environmental Research 162 105-134.

Gibb N, Tillin HM, Pearce B, Tyler-Walters H (2014). Assessing the sensitivity of *Sabellaria spp.* to pressures associated with marine activities. JNCC report No. 504

Gislason, H (1994). Ecosystem effects of fishing activities in the North Sea. Marine Pollution Bulletin, 29, 520-527.

Grieve C, Brady DC, Hans Polet IR (2011). Best Practices for Managing, Measuring and Mitigating the Benthic Impacts of Fishing. Final Report to the Marine Stewardship Council.

Grieve C, Brady DC, Polet H (2014). Review of habitat dependent impacts of mobile and static fishing gears that interact with the sea bed – Part 1. Marine Stewardship Council Science. Vol. 2, 18–88.

Gubbay S, Knapman PA (1999). A review of the effects of fishing within UK European Marine Sites. Produced for: The UK Marine SACs Project. Available [here](#).

Glawys LI, Jennings S, Kaiser MJ, Davies TW, Hiddink JG (2014). Quantifying recovery rates and resilience of seabed habitats impacted by bottom fishing. *Journal of Applied Ecology*.

Hall K, Paramor OAL, Robinson LA, Winrow-Giffin A, Frid CLJ, Eno NC, Dernie KM, Sharp RAM, Wyn GC, Ramsay GC (2008). Mapping the sensitivity of benthic habitats to fishing in Welsh waters – development of a protocol; CCW (Policy Research) Report No: 8/12. 85pp.

Hendrick VJ, Foster-Smith RL, Davies AJ (2011). Biogenic Reefs and the Marine Aggregate Industry. Marine ALSF Science Monograph Series No. 3. MEPF 10/P149. (Edited by R. C. Newell & J. Measures). 60pp. ISBN: 978 0 907545 46 0.

Hervás A, Nimmo F, Southall T, Macintyre P (2012). The SSMO Shetland inshore brown & velvet crab, lobster and scallop fishery. MSC Public Certification Report.

Hiddink JG, Jennings S, Kaiser MJ, Queiros AM, Duplisea DE, Piet GJ (2006). Cumulative impacts of seabed trawl disturbance on benthic biomass, production, and species richness in different habitats. *Canadian Journal of Fisheries and Aquatic Science*. Vol. 63, 721–736.

Hinz, H., Murray, L. G., Malcolm, F. R., & Kaiser, M. J. (2012). The environmental impacts of three different queen scallop (*Aequipecten opercularis*) fishing gears. *Marine environmental research*, 73, 85-95.

Howarth, LM, Stewart, BD (2014). The dredge fishery for scallops in the United Kingdom (UK): effects on marine ecosystems and proposals for future management. Report to the Sustainable Inshore Fisheries Trust. Marine Ecosystem Management Report no. 5, University of York, 54 pp.

Humborstad, O-B, Nøttestad, L, Løkkeborg, S, Rapp, HT (2004). RoxAnn bottom classification system, sidescan sonar and video-sledge: spatial resolution and their use in assessing trawling impacts. *ICES Journal of Marine Science*, 61:53-63.

International Maritime Organisation (IMO) (2012). Guidance for minimising the transfer of invasive aquatic species and biofouling (hull fouling) for recreational craft.

Institute of Estuarine and Coastal Studies (IECS) (1995). Marine environmental baseline survey and assessment, Race Bank, east coast, UK. IECS unpublished report to the Environment Agency: Institute of Estuarine and Coastal Studies (IECS).

Institute of Estuarine and Coastal Studies (IECS) (1999). Biological baseline survey of Inner Dowsing (Area 439) & North Dowsing (Area 400). Report prepared for Entec UK for Hanson Aggregates Marine Ltd.: Institute of Estuarine and Coastal Studies (IECS).

JNCC and Natural England (2010). Inner Dowsing, Race Bank and North Ridge Selection Assessment Document Version 5.0. Available [here](#).

JNCC and Natural England (2013). Inner Dowsing, Race Bank and North Ridge candidate Special Area of Conservation. Formal advice under Regulation 35(3) of the Conservation of Habitats and Species Regulations 2010 (as amended), and Regulation 18 of the Offshore Marine Conservation Regulations (Natural Habitats, &c.) Regulations 2007 (as amended). Available [here](#).

Kaiser, MJ, Clarke, KR, Hinz, H, Austen, MCV, Somerfield, PK, Karakassis, I (2006). Global analysis of response and recovery of benthic biota to fishing. Marine Ecology Progress Series. Vol. 311, 1-14.

Kaiser, MJ, Collie, JS, Hall, SJ, Jennings, S, Roiner, IR (2002). Modification of marine habitats by trawling activities: prognosis and solutions. Fish and fisheries. Vol. 3, 114-136.

Kaiser, MJ, Edwards, DB, Armstrong, PJ, Radford, K, Lough, NEL, Flatt, RP, Jones, HD (1998). Changes in megafaunal benthic communities in different habitats after trawling disturbance. ICES Journal of Marine Science: Journal du Conseil. Vol. 55(3), 353-361.

Kaiser, MJ, Spencer, BE (1996). The effects of beam-trawl disturbance on infaunal communities in different habitats. Journal of Animal Ecology: 348-358.

Kerckhof, F, Degraer, S, Norro A, Rumes, B (2011). (Chapter 4. Offshore intertidal hard substrata: a new habitat promoting non-indigenous species in the Southern North Sea: an exploratory study.

Lambert, G., Jennings, S., Kaiser, M., Davies, T. and Hiddink, J., 2014. Quantifying recovery rates and resilience of seabed habitats impacted by bottom fishing. Journal of Applied Ecology, 51(5), pp.1326-1336.

Lambert, G, Murray, LG, Hiddink JG, Hinz H, Salomonsen, H, Moorhead, EK and Kaiser, MJ (2015). Impact of scallop dredging on benthic communities and habitat features in the Cardigan Bay Special Area of Conservation. Part III – Impact on epifauna. Fisheries & Conservation report No. 61, Bangor University. pp.61

Lart, W (2012). Fishing spatial-temporal pressures and sensitivities analysis for MPA Fishing Industry Collaboration Pilot FES 252: Report on Seafish workshop on the physical effects of fishing activities on the Dogger Bank.

Last, K., Hendrick, V., Sotheran, I., Foster-Smith, B., Foster-Smith, D. and Hutchison, Z (2012). Assessing the impacts of shrimp fishing on *Sabellaria spinulosa* reef and associated biodiversity in The Wash and North Norfolk SAC, Inner Dowsing Race Bank North Ridge SAC and surrounding areas. *Report for Natural England*. Available at:
https://www.dassh.ac.uk/dataDelivery/filestore/8/9/0/4_07c7622fb8c86d6/8904_427993f05ebf6ef.pdf

Lewin, W. C., Arlinghaus, R. and Mehner, T. (2006). Documented and potential biological impacts of recreational fishing: insights for management and conservation. *Reviews in Fisheries Science*, 14(4), 305-367.

Lewin, W.C., Weltersbach, M.S., Ferter, K., Hyder, K., Mugerza, E., Prellezo, R., Radford, Z., Zarauz, L. and Strehlow, H.V. (2019). Potential environmental impacts of recreational fishing on marine fish stocks and ecosystems. *Reviews in Fisheries Science & Aquaculture*, 27(3), pp.287-330.

Lewin, W. C., Weltersbach, M. S., Denfeld, G. and Strehlow, H. V. (2020). Recreational anglers' perceptions, attitudes and estimated contribution to angling related marine litter in the German Baltic Sea. *Journal of Environmental Management*, 272, 111062.

Limpenny DS, Foster-Smith RL, Edwards TM, Hendrick VJ, Diesing M, Eggleton JD, Meadows WJ, Crutchfield Z, Pfeifer S, Reach IS (2010). Best methods for identifying and evaluating *Sabellaria spp* and cobble reef. Aggregate Levy Sustainability Fund Project MAL0008. Joint Nature Conservation Committee, Peterborough, 134 pp. ISBN: 978-0-907545-33-0.

Lloret, J., Cowx, I.G., Cabral, H., Castro, M., Font, T., Gonçalves, J.M., Gordo, A., Hoefnagel, E., Matic-Skoko, S., Mikkelsen, E. and Morales-Nin, B. (2018). Small-scale coastal fisheries in European Seas are not what they were: ecological, social and economic changes. *Marine Policy*, 98, pp.176-186.

Løkkeborg, S (2005). Impacts of trawling and scallop dredging on benthic habitats and communities. FAO Fisheries Technical Paper. No. 472. Rome, FAO. 2005. 58p.

Lozano, RL and Mouat, J (2009). OSPAR Marine Litter in the North-East Atlantic Region.

MBIEG (2020). Assessing the physical impact of seining gear on protected features in UK waters. A report produced by The Marine Biological Association (MBA) on behalf of the Marine Biodiversity Impacts Evidence Group, Project No: ME6015, 71pp

Murray, LG, Lambert, GI, Bennell, J, Salomonsen, H, Kaiser, MJ (2015). Impact of scallop dredging on benthic communities and habitat features in the Cardigan Bay Special Area of Conservation. Part II – Physical environment. Fisheries & Conservation report No. 60, Bangor University. pp.23.

O'Neill, FG, Summerbell, K (2011). The mobilisation of sediment by demersal otter trawls. Original Research Article. Marine Pollution Bulletin, Volume 62, Issue 5, May 2011, Pages 1088-1097.

O'Neill, FG, Robertson, M, Summerbell, K, Breen, M, Robinson, LA (2013). Mobilisation of sediment and benthic infauna by scallop dredges. Marine Environmental Research, 90: 104–112.

OSPAR (2009). OSPAR assessment of the impacts of shipping on the marine environment.

OSPAR (2012). Guidelines on best environmental practice (BEP) in cable laying and operation.

Pearce, B, Taylor, J, Seiderer, LJ (2007). Recoverability of *Sabellaria spp* following aggregate extraction. Aggregate Levy Sustainability Fund MAL0027. Marine p.31 Ecological Surveys Limited, 24a Monmouth Place, Bath, BA1 2AY. 87pp. ISBN 978-0-9506920-1-2.

Pinnegar, J.K., Polunin, N.V.C., Francour, P., Badalamenti, F., Chemello, R., Harmelin-Vivien, M.L., Hereu, B., Milazzo, M., Zabala, M., D'anna, G. and Pipitone, C., (2000). Trophic cascades in benthic marine ecosystems: lessons for fisheries and protected-area management. Environmental Conservation, pp.179-200.

Polet, H. and Depestele, J. (2010). Impact assessment of the effects of a selected range of fishing gears in the North Sea. Stichting Noordzee, WNF.

Potts, T, Hasting, E (2011). Marine Litter Issues, Impacts and Actions. A report commissioned by Marine Scotland.

Queiros, AM, Hiddink, JG, Kaiser, MJ, Hinz, H (2006). Effects of chronic bottom trawling disturbance on benthic biomass, production and size spectra in different habitats. Journal of Experimental Marine Biology and Ecology. Vol. 335, 91–103.

Rayment, WJ (2001). Venerid bivalves in circalittoral coarse sand or gravel. In Tyler-Walters H and Hiscock K (eds) Marine Life Information Network: Biology and Sensitivity Key Information Reviews, [on-line]. Plymouth: Marine Biological Association of the United Kingdom. www.marlin.ac.uk/habitat/detail/63

Rijnsdorp, AD, Hiddink, JG, van Denderen, PD, Hintzen, NT, Eigaard, OR, Valanko, S, Bastardie, F, Bolam, SG, Boulcott, P, Egekvist, J, Garcia, C (2020). Different

bottom trawl fisheries have a differential impact on the status of the North Sea seafloor habitats. *ICES Journal of Marine Science*.

Rijnsdorp, AD (2015). Flyshoot fishery in relation to sea floor protection of the Frisian front and Central Oyster ground areas.pdf

Rijnsdorp, A (2013). BENTHIS (Benthic Ecosystem Fisheries Impact Study) Deliverable 1.1b. Benthic impact of the perspective of the fisheries. In: Report on benthic ecosystem processes and the impact of fishing gear: p.1-35;

Roberts, C, Smith, C, Tillin, H, Tyler-Walters, H (2010). Review of existing approaches to evaluate marine habitat vulnerability to commercial fishing activities. Environment Agency Report: SC080016/R3

Schernewski, G., Balciunas, A., Gräwe, D., Gräwe, U., Klesse, K., Schulz, M., Wesnigk, S., Fleet, D., Haseler, M., Möllman, N. and Werner, S. (2018). Beach macro-litter monitoring on southern Baltic beaches: results, experiences and recommendations. *Journal of Coastal Conservation*, 22(1), pp.5-25.

Schwinghamer, P, Gordon Jr, DC, Rowell, TW, Prena, J, McKeown, DL, Sonnichsen, G, Guigné, JY (1998). Effects of experimental otter trawling on surficial sediment properties of a sandy-bottom ecosystem on the Grand Banks of Newfoundland. *Conservation Biology*, 12: 1215–1222.

Sciberras, M, Hiddink, JG, Jennings, S, Szostek, CL, Hughes, KM, Kneafsey, B, Clarke, LJ, Ellis, N, Rijnsdorp, AD, McConnaughey, RA, Hilborn, R (2018). Response of benthic fauna to experimental bottom fishing: A global meta-analysis. *Fish and Fisheries*, 19(4), pp.698-715.

Stephenson, F, Mill, AC, Scott, CL, Polunin, NVC, Fitzsimmons, C, (2017). Experimental potting impacts on common UK reef habitats in areas of high and low fishing pressure. *ICES (Int. Counc. Explor. Sea) J. Mar. Sci.* 74 (6), 1648–1659.

Tillin, HM, Hull, SC, Tyler-Walters, H (2010). Development of a sensitivity Matrix (pressures-MCZ/MPA features). Report to the Department of Environment, Food and Rural Affairs from ABPMer, Southampton and the Marine Life Information Network (MarLIN) Plymouth: Marine Biological Association of the UK. Defra Contract No. MB12 Task 3A, Report No. 22.

Van Marlen, B, Wiegerinck, JAM, van Os-Koomen E, van Barneveld E (2013). Catch comparison of flatfish pulse trawls and a tickler chainbeam trawl. *Fisheries Research*. Vol. 151, 57– 69.

Vanstaen, K, Breen, P (2014). Defra project MB0117: Understanding the distribution and trends in inshore fishing activities and the link to coastal communities.

Vorberg, R (2000). Effects of shrimp fisheries on reefs of Sabellaria spp (Polychaeta). ICES Journal of Marine Science: Journal du Conseil, 57(5), 1416-1420.

Waardenburg, B (2017). Impact of demersal seine fisheries in the Natura 2000 area Dogger Bank. WWF Netherlands.

Walmsley, SF, Bowles A, Eno, NC, West, N (2015). Evidence for Management of Potting Impacts on Designated Features. Report Commissioned by Defra's Marine Biodiversity Impact Evidence Group. Reference: MMO1086

Annex 1 - MMO methodology

The need for assessment

In 2012, the Department for Environment, Food and Rural Affairs (Defra) announced a revised approach to the management of commercial fisheries in European marine sites (EMS)⁴⁶. The objective of this revised approach is to ensure that all existing and potential commercial fishing activities are managed in accordance with the provisions of Article 6 of the Habitats Directive⁴⁷. The revised approach was extended to include management of commercial fisheries in marine conservation zones (MCZ) in 2014⁴⁸.

This approach was being implemented using an evidence based, risk-prioritised, and phased basis. Risk prioritisation is informed by using a matrix of the generic sensitivity of the sub-features of EMS to a suite of fishing activities. These activity/sub-feature interactions have been categorised according to specific definitions, as red, amber, green or blue⁴⁹.

Activity/sub-feature interactions identified within the matrix as amber required a site-level assessment to determine whether management of activity is required to conserve site features. Activity/sub-feature interactions identified within the matrix as green also require a site level assessment if there are “in combination effects” with other plans or projects.

Site-level assessments are carried out in a manner consistent with the requirements of Article 63 of the Habitats Regulations for EMS and the requirements of section 126 of the Marine and Coastal Access Act 2009 for MCZ. For EMS the assessments will determine whether, in light of the site’s conservation objectives, fishing activities are having an adverse effect on the integrity of the site. For MCZ the assessments will determine whether there is a significant risk of fishing activities hindering the conservation objectives of the site.

Assessment process

The fisheries assessments have three stages:

Part A: A coarse assessment using generic sensitivity information to identify which fishing activities can be discounted from further assessment (Part B) as they are not taking place or not a significant concern.

Part B: An in-depth analysis to assess the effects of remaining pressures on the features of the site

Part C: An in-combination assessment between all fishing and non-fishing activities occurring.

⁴⁶ www.gov.uk/government/publications/revised-approach-to-the-management-of-commercial-fisheries-in-european-marine-sites-overarching-policy-and-delivery

⁴⁷ Council Directive 92/43/EEC on the Conservation of natural habitats and of wild fauna and flora

⁴⁸ The MMO responsibilities in relation to management of MCZs are laid out in Sections 125 to 133 of the Marine and Coastal Access Act 2009

⁴⁹Managing Fisheries in MPAs matrix: www.gov.uk/government/publications/fisheries-in-european-marine-sites-matrix

Sources of evidence

Evidence used in the assessments falls into two broad categories:

1. Fishing activity information. This includes patterns, intensity, and trends of fishing activities and types of gear used.
2. Ecological information, in particular the location, condition and sensitivity of designated features.

Fishing activity information

VMS data

VMS data are derived from positional information reported by UK and EU member state vessels carrying the EU mandated vessel monitoring system (VMS). Since 2015 all commercial fishing vessels of 12 metres and over in length have been required to report their position, course and speed at regular intervals using VMS. Prior to 2015 this requirement applied to commercial fishing vessels of 15 metres and over.

VMS data were analysed in ArcGIS. VMS reports not associated with fishing activity were removed. These included reports with speeds greater than 6 knots (indicating non-fishing) and reports from vessels known to be performing guard ship duties for marine developments.

For UK vessels gear type and landings were assigned to VMS data by matching each report to gear types recorded in relevant landings declarations, logbooks and the Community Fishing Fleet Register.

For EU member state vessels only gear types are assigned to the VMS data as individual vessel landings are not available.

Landings data

Landings data are recorded at International Council for the Exploration of the Sea (ICES) statistical rectangle⁵⁰ level through landings declarations and logbooks.

In areas where a high proportion of landings came from vessels with VMS, landings data from vessels with VMS were linked to VMS-derived location reports to provide spatial estimates of where landings were derived from within an ICES rectangle (see VMS data above).

For vessels that do not require VMS (<12 m in length) or EU member state vessels where landings are not assigned to VMS reports (see VMS data above), landings from within specified areas (e.g. MPA's or area of feature) are estimated using the proportion of VMS reports (for VMS vessels) or the relative size of the MPA/Feature area compared to the sea area of the containing ICES rectangle(s).

⁵⁰ ICES statistical rectangles are part of a widely used grid system for North Eastern Atlantic waters. For more information see: <https://www.ices.dk/data/maps/Pages/ICES-statistical-rectangles.aspx>

Landings data are analysed to determine quantities of landings by gear group and vessel size group.

Spatial footprint analysis

See Annex 2 for how spatial footprint analysis using Pr-values were calculated.

Vessel Sightings data

Sighting information is recorded into the Monitoring Control and Surveillance System (MCSS). It is collected by various bodies such as MMO coastal staff, IFCAs, Navy patrols and other relevant agencies and contains the following:

1. Date and time of sighting
2. Reporting body
3. Vessel name, ID, gear type
4. Approximate location of vessel
5. Approximate speed of vessel
6. Whether the vessel is: Laid/tied up, steaming or fishing.

SPUE Fisheries sightings data

Sightings data between 2010 and 2012 were collated and analysed to produce Sightings Per Unit Effort figures for a Defra commissioned Cefas project published in 2014 to better understand trends in inshore fisheries⁵¹.

These data were displayed as national layers of sightings (of certain fishing activities; trawling, potting, netting etc.) per unit effort.

MMO and IFCA expert opinion on fishing activity

MMO marine officers and IFCA inshore fisheries and conservation officers provided information on fishing activity within MPAs. Information included number and size of vessels fishing, target species, type and amount of fishing gear used and seasonal trends in activity. Confidence levels were provided alongside expert opinion and estimates were provided where exact numbers were not known.

Fishing Industry Information

Where possible and achievable, information from the fishing industry regarding current fishing locations, intensity and gear types has been used to build the evidence base for the assessment.

⁵¹<http://randd.defra.gov.uk/Default.aspx?Menu=Menu&Module=More&Location=None&Completed=1&ProjectID=18126>

FisherMap data

Source: 2012 Marine Conservation Zone Project Stakmap Commercial Fishing under 15 m vessels lines summary by month. In 2012, the FisherMap project conducted interviews with almost 1000 skippers of the under 15 m fishing fleet, with the aim of mapping the activities of the commercial fishing fleet. Of those interviewed, 594 gave their permission for their data to be shared with third parties.

The data was presented as a year's activity, collected from a series of monthly totals of vessel visits, per grid cell. Summary data is provided as a series of monthly totals of vessel visits per grid cell. FisherMap data and expert opinion is used to calculate numbers of under 15 m vessels operating in a given site.

Ecological information

The fisheries assessments use the conservation advice packages produced by Natural England and the Joint Nature Conservation Committee. These provide information on the features of the site, their area and conditions. The packages also contain advice on operations and supplementary advice documents which allow the assessment of which pressure/gear combinations a feature may be sensitive too.

For some assessments, further ecological information has also been provided by Natural England. This information is available in the relevant assessments.

Sensitivity, vulnerability and site integrity

The following definitions of sensitivity, vulnerability and site integrity are used in MMO assessments.

Sensitivity is defined as:

a measure of tolerance (or intolerance) to changes in environmental conditions.⁵²

Vulnerability is defined as:

a combination of the sensitivity of a feature to a particular pressure/activity, and its exposure to that pressure/activity.

Site integrity is defined as:

The integrity of a site is the coherence of its ecological structure and function, across its whole area, that enables it to sustain the habitat, complex of habitats and/or the levels of populations of the species for which it was designated, and reference the following guidance document.⁵³

⁵² Tilin *et al.* 2010; Roberts *et al.* 2010.

⁵³ <https://www.gov.uk/guidance/appropriate-assessment>

Annex 2 - Assumptions used to calculate spatial footprint (Pr-values)

1. Pr-value background

1.1. Introduction

MMO are required to assess the impacts of all fisheries on designated features and habitats within marine protected areas (MPAs) in English waters.

The application of a “footprint” approach has been promoted by previous authors (such as Jennings *et al.*, 2012⁵⁴) as a method to quantify fishing pressure within an area of interest (AOI) such as a ‘fishing impact equation’ where:

$$\text{Fishing footprint (Pr)} = \frac{\text{Fishing effort within AOI} \times \text{Area fished by individual vessel in 1 day}}{\text{Total area of MPA/feature}}$$

Generating a “fishing footprint value” (Pr) aims to define the level of pressure for a single average day of effort for a reference vessel or fisher (land-based) within a fleet, taking into account the gear used. This value could be multiplied by the number of vessels or fishers to give the total pressure for a particular gear over a specific time period e.g. a calendar year.

This aims to inform assessments concerning the level of impact that is acceptable for maintaining integrity of the site or feature. This approach can also be used to help define the spatial extent of the fisheries activities (in relation to feature size) or simply identify where interactions exist with features (which may in itself signify adverse effect and warrant management measures). The equation can also be used to model “worst case” scenarios to help define upper limits of potential impact, which can be refined to more realistic levels with local expert judgement.

However the factors involved in calculating the area of interaction and level of impact can be complex depending on the range of vessels, fishing effort and gear types used in the area, temporal or spatial patterns of activity within the fishery, the frequency of impacts and resilience of the habitats concerned, and any cumulative impacts of different types of gear. The incorporation of these factors will need to be considered when calculating the equation, along with the availability and robustness of data to provide such information for current and future assessments.

In order to calculate the fishing pressure effectively for each gear, a clear understanding of the three parameters that define the fishing pressure must be obtained.

1.1.1. Fishing effort

In order to calculate fishing effort there are two specific variables that must be defined for each gear type:

⁵⁴ Jennings, S., Lee, J., Hiddink, J.G., 2012. Assessing fishery footprints and the trade-offs between landings value, habitat sensitivity, and fishing impacts to inform marine spatial planning and an ecosystem approach. *ICES J. Mar. Sci.* 69, 1053–1063. doi:10.1093/icesjms/fss050

- **Effort** (the number of effort units for a particular gear type) and
- **Area of interaction** (the area of contact from a unit of gear)

A source of effort data is vessel monitoring system (VMS) data as this represents high quality independent data that can be linked to logbook data for UK vessels to verify and merge catch and effort datasets. Area of interaction is defined as the actual impact of the individual gear type based on the proportion of gear in contact with the bottom and this information can be sourced from scientific literature and/or interviews (see section 3.1 for further details).

1.1.2. Area of interest

The area of interest (AOI) could be defined as the MPA itself or designated features within a specific MPA. Data sources on the distribution and extent of designated features could be obtained from statutory nature conservation bodies (SNCBs) such as Natural England and the Joint Nature Conservation Committee (JNCC).

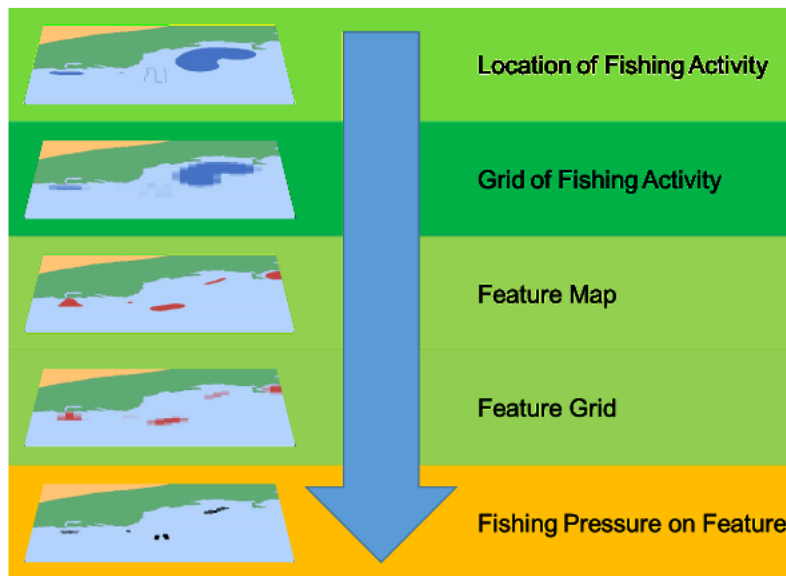
1.2. Developing the equation further

In order to determine the level of impact of fishing activity on designated features, the sensitivity of the feature should be incorporated into the proposed fisheries footprint calculation to help determine the extent to which the interaction is likely to cause an adverse effect. The sensitivity of the feature may be influenced by the time of recovery of a feature, the level of natural disturbance, cumulative impacts etc. This was identified through the fisheries European Marine Site (EMS) matrix and further scientific literature reviews.

Fishing effort also varies in terms of both the spatial and temporal distribution, potentially leading to clustering and non-uniform distribution of fishing effort across a single feature. Therefore, gaining an understanding of intensity of fishing on a feature would be useful in identifying potential cumulative impacts.

To incorporate clumping or non-uniform distribution of fishing effort a geospatial system was developed.

Figure 26: An example of input layers and stages for geospatial calculations



Spatial and temporal data was obtained in the form of VMS data to map fishing activity (effort). Area of interaction with the seabed from different gears was calculated using scientific literature and interviews with informed individuals. Feature maps of designated features within MPAs were obtained from SNCBs. From this the following can be calculated for the different gear types:

- Single VMS report gear footprint (m^2): This calculates the gear fishing footprint equivalent to a single VMS report across a cell area (0.2025 km^2) over a 2 hr time frame.
- Total VMS report area (km^2): This calculates the sum of unique cell areas (0.2025 km^2) where VMS reports occur.
- Total gear footprint (km^2): This is the total area impacted by fishing gear. This is calculated by multiplying the total number of VMS reports by cell area (0.2025 km^2) and the single VMS report gear footprint.
- Pr-value: Total extent of AOI impacted by gear (as a ratio). This is calculated by dividing total gear footprint by the AOI.
- Pr-value percentage: Percentage of AOI impacted by gear.

2. Analysis

2.1. Single VMS report gear footprint

The types of gear currently included in the gear calculators which calculates the single VMS report gear footprint are described in Table 37.

Table 37: A description of gear and the gear code used

IFISH Code	Gear	Brief Description
DRB	Boat dredges	Two types; one that is dragged along seabed, another that is like a benthic scoop that penetrates the sea bottom. Targets mussels, clams, scallops, crab etc.
FPO	Pots	Cages/baskets made from various materials and come in various sizes. Mainly set on the bottom, sometimes designed for mid-water use. Pots target fish, crustaceans and cephalopods.
GN/GNS	Gillnets (not specified) /Set gillnets (anchored)	A gillnet is a wall of netting that hangs in the water column. Set gillnets are anchored in the seabed and held down by the heavy rope line. They can be either vertical (with a float line) or flat (without a float line). Targets coastal species.
HMD	Mechanized dredges	Hydraulic dredges dig and wash out mussels from the seabed. It is considered a harvesting machine when the same gear collects the mussels and hauls them on board.
OTB	Otter trawls - bottom	Dragged along bottom and has an extended top panel to stop fish escaping upwards. Targets bottom and demersal species.
OTT	Otter twin trawls	Two identical trawls fixed together to increase the fishing area. Two otter boards to hold mouths open, one at each far end. The connection between the two trawls is a rope which joins the connection between the two pulling. Usually targets shrimp.
SDN	Danish seine	A weighted rope is anchored at one end by a marker buoy, while the other is attached to the vessel. The vessel sweeps in a circular motion to deploy the rope and the attached net. Once deployed, the gear is towed in, and the net winched onto the vessel to collect species forced into the path of the net.
SSC	Scottish seine	Similar to the above but hauled while the vessel is stationary using its own engine power rather than an anchor.
TBB	Beam trawls	Mouth of trawl is permanently held open by a beam with guides/skids attached. This disturbs bottom fish which rise up and get caught.
TBN	Nephrops trawls	Adapted to be selective for Nephrops with mall holed mesh. Some have devices to allow the inevitable larger by-catch to escape.

Each gear type has a gear calculator which calculates the gear fishing footprint for a cell area over a 2 hour time frame. A cell is 450 m by 450 m (20250 m²) or 0.2025 km², 2 hours was

chosen as it is the maximum time allowed between VMS reports. This is calculated as 0.083 or one twelfth of a day.

The calculation is as follows for trawls or dredge gears:

$$\text{Single VMS report} = \frac{\text{Total width of gear (m)} * \text{Total length hauled per day (m)}}{\text{Area of cell size (20250 m}^2\text{)}} * 2 \text{ hr period (0.083)}$$

The calculation is as follows for nets and lines, pots and traps, hand-gathering or single position gears:

$$\text{Single VMS report} = \frac{\text{Area of impact from one unit of gear (m}^2\text{)} * \text{No.of operations in one day}}{\text{Area of cell size (20250 m}^2\text{)}} *$$

2 hr period (0.083)

This gives an estimate of the area (in m²) impacted by gear from a single VMS report based on the different fishing gears (Table 38). However, this does assume the same size gear and amount of operations/hauls which occur for each gear type regardless of other variables (e.g. boat length, engine power, bylaws in place etc.). See section 3.1 for assumptions made about the gear calculations.

Table 38: Estimate of different gears fishing footprint across a cell area for a two-hour period.

Gear	Single VMS report gear fishing footprint over cell area (m²)
TBB	1.336195
OTT	0.225177
DRB	0.437237
OTB	0.115868
OT	0.115868
HMD	0.057756
TBN	0.034159
GNS	0.001808
GN	0.001808
FPO	0.00001
SDN	0.003689
SSC	0.005849

2.2. Pr-value model

The Pr-value model requires several datasets as inputs including:

- Annual UK VMS data for >12 m vessels
- Annual non-UK VMS data >12 m vessels
- Single VMS report gear footprint calculations
- MPA sites and designated feature data

Assumptions about the datasets are included in Section 3.

The Pr-value model has the following steps:

1. The UK and non-UK VMS data is clipped to the area of interest (MPA site or designated feature within site).
2. VMS reports which are denoted as 'fishing' are chosen (vessels travelling between >0 and <6 knots).
3. VMS reports from the same vessels which are less than 2 hours apart (7080 seconds exactly, see Section 3.4 for explanation) are excluded.
4. The processed VMS data (VMS reports = fishing and ≥ 2 hours) is joined to the gear calculations data.
5. A grid is created across the area of interest, with cell sizes of 450 m by 450 m.
6. The grid and processed VMS data are joined together.
7. Gear not included in the current gear calculators is excluded.
8. The cell area is calculated as 0.2025 km² for each cell.
9. Total gear footprint is calculated by multiplying single VMS report gear footprint by the cell area (0.2025 km²). This is then multiplied by the number of VMS reports per gear type.
10. The VMS report area and total gear footprint is summed by gear type.
11. A summary table is created which includes:
 - AOI field (km²)
 - AOI name (text)
 - Total VMS report area (km²): Sum of unique cell areas (0.2025 km²) where VMS reports occur.
 - Total gear footprint (km²): Total area impacted by fishing gear.
Total no. of fishing VMS reports * cell area (0.2025) *
single VMS report gear footprint
 - Pr-value: Total extent of AOI impacted by gear. $\frac{\text{Total gear footprint}}{\text{AOI}}$
 - Pr-value percentage (%): Percentage of AOI impacted by gear. $\frac{\text{Total gear footprint}}{\text{AOI}} * 100$

3. Pr-value Assumptions

3.1 Gear Calculators

A cell is 450 m by 450 m or 0.2025 km². Two hours was chosen as it is the maximum time allowed between VMS reports. These were chosen so that a beam trawler (the largest swept area) will have covered the whole cell in 2 hours.

When calculating P-values the gear type assigned by the MMO statistical team is used. If no gear has been assigned, then if the vessel has an assigned statistical gear in the same year

that gear is used. If no match can be found, then the primary gear type assigned in the Food and Agriculture Organization of the United Nations (FAO) fishing vessel database is used.

DRB:

- Based on a 16.28 m scallop vessel with 2 x 6.7 m dredge bars each with two shoes at 720 mm wide. Each dredge bar has 8 x 76 cm dredges (Lart, 2012).
- Number of vessels and days spent fishing: derived from VMS/landings records.

FPO:

- Size of pot: based on MMO coastal officer advice – 120 cm x 70 cm.
- Number of vessels and days spent fishing: derived from VMS/landings records.
- Number of pots used by vessels: derived from local fisherman – 30 pots per day.

GN/GNS:

- Based on a vessel shooting 10 tiers each 132 m. Each tier has 2 anchors at 1.0 m x 0.5 m (MMO coastal officer has corrected this from 0.5 cm x 0.5 cm). Foot rope 3 m wide drag. Info derived from Seafish report on a workshop on the physical effects of fishing activities on Dogger Bank and Annexes to: Feasibility study on applying a spatial footprint approach to quantifying fishing pressure.
- 5.5 nets hauled per day. Info derived from Seafish report on a workshop on the physical effects of fishing activities on Dogger Bank and MMO coastal.

HMD:

- Based on 1 cage with a total width of approximately 1.9 m. Data from <https://spo.nmfs.noaa.gov/sites/default/files/pdf-content/MFR/mfr444/mfr4441.pdf>
- Haul duration 10 – 12 hrs. Data from <http://www.seafish.org/media/Publications/SR348.pdf>
- Haul speed 4 knots. Data from <http://www.seafish.org/media/Publications/SR348.pdf>

OT/OTB:

- Based on a vessel with 4 m net width, two 0.65 m otter boards, 60% ground rope interaction. Info derived from Seafish report on a workshop on the physical effects of fishing activities on Dogger Bank and Annexes to: Feasibility study on applying a spatial footprint approach to quantifying fishing pressure. 4 m net width and 0.65 m board width comes from MMO coastal team.
- Number of vessels and days spent fishing: derived from VMS/landings records.

OTT:

- Based on a vessel with 2 * 4 m trawl, two 0.65 m otter boards, 60% ground rope interaction and 1 clump of 0.6 m. Info derived from Seafish report on a workshop on the physical effects of fishing activities on Dogger Bank and Annexes to: Feasibility study on applying a spatial footprint approach to quantifying fishing pressure. 4 m net width and 0.65 m board width comes from MMO coastal team (Inner Dowsing). Haul duration 4 hours, from MMO officer.
- Haul speed 4 knots, from MMO officer.

TBB:

- Based on a vessel with 2 x 12 m trawl, four 720 mm shoes and 2 tickler chains with 60% interaction with the seabed. Info derived from Seafish report on a workshop on the physical effects of fishing activities on Dogger Bank and Annexes to: Feasibility study on applying a spatial footprint approach to quantifying fishing pressure.
- Haul duration 4 hours. Info derived from Seafish report on a workshop on the physical effects of fishing activities on Dogger Bank and MMO coastal.
- Haul speed 4 knots. Info derived from Seafish report on a workshop on the physical effects of fishing activities on Dogger Bank and MMO coastal.

TBN:

- Based on a vessel with 2 x 3.5 m beam trawls, 4 x 0.2 m feet and 60% ground rope interaction. Info derived from Annexes to: Feasibility study on applying a spatial footprint approach to quantifying fishing pressure.
- Haul duration 2 hours. Info derived from Annexes to: Feasibility study on applying a spatial footprint approach to quantifying fishing pressure.
- Haul speed 1.5 knots. Info derived from Annexes to: Feasibility study on applying a spatial footprint approach to quantifying fishing pressure.

SDN:

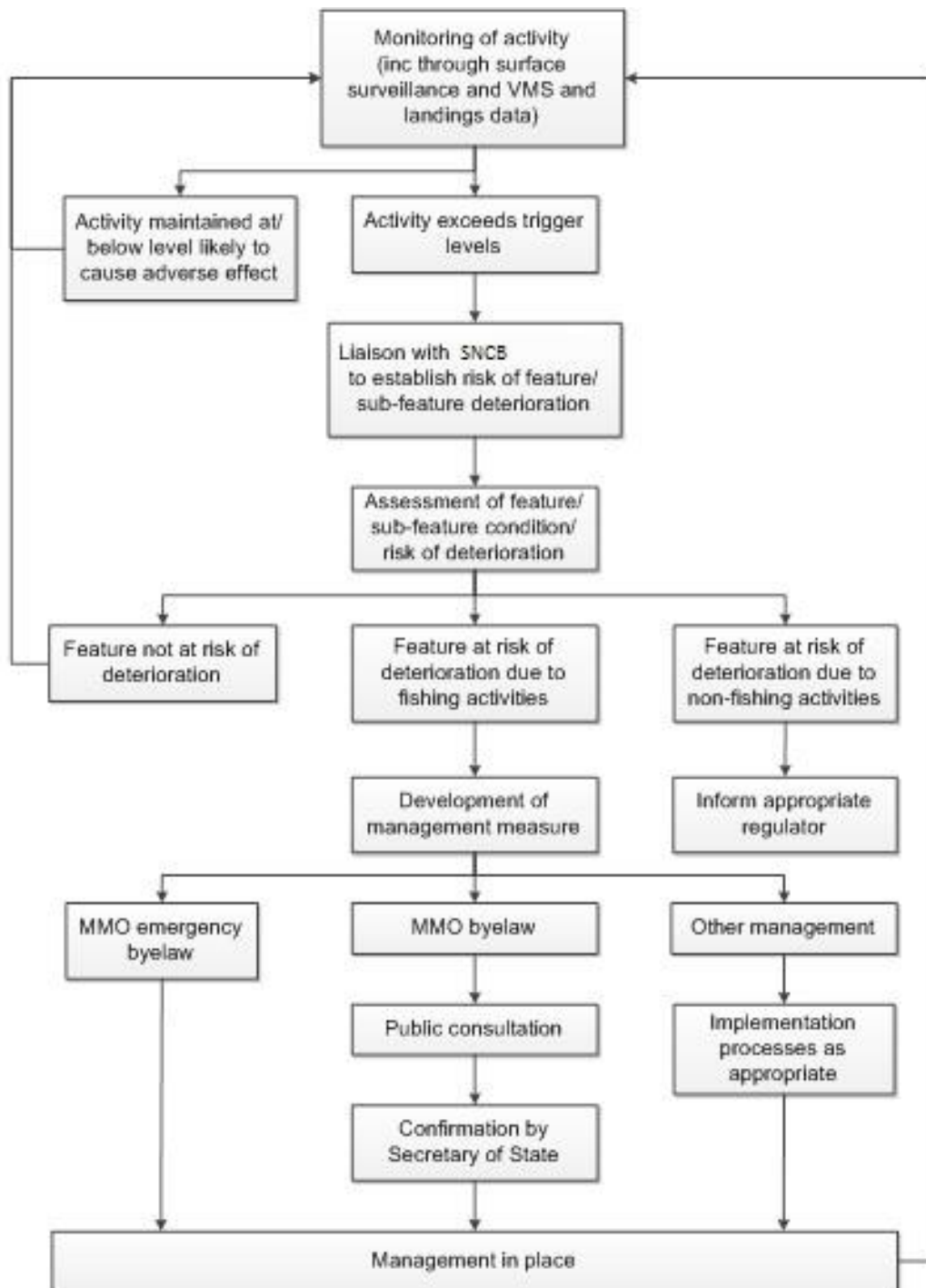
- As this gear is not trawled and the total area impacted by the gear is available, this has been considered a net / line gear.
- Specific details were not available however the marine institute estimates the area of impact as being on average 2.25 km². <https://www.seafish.org/document/?id=6425bbfc-06f9-4205-9e50-f6839e6206ea>
- SDN only takes place in daylight hours. Hauls per day (4) is estimated on the average number of daylight hours (12) and the haul duration ~ 3 hours.

SSC:

- As this gear is not trawled and the total area impacted by the gear is available (below), this has been considered a net / line gear.
- Specific details were not available however according to Seafish a seiner with 13 coils of rope lead to 2854 m² area of impact. <https://www.seafish.org/responsible-sourcing/fishing-gear-database/gear/ssc-scottish-seine/>
- SSC only takes place in daylight hours. Hauls per day (5) is estimated on the average number of daylight hours (12) and the haul duration ~ 2.25 hours (median of 1.5 – 3 hours as described by Seafish: <https://www.seafish.org/responsible-sourcing/fishing-gear-database/gear/ssc-scottish-seine/>

Annex 3 - Monitoring and Control Process

Figure 27: Monitoring and control process



Annex 4 - MMO MPA management decision tree

MMO has legal obligations in relation to Marine Conservation Zones (MCZs) and European Marine Sites (EMS) which include Special Areas of Conservation (SAC) and Special Protection Areas (SPA):

- Under the Conservation of Habitats and Species Regulations, 2017 reg 9 and Conservation of Offshore Marine Habitats and Species Regulations, 2017 reg 6 to secure compliance with the requirements of the Habitats Directive. Of particular relevance to marine conservation is section 6(2): to avoid the deterioration of habitats and disturbance of designated species; and
- Under the Marine and Coastal Access Act, 2009 125(2)(a) to consider using its functions that best furthers the conservation objectives stated for the MCZ.

Site-level assessments are carried out in a manner consistent with the requirements of section 126 of the Marine and Coastal Access Act 2009 for MCZ and Article 6(3) of the Habitats Directive for EMS, which relates to assessments completed for plans and projects.

The assessments determine whether:

- For EMS, fishing activities are having an adverse effect on the integrity of the site.
- For MCZ, there is a significant risk of fishing activities hindering the conservation objectives of the site.

If this is the case MMO has a legal duty to appropriately manage these activities to ensure compliance with the Habitats Regulations and MaCAA. Proposed management measures generally fall within one of two approaches:

1. Reduction of pressures associated with fishing gear(s) of concern:
 - zoned management (partial site/feature prohibition of these gears)
2. Removal of pressures associated with fishing gear(s) of concern:
 - a whole site/feature prohibition of these gears

To determine the most appropriate management approach the MMO adheres to the decision tree detailed in Figure 28.

MMO make use of the best available habitat/biotope data and JNCC's Marine Evidence-based Sensitivity Assessment (MarESA), along with SNCB conservation advice and scientific evidence provided in the fisheries assessment to identify areas within the site where the feature(s) may be suitably resilient, or of lower sensitivity (resistant) to the pressures associated with the fishing gear(s) of concern. If the feature(s) present are identified from the scientific evidence-base as being potentially more resilient and/or resistant to the gear(s) of concern, consideration is given to areas within the site where these feature(s) are present that could potentially remain open, either with or without restrictions on fishing activity levels.

As an independent coastal State, MMO in its assessment of the evidence must fulfil its regulatory duties and responsibility to ensure compliance with the Conservation of Offshore Marine Habitats and Species Regulations 2017; to exercise all relevant functions to ensure compliance with the Habitats Regulations. Having regard to the current conservation status of

the site/feature(s) and the site's conservation objectives, MMO must be satisfied beyond reasonable scientific doubt, that fishing gear(s) of concern can continue within the site/feature(s) at reduced capacity (zoned management approach) and/or restricted activity levels without compromising compliance with its regulatory duties and responsibilities. MMO will act in accordance with the precautionary principle, and this includes when the scientific evidence-base indicates that an adverse effect on the integrity of the site (EMS) and/or significant risk of hindering the conservation objectives of the site (MCZs) cannot be avoided.

If MMO is not satisfied, beyond reasonable scientific doubt, that a reduce/limit management approach will allow compliance with its duties under the Habitats Regulations/MaCAA, a removal of pressure will be deemed necessary for the gear/feature interaction(s) and a full site/feature(s) prohibition of the fishing gear(s) of concern will be implemented (whole site prohibition).

Figure 28: MMO MPA management decision tree

