



Performance Standard for Laboratories Undertaking Chemical Testing of Soil

Environment Agency
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Foreword

The Environment Agency has established its Monitoring Certification Scheme: MCERTS to deliver high quality environmental measurements. The scheme provides for the product certification of instruments, the competency certification of personnel and the accreditation of laboratories based on international standards.

For the chemical testing of soil where results are to be submitted to us for regulatory purposes, we require laboratories to be accredited to the current version of the European and international standard, ISO/IEC 17025 for this MCERTS performance standard. Accreditation is undertaken by an appropriate national organisation, which in the United Kingdom is the United Kingdom Accreditation Service (UKAS). The MCERTS performance standard provides an application of ISO/IEC 17025 specifically for the chemical testing of soil and covers:

- performance targets
- the selection and validation of methods
- sampling pre-treatment and preparation
- participation in proficiency testing schemes
- the reporting of results and information.

The benefits of MCERTS for the chemical testing of soil are that the scheme:

- provides formal accreditation of laboratories in accordance with European and international standards
- provides assurance to all stakeholders (including industrial process operators, laboratories, regulators and the public) of the reliability of data from tests
- establishes a level playing field in this competitive market, based on the Environment Agency's requirements
- indicates that the chemical testing of soil is a critical component in producing defensible data for regulatory purposes
- promotes and raises the professional standing of laboratories by establishing "quality standards" to which all should aspire and be judged.

Some of the requirements of the performance standard are described in general terms. This is to allow a degree of flexibility for a laboratory and to allow the laboratory to take advantage of technological developments. In this way, a laboratory is not excluded simply because, for example, it lacks specific equipment. However, along with this flexibility is the need for the provision of appropriate information. For example, if test data are to be generated for a specific site over an extended period it is essential that consistent and meaningful comparisons can be made. Where we assess data for regulatory purposes, all relevant information must be recorded and be available to us, if requested.

Most of the requirements of this performance standard are aimed at laboratory activities. However, if data are to be submitted to us for regulatory purposes, the procurer of the analytical services must ensure that the requirements are satisfied and that the appropriate information is provided to us, if requested.

It is recognised that variations due to sampling can be greater than those introduced via analysis; however, this performance standard does not specifically cover sampling or the competency of personnel in relation to sampling procedures and strategies.

If you have any questions regarding the accreditation process, or would like further information on how to apply, please contact:

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For more information on MCERTS and for copies of the performance standards and further guidance, see our web-site at:

www.mcerts.net

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MCERTS Performance Standard for Laboratories Undertaking Chemical Testing of Soil

Introduction

The extension of MCERTS to include the chemical testing of soil is built on proven international standards to ensure that the quality of test data is high. This performance standard details the requirements for a laboratory undertaking the chemical testing of soil and the procurer of analytical services to the MCERTS performance standard.

The general requirements for the competence of testing and calibration laboratories are described in the European and international standard ISO/IEC 17025. Where data are submitted to the Environment Agency for regulatory purposes, those data shall be generated using methods that have been accredited to the European and international standard ISO/IEC 17025 for this MCERTS performance standard. Such methods shall be included within an accredited laboratory's scope of activities. This performance standard contains requirements that a laboratory must meet if it wishes to demonstrate that it operates a management system, is technically competent and able to generate valid results, and wishes to be considered as a laboratory registered under the MCERTS performance standard for the chemical testing of soil. In addition, there are also requirements for a procurer of analytical services that wishes to submit data to the Environment Agency for regulatory purposes.

ISO/IEC 17025 recognises - in clause 1.6 (Note 1) - that it might be necessary to explain or interpret certain requirements in the international standard to ensure that the requirements are applied in a consistent manner. The MCERTS performance standard provides criteria for applying ISO/IEC 17025 in the specific field of the chemical testing of soil. In producing this MCERTS performance standard, guidance given in Annex B of ISO/IEC 17025 has been followed.

The MCERTS performance standard for laboratories undertaking the chemical testing of soil does not restate all the provisions of ISO/IEC 17025 which must be complied with. It states only those additional requirements which must also be complied with, in order for a laboratory to become registered under MCERTS for the chemical testing of soil.

The clause numbers in this document align with those of ISO/IEC 17025:2005, and may not be the same as those in other dated versions of ISO/IEC 17025. The text of ISO/IEC 17025 is not repeated, and where no additional requirements are needed, this is stated.

1 Scope

- 1.1** The chemical testing of soil can be undertaken for a wide range of parameters using a wide range of methods. The methods that a laboratory uses to generate data that are submitted to the Environment Agency for regulatory purposes shall be accredited to ISO/IEC 17025 for this MCERTS performance standard. These methods shall be defined in the laboratory's scope of activities.
- 1.2** This performance standard is applicable to all laboratories and procurers of analytical services where results, generated for the chemical testing of soil, are submitted to the Environment Agency for regulatory purposes.
- 1.3** No additional requirements to ISO/IEC 17025.
- 1.4** When a laboratory satisfies all of the appropriate requirements of this performance standard, that laboratory will have demonstrated that it meets the Environment Agency's MCERTS requirements for the chemical testing of soil.
- 1.5** No additional requirements to ISO/IEC 17025.
- 1.6** If a laboratory complies with the appropriate requirements of this performance standard, it will be regarded by the Environment Agency as demonstrating its competence and of being capable of undertaking the chemical testing of soil to the Environment Agency's requirements, for its published scope of activities. Its details shall be defined in a scope of accreditation published on the UKAS website.

2 References

2.1 Normative references

ISO/IEC 17025 - General requirements for the competence of testing and calibration laboratories.

2.2 Text references

- a) ISO TR 13530:1997 "Water Quality - A Guide to Analytical Quality Control for Water Analysis".
- b) "A Manual on Analytical Quality Control for the Water Industry", R. V. Cheeseman and A. L. Wilson, revised by M. J. Gardner, NS 30, Water Research Centre, 1989. ISBN 0-902156-85-3
- c) "Valid Analytical Methods and Procedures", C. Burgess, Royal Society of Chemistry, 2000. ISBN 0-85404-482-5.
- d) "Guidelines for achieving quality in trace analysis", M. Sargent and G. MacKay, Royal Society of Chemistry, 1995, ISBN 0-85404-402-7.
- e) "Technical aspects of site investigation", Environment Agency R&D Technical Report P5-065/TR, 2000.
- f) "Secondary model procedures for the development of appropriate soil sampling strategies for land contamination", Environment Agency R&D Technical Report P5-066/TR, 2000.

- g) "Sampling as a source of measurement uncertainty: techniques for quantification and comparison with analytical sources". *Journal of Analytical Atomic Spectrometry*, M. H. Ramsey, 1998, **13**(2), 97-104.
- h) "Quantifying Uncertainty in Analytical Measurement". *Eurachem/CITAC Guide CG4*, second edition, 2000. ISBN 0-948926-15-5 (www.eurachem.ul.pt).
- i) "An aid to accreditation" *Eurachem/CITAC Guide to quality in analytical chemistry*, Edition 2002, (www.eurachem.org).
- j) "The fitness for purpose of analytical methods". A laboratory guide to method validation and related topics. *Eurachem Guide* 1998, (www.eurachem.org).
- k) "Development and Harmonisation of Measurement Uncertainty Principles – Part (d): Protocol for uncertainty evaluation from validation data." V J Barwick, S L R Ellison, LGC/VAM/1998/088.
- l) "Handbook for Calculation of Measurement Uncertainty in Environmental Laboratories". Version 1.3, Nordtest Report TR 537 October 2003.
- m) "The J-chart: a simple plot that combines the capabilities of Shewhart and cusum charts, for use in analytical quality control". Analytical Methods Committee technical brief No.12, Royal Society of Chemistry 2003.
- n) "Quality Control Charts in Routine Analysis", M J Gardner, WRc Report CO4239 1996.
- o) "Guidelines for the In-House Production of Reference Materials" – version 2, B Brookman, R Walker 1998 LGC/VAM/1998/040.
- p) "Applications of Reference Materials in Analytical Chemistry" - V. Barwick, S. Burke, R. Lawn, P. Roper and R. Walker Royal Society of Chemistry, Cambridge, 2001 ISBN 0-85404-448-5.
- q) "Soil Science: Methods and Applications" - D.L.Rowell. Longmans, 1994 ISBN 0-582-08784-8.

3 Terms and definitions

In the context of this performance standard, the following terms and definitions apply: It is recognised that some terms used in this document may have slightly different meanings to those used in other publications.

Air-dried sample - Sample that has been dried at ambient temperatures not exceeding 30°C.

Analytical Quality Control (AQC) - The overall process of ensuring that the application of an analytical method is controlled within specified tolerances.

Assisted-dried sample - A sample that has undergone a specified accelerated drying process. This may involve oven-assisted drying at a specified temperature, freeze-drying or some other process.

As submitted basis - The sample as it is received at the laboratory for direct analysis of either the entire sample, or of a representative sub-sample, i.e. without further sample pre-treatment.

Batch - A number of samples prepared for a discrete analytical run.

Bias - Bias, which may be positive or negative is the difference (expressed as a percentage) between the mean of a number of determinations obtained under repeatability conditions and the true or accepted concentration.

$$\% \text{Bias} = \frac{(\text{mean of determinations} - \text{true or accepted value}) \times 100}{\text{True or accepted value}}$$

Bias can be estimated where appropriate certified reference materials are available and a stated (certified) concentration has been quoted. Recovery data can be used to estimate bias via spiking experiments (see spiking recovery).

Certified Reference Material (CRM) - Reference material, accompanied by a certificate, one or more of whose property values are certified by a procedure which establishes its traceability to an accurate realization of the unit in which the property values are expressed, and for which each certified value is accompanied by an uncertainty at a stated level of confidence. [ISO/IEC-Guide 30]

Concentration - Concentration is usually expressed as mass per unit mass, for example mg kg⁻¹. It may be quoted on an “as submitted” basis, a “wet weight” basis, or on a “dry weight” basis. (In certain circumstances the term concentration is not appropriate, for example in the determination of pH values).

Critical level of interest - This is the concentration value around which a decision is often required, for example is the concentration above or below a certain value. It may be for example a “soil guideline value” (which may be found for certain parameters on the Environment Agency website), a regulatory limit, or some other concentration of importance. A method is usually deemed acceptable if, when used properly, it is capable of establishing within defined limits of bias and precision, whether a concentration is above or below the critical level of interest.

Laboratory - A laboratory, or sub-contracting laboratory, that undertakes the chemical testing of soil.

Parameter - Within the sample, this is the determinand, measurand, analyte, substance, or group of substances, the concentration of which needs to be determined. It shall be clearly and unambiguously defined.

Performance characteristics - Those performance values, such as precision, bias (or recovery, as appropriate) and limit of detection that need to be estimated before a method is used routinely.

Precision - This is the distribution of a number of repeated determinations, expressed in this document as the % relative standard deviation (RSD).

$$\% \text{RSD} = \frac{S \times 100}{M}$$

Where S = total standard deviation, M is the mean of results, both as indicated in section 5.4.5.3

Repeatability conditions - Those conditions where analyses are carried out in one laboratory by one or more analysts, using the same equipment and reagents, within a short period of time.

Sample - That (uniquely identified) material removed from a site and submitted to the laboratory for analysis.

Spiking recovery - The addition of a known quantity of a parameter to a sub-sample, followed by analysis to establish that fraction or percentage recovered by the use of a defined method.

$$\% \text{Recovery} = \frac{(D - P) \times 100}{A}$$

Where:

D = the measured amount of parameter in sub-sample following addition of known amount of parameter;

P = the measured amount of parameter in sub-sample prior to addition of known amount of parameter;

A = amount of parameter added to sub-sample.

This technique is often used as the only viable option for the analyst when appropriate certified reference materials are not available and bias cannot be determined directly. When this is so, bias is calculated from

$$\% \text{Bias} = \% \text{Recovery} - 100$$

Statistical control - When the result or results of quality control samples are shown to be within defined limits of recognised acceptability, a method is said to be in statistical control. When these limits are breached, the method is considered out of control.

Sub-sample - A representative or homogenised portion of the sample. This portion is used in the analysis.

Wet-weight basis - The sample as it is received at the laboratory for direct analysis either of the whole sample or of a representative sub-sample, i.e. without further sample pre-treatment.

4 Management requirements

4.1 Organisation

4.1.1 No additional requirements to ISO/IEC 17025.

4.1.2 For data to be submitted to the Environment Agency for regulatory purposes, the laboratory shall carry out its testing and calibration activities in such a way as to meet the requirements of this performance standard.

4.1.3 - 4.3.3.4 No additional requirements to ISO/IEC 17025.

4.4 Review of requests, tenders and contracts

4.4.1 No additional requirements to ISO/IEC 17025.

4.4.1(a) For data to be submitted to the Environment Agency for regulatory purposes, the requirements of the methods to be used shall be clearly and unambiguously defined and documented. The laboratory shall demonstrate that the requirements of the methods to be used shall be understood by those who undertake the analysis.

Note: The laboratory may or may not be aware that the data it generates will be submitted to the Environment Agency for regulatory purposes. However, the laboratory's customer or procurer of the analytical service should be aware that if it wishes to submit the data to the Environment Agency for regulatory purposes, then the requirements of this performance standard need to be satisfied.

4.4.1(b) No additional requirements to ISO/IEC 17025.

4.4.1(c) For data to be submitted to the Environment Agency for regulatory purposes, the appropriate test and/or calibration method shall be selected and shall satisfy the requirements of this performance standard.

4.4.2 - 4.4.5 No additional requirements to ISO/IEC 17025.

4.5 Sub-contracting of tests and calibrations

4.5.1 A laboratory may sub-contract the chemical testing of soil to another laboratory. It is the responsibility of the laboratory to ensure that the sub-contracted laboratory is registered under MCERTS for the scope of work sub-contracted. The provisions of this clause do not apply to samples submitted to a laboratory by an external quality control or inter-laboratory proficiency-testing scheme organiser.

4.5.2 - 4.13.1.4 No additional requirements to ISO/IEC 17025.

4.13.2 Technical records

4.13.2.1 The laboratory shall retain records for a defined period of time of not less than six years. This period of time shall take into account the need of the customer (procurer of the analytical services) and the need to submit these records to the Environment Agency, if requested.

4.13.2.2 - 4.15.2 No additional requirements to ISO/IEC 17025.

5 Technical requirements

5.1 - 5.2.5 No additional requirements to ISO/IEC 17025.

5.3 Accommodation and environmental conditions

5.3.1 Equipment, reagents and samples shall be protected from damage or degradation, during collection, transportation and subsequent storage, as appropriate.

Note: There may be methods specifying the procedures necessary for protecting the integrity of samples and reagents during transportation and storage such as collection into suitable containers and storage out of direct sunlight at specified temperatures etc.

The laboratory shall have procedures in place and use appropriate practices to ensure that conditions do not adversely affect the measurement result.

5.3.2 - 5.3.5 No additional requirements to ISO/IEC 17025.

5.4 Test and calibration methods and method validation

5.4.1 General

The laboratory shall demonstrate and provide justification that suitable methodology (including sample pre-treatment and preparation) has been used in the analysis of a particular matrix and parameter and that it is appropriate with respect to the concentration of the parameter in the sample. The laboratory shall demonstrate and provide justification that method validation procedures have been undertaken in such a manner as is appropriate to the sample matrix undergoing analysis. Full details of the method and method validation procedures shall be made available to the Environment Agency, if requested.

5.4.2 Selection of methods

The Environment Agency will not prescribe those analytical methods that a laboratory should use, but the method used shall be appropriate for the matrix and parameter at the level of concentration being analysed. Where results are submitted to the Environment Agency for regulatory purposes, a clear and unambiguous description of the method used to generate the results shall be provided to the Environment Agency, if requested. This description, which need not be fully comprehensive, shall comprise more than the title of the method and shall clearly indicate the parameter, scope, principle and matrix for which the method is applicable.

The description of the method, parameter and matrix shall be sufficiently detailed to allow direct comparisons with similar methods, parameters and matrices that might be used and determined by other analysts or laboratories. For example, when an extraction technique is used to isolate or concentrate a particular parameter, the name of the solvent or full details of the composition of the solvent mixture shall be given. Also, the amount of soil taken for analysis and the amount of solvent used in the extraction shall be reported. In addition, where the analytical determination of an extract is undertaken and, for example, this involves the use of a specific wavelength or mass number, then details shall also be given.

The term “contaminated land” will not be sufficient to describe the matrix, which shall include reference to the major constituents and components.

Note: The description of the sample matrix may, for example, be in the form “an organic-rich (predominantly loam) soil visibly contaminated with hydrocarbons” or “an industrial soil of mostly clay and sand (containing brick debris) from a former manufacturing site” possibly with a statement of the manufactured product.

A fully documented method shall be made available to the Environment Agency, if requested.

5.4.3 - 5.4.4 No additional requirements to ISO/IEC 17025.

5.4.5 Validation of methods

5.4.5.1 No additional requirements to ISO/IEC 17025.

5.4.5.2 Before any method for a particular matrix and parameter is used for generating data for submission to the Environment Agency that method shall be accredited to ISO/IEC 17025 for this performance standard. Only those results generated using these methods will be eligible for submission to the Environment Agency for

regulatory purposes. The process of full validation provides confidence that the established performance characteristics are based on robust experimental determinations and are statistically sound.

General approach

Validation procedures include a number of operations. These shall include the analyses of, where available and appropriate, matrix certified reference materials relevant to the matrices, parameters and range of parameter concentrations under investigation. The method shall be validated for each parameter analysed on matrices likely to be analysed within the laboratory. This validation shall include at least three different soil matrices. Sample pre-treatment and preparation is an important part in the validation process and shall be considered, as this may not be monitored by the use of certified reference materials. In these cases a separate exercise to determine the effects of sample pre-treatment and preparation shall be undertaken.

Note: Whilst it is not expected that every sample submitted should require its own validated method, it is recognised that a single validated method established for one particular matrix but used for every sample, irrespective of its matrix, is not appropriate.

In the absence of suitable certified reference materials, recovery estimates relevant to the matrix and parameter under investigation shall be determined by the use of spiking experiments. Where possible these experiments shall cover the entire method (including pre-treatment, extraction and determination). The addition of a parameter to a sub-sample followed by immediate extraction is not a satisfactory test for estimating spiking recovery, as sufficient time must elapse to allow possible matrix-parameter interactions to occur. The laboratory shall demonstrate that its use of spiking experiments and the spiking procedures employed is appropriate.

Where a suitable certified reference material is, initially, not available, but then, after recovery estimates have been undertaken, becomes available, then the newly available certified reference material shall be used to check the bias is satisfactory. The number of replicate determinations required to detect a bias at a target level (for example MCERTS performance targets) can be estimated if the precision of the measurement is known.

For spiking experiments, the concentrations of the solutions used in the validation procedures shall be appropriate to the concentrations found in samples being routinely analysed. Recovery estimates shall be obtained using two significantly different but appropriate concentration levels, for example, at 20 % and 80 % of the expected range. All solutions shall either be taken from bulk stock solutions that are known (and have been shown) to be stable over the entire period of testing or, if solutions are not stable over the entire period of testing, be prepared immediately before the analysis of each validation batch. The traceability of these solutions shall have been established.

When Isotope Dilution Mass Spectrometry [IDMS] is employed, with appropriate labelled analogues of the determinands spiked into all samples - including calibration standards and matrix AQC standards,- then equilibrated appropriately before sample preparation is undertaken, the results obtained will be recovery corrected. The recovery corrected values of spiked samples and/or CRMs obtained in this manner shall be used to estimate bias against the certified CRM and/or added spike.

5.4.5.3 Validation procedure

For the method, parameter and matrix, the performance characteristics shall be determined with a minimum of ten degrees of freedom. This shall be carried out by analysing each certified reference material or spiked samples in duplicate in different analytical batches. 11 batches of duplicates will guarantee a minimum of ten degrees of freedom. However, it may be that 10 degrees of freedom will be achieved in less than 11 batches, this can be checked after each batch of results (see references a) and b) for appropriate procedures). Validation shall be undertaken in a period of time not less than six days and no more than three months.

Note 1: This procedure is often termed an 11 x 2 test, as 11 batches containing 2 replicates of each test material are analysed.

Precision should then be estimated using analysis of variance (ANOVA), from which different sources of error (for example within batch and between batch random errors) can be estimated and combined to give a total error as a standard deviation. Details of the statistical procedures for ANOVA and recovery (bias) estimation are given in references a) and b), see also Annex B of this performance standard.

Note 2: The use of a validated method for one particular soil matrix may not be suitable for the analysis of a different soil matrix. This may also be the case when analysing samples of the same soil matrix containing significantly different concentrations of the same parameter.

The laboratory shall demonstrate that the certified reference material for the matrix, methodology, parameter and concentration of parameter being analysed is appropriate.

When a method has been validated, its stated performance shall reflect the routine capability of the method. That is, when the method is used routinely, its day to day performance shall be typical of and maintained at the level of the stated validation performance.

The limit of detection of a method used to analyse highly contaminated samples may be higher than the limit of detection of a method used to analyse slightly contaminated samples. The reported limit of detection shall be fit for the intended purpose and appropriate to the concentration level of interest required of the analysis. The limit of detection shall be calculated as described in Annex B1. The limit of detection should never be used in isolation of other method validation data to judge the appropriateness of a method.

Note 3: The maximum value of the limit of detection usually regarded as being fit for purpose is 10 % of the concentration regarded as the critical level of interest.

Revalidation

After an analytical method has been validated and accredited, it is inevitable that in time some modification of procedures will take place. Any modifications to a method routinely used within a laboratory may affect the resulting performance. Any changes made to a method already accredited against the MCERTS requirements shall be notified to the national accreditation organisation. These changes could range from replacing a piece of equipment to a fundamental procedural modification, such as a different extraction procedure.

Minor changes to the analytical system may not require revalidation, but care should be taken to ensure the cumulative effects of several changes do not affect system

performance by, for example, closely monitoring internal and external AQC, and reanalysing CRMs used for validation.

If an instrument is being replaced by one of the same model, and performance is not expected to fundamentally change, laboratories need only demonstrate that the new instrument performs as well as the old instrument. This could be achieved, for example, by analysing several replicates of a representative matrix such as a spiked soil, a CRM or a soil AQC sample.

If a fundamental change is made to the analytical procedure or the equipment used then a full validation on a minimum of 3 matrices is required in accordance with this performance standard. These changes may include, for example, replacing ICPOES with ICPMS, using a new extraction technique etc.

It is recognised that an intermediate degree of validation should be carried out if significant changes are made to a method that are not considered fundamental to performance. A partial validation shall be performed (for example analysis of 6 batches of duplicates), using only one spiked sample from the lower end of the calibration range, or preferably a CRM, for all appropriate matrices. If a laboratory judges that this level of validation is required, then it shall notify and gain the approval of UKAS. Laboratories shall ensure that the amendments to the analytical system and any procedures that may be affected are included in the revalidation.

Performance criteria

The Environment Agency has specified that the following performance characteristics are acceptable for the validation of methods for the chemical testing of soil, bearing in mind the need to take meaningful decisions, current analytical capabilities and other likely sources of variation.

- The bias (or systematic error) of individual results determined for the entire method shall not be significantly greater than the figure indicated in Annex A (Tables 1 – 4) expressed as a percentage. The certified reference value of the certified reference material shall be used as the true or accepted value when calculating bias. If a critical level of interest is known, the target bias value used can be taken as one-twentieth of the critical level of interest and either bias value used whichever is the greater. Laboratories shall demonstrate that the bias satisfies the stated requirement at the critical level of interest.
- The precision, as expressed as the % relative standard deviation, of individual results determined for the entire method shall not be significantly greater than the figure indicated in Annex A (Tables 1 – 4). Precision shall be estimated using analysis of variance to determine total standard deviation. If a critical level of interest is known, the target precision value used can be taken as one-fortieth of the critical level of interest and either precision value used whichever is the greater. Laboratories shall demonstrate that the precision satisfies the stated requirement at the critical level of interest.

Testing for significance shall be carried out as described in Annex B2. If, for a particular parameter, testing shows a significant difference exists, then further method development or refinement is required, or a different analytical method used.

Annex A (Tables 1 – 4) specifies the performance characteristics for a selection of parameters (which are not to be regarded as exhaustive).

Note 4: In the context of soil analysis, bias and recovery can be regarded as synonymous. The use of certified reference materials is preferred to spiked soils, but if appropriate certified reference materials are not available (either for the matrix, parameter or parameter concentration under investigation) then spiking experiments may offer the only suitable means of estimating recovery. Where the analysis involves preparation and steps (for example drying and grinding) that are not required for the certified reference material, then a combination of CRM analyses and spiked soil analyses may be required.

When a laboratory requests accreditation of additional parameters not listed in Annex A of this standard, the following performance requirements shall be enforced:

Metals	– 7.5% precision and 10% bias
Organometallics	– 15% precision and 30% bias
Inorganics	– 10% precision and 20% bias
Organics	– 15% precision and 30% bias

If a laboratory is unable to meet these requirements due to matrix effects or fitness for purpose issues it shall propose alternative performance characteristics and submit them to the Environment Agency via UKAS for assessment.

Unvalidated matrices

If a laboratory subsequently receives soils of a different matrix to those on which full validation has been carried out, then for each sample type/matrix the laboratory shall undertake replicate analyses and carry out spike recovery tests with batches of samples as they are received. Spikes and replicates should be included at random, each at a minimum frequency 1 per 20 samples. If the batch size is less than twenty, one spike and one replicate sample shall still be required.

The results from the spiked and unspiked samples should be used to calculate the percentage recovery. Results of sample spikes and replicates should be reported.

For some parameters a high background concentration may make it difficult to assess spike recovery. An alternative approach may therefore be required, such as pre-dilution of spiked samples before analysis.

An alternative to spiking and replicating samples for unvalidated matrices is the use of isotopically labelled surrogate compounds to establish the recovery of each parameter for each sample. A known amount of the isotopically labelled surrogate compound shall be added to every sample prior to sample analysis. The recovery of the surrogate compound shall fall within acceptable limits and should be reported with the associated sample results.

5.4.6 Estimation of uncertainty of measurement

No additional requirements to ISO/IEC 17025.

Note: Useful information regarding the estimation of measurement uncertainty is given in references h), k) and l).

5.4.6.1 – 5.5.12 No additional requirements to ISO/IEC 17025.

5.6 Measurement traceability

5.6.1 - 5.6.2.1.2 No additional requirements to ISO/IEC 17025.

5.6.2.2 Testing

5.6.2.2.1 Equipment shall be calibrated, and if appropriate with each batch of samples, using measurement standards that are traceable to national or international standards except where they have been derived from natural physical constants, or where this degree of traceability is not possible.

For instrumental methods, calibration solutions may be taken through the entire method or be prepared solely for the determination stage. In either case, solutions shall be matched to the sample extract solution to be determined, both in terms of acid strength and content or solvent composition. In addition, the calibration shall cover the range of interest for the samples being analysed, and should, ideally, be linear over that range. At least three calibration points (not including the calibration blank) are required, but more shall be necessary for a non-linear calibration. Calibration solutions, and standard solutions used for quality control purposes, should, where possible, be prepared by different analysts and from different lots or sources of materials.

The response of instruments may fall due to, for example, deterioration in a detector. This may not be immediately obvious from internal quality control sample results but might coincide with deterioration in both precision and limit of detection of the analytical system. The initial calibration should, therefore, meet with appropriate predefined system suitability limits. Examples include the use of peak area or signal to noise ratio and for chromatographic methods criteria for acceptable peak shape and peak resolution for closely eluting peaks.

Confirmation of the continuing validity of calibration shall be achieved by analysis of calibration check standards regularly throughout the analytical batch according to a defined procedure. The instrument shall not be re-calibrated using the check standard. If a check standard fails to meet appropriate predefined limits the instrument shall be recalibrated and affected samples reanalysed, unless it can be demonstrated that the results would not be affected. Where appropriate, procedures shall be in place to ensure calibration is valid through to the end of an analytical run.

At least one blank sample, containing negligible amounts of the parameters of interest, should be taken through the entire analytical system (including sample preparation if appropriate) with each batch of samples. Laboratories shall demonstrate, according to written procedures, how blank samples are utilised. Blank sample results that show evidence of contamination shall be investigated and may require the analysis of the entire batch of samples to be repeated.

5.6.2.2.2 – 5.6.3.4 No additional requirements to ISO/IEC 17025.

5.7 Sampling

5.7.1 A sample shall be analysed using either all of the sample or a representative or homogenised sub-sample. If a parameter is known to be unstable, or suspected of being unstable, or begins to degrade once the sample has been taken, then the analysis shall be carried out without undue delay. The analysis shall be undertaken on a sub-sample of the sample as removed from the site or preserved or stabilised on site. The results of this analysis shall then be converted to, and reported on, a dry-weight basis of the sample submitted to the laboratory. Thus, the sample shall be analysed on a “wet-weight” or “as submitted” basis, but results reported on a dry-weight basis, and this fact recorded. The procedures used to establish the dry-weight basis shall be defined and reported by the laboratory, as shall drying temperature.

When a sample is stabilised, or preserved and subsequently analysed, then this fact shall be recorded when the results are reported and details of the stabilising or preserving agent shall be recorded. Where a party independent of the analysing laboratory performs this activity (for example the provider of the samples), the laboratory should obtain this information and report it as above.

When a sample has been dried and is subsequently analysed, sufficient information shall be provided to establish the stability of the parameter analysed. Such information shall provide justification for analysing the dried sample, rather than analysing the sample on a "wet-weight" or "as submitted" basis.

Note 1: This information may be in the form of a statement, or describe the work undertaken to justify the approach adopted.

The procedures used to prepare dried samples shall be defined and reported by the laboratory. This may include air-drying and/or assisted-drying. The temperature at which the drying is to be undertaken shall be appropriate to the parameter being determined to ensure that the parameter does not undergo degradation or is lost from the sample. If a sample is to be described as air-dried the drying temperature shall not be more than 30 °C.

When a sample is removed from the site, it often contains a variety of substances and constituents other than the soil and contaminants under investigation. If, after drying, the sample is to be crushed, ground and/or sieved, or certain constituents removed, then appropriate details of the sieve and any material remaining on the sieve, or any constituent parts removed, shall be recorded and reported. These details shall include, for example, the amount, type and nature of such materials.

It shall be noted whether any constituent parts removed or material not crushed, ground and/or sieved undergoes the same analysis as that carried out on material that is crushed, ground and/or sieved. All relevant information (including details of the sieve) shall be provided to establish whether the analysis of the sub-sample relates to all, or constituent parts, of the sample submitted to the laboratory.

Note 2: It is recognised that different practices exist for sample preparation and pre-treatment and that these practices depend on the nature of the sample submitted, the site from where samples are taken and the need for the analysis. It is of paramount importance that relevant information is reported.

5.7.2 - 5.8.4 No additional requirements to ISO/IEC 17025.

5.9 Assuring the quality of test and calibration results

Having verified that the method performance criteria prescribed in Annex A have been satisfied, on-going performance shall be monitored to:

- demonstrate that the method performance required by this performance standard is maintained in a statistically controlled manner
- identify at an early stage any changes (especially deterioration) in performance
- provide historical verification of this performance (i.e. records are kept)
- enable aspects of measurement uncertainty to be estimated.

Carrying out quality control analyses where on-going precision and bias are monitored can fulfil these AQC objectives.

AQC falls into two main categories – external and internal.

The laboratory shall participate in an appropriate external quality control or inter-laboratory proficiency-testing scheme. Where possible, samples from the scheme organiser should reflect typical matrices and parameter concentrations analysed within the laboratory.

Note 1: The Environment Agency will encourage scheme organisers to provide appropriate samples (in terms of matrices, parameters and concentrations of parameters) for distribution that reflect real-life situations and site investigations where data need to be generated.

As far as is possible, the methods, used by the laboratory to generate analytical data for the chemical testing of soil which are submitted under MCERTS, shall be the same as those methods used by the laboratory for the analysis of samples distributed by the proficiency-testing scheme organiser. In addition, as far as is possible, samples distributed by the proficiency-testing scheme organiser should be treated by the laboratory in the same manner as normal routine samples submitted for chemical testing of soils. For example, procedures for registration, storage, analysis and the recording and reporting of results should be similar.

Full details of the scheme, including the number of samples, parameters and analyses to be undertaken by the laboratory and the types of matrices to be analysed, shall be made available. The reports of the results of all analyses submitted by the laboratory to the scheme organiser shall be made available.

The laboratory shall have a documented system in operation to review, investigate and address the results submitted to the proficiency scheme organiser that are considered unsatisfactory, and examine trends in performance. If a significant deterioration in method performance is detected and cannot be corrected within a reasonable period of time the method should be re-validated.

This review procedure should take into consideration the number of other laboratories participating in the scheme and whether these laboratories use the same or similar analytical methods.

For internal quality control, the performance of each analytical method shall be verified for each batch of samples analysed. Control samples shall be analysed within the analytical batch with which they have been prepared.

Note 2: The internal quality control requirements for hydrocarbon banding methods can be found in appendix 5

In each analytical batch, a minimum of 5% of samples shall be laboratory control samples. Laboratory control samples may be certified reference materials, reference materials, in-house reference materials or spiked soils, in that order of preference. If the batch size is less than twenty, one laboratory control sample per batch is still required.

In order to monitor the variation of laboratory control samples, results shall be recorded or plotted on quality control charts (see Annex C). These charts shall be reviewed regularly, and the control limits updated as necessary (see Annex C). To be able to demonstrate statistical control, a minimum of 30 points are required to be plotted in a 12 month cycle, spread evenly over the period.

Note 3: When control limits are updated, estimates of measurement uncertainty should also be updated

If an analytical procedure is carried out infrequently, it shall be necessary to employ a greater degree of AQC to ensure statistical control of the method is maintained. The approach taken shall be fully justified.

Note 4: Examples of greater degree of quality control include increasing the number of control samples in a batch, use of the standard additions approach, and use of isotopically labelled surrogate compounds in organic analysis.

If laboratories use their own reference materials or spiked soils, the actual values used must conform to the traceability criteria as described in ISO/IEC 17025. The following types of control material may be suitable:

- **Certified Reference Material** – A sample of the target matrix, the concentration of parameter being certified to a quoted uncertainty and preferably traceable to an international/national Standard.
- **Reference Material** – A sample of the target matrix, the concentration of determinand having been characterised to a quoted uncertainty.
- **In-house Reference Material** – A sample produced by the laboratory, often by blending soils, each containing known concentrations of different parameters of interest. It is vital that the sample is fully homogenised so that variations in repeat analyses reflect the analytical method performance and not inhomogeneity of the sample. The amount of material should be large enough to provide consistent and stable samples for as long a period as possible. An advantage of using in-house reference materials is the ability to match the parameter concentration and matrix of the material to samples normally encountered in the laboratory.

Note 5: Guidance on the production of in-house reference materials can be found in references o) and p).

Note 6: Traceability for this material may be achieved by characterisation against a certified reference material, for example during method validation or by comparison with the analysis of the material by accredited third-party laboratories.

- **Spiked Soil** – A soil representative of the matrix being analysed, to which a known quantity of a parameter standard solution is added before analysis. A parameter that has been added to the soil may not be present in its “natural” state or fully integrated with the sample matrix, but (for example when the parameter is unstable or volatile) this may be the only option available. Standards used for spiking the sample should be from a different source to that used for calibration. Suitable contact times between spiking and extraction should be determined to provide interaction between spike and sample and to ensure there is no degradation of the parameter.

Note 7: Estimates of bias are often complicated with “recovery” terms, especially if the method involves an extraction stage. An estimate of precision is easily obtainable, but the apparent precision of the spike is a combination of the precision of the sample and that of the spiked sample.

- **Other Options** - Duplicate analyses of individual samples as submitted to the laboratory should be considered when a test is carried out infrequently, as should the use of duplicate control charts. Standard addition techniques may be appropriate. Other alternative procedures or a combination of approaches may be necessary to demonstrate control of infrequently performed tests.

For the individual parameters listed in Annex A (Tables 1-4) quality control results shall be plotted on appropriate control charts.

In multi-parameter organic methods (for example volatile organic compounds (VOCs) or semi-volatile organic compounds (SVOCs)), all single parameters listed in Annex A of this performance standard shall be plotted on a control chart. A minimum of 20% of parameters not listed in Annex A shall be plotted. Groups of parameters with similar properties shall have representative parameters selected for plotting. The selection of these parameters should include critical parameters, for example those most likely to be laboratory contaminants, and for chromatographic methods, parameters that elute at the beginning and end of a chromatogram, or those whose peaks are poorly resolved. Laboratories shall justify their approach. All other parameters, that is those not selected for plotting on control charts, shall be present in the laboratory control sample at appropriate concentrations and be determined. These results shall be recorded and shall be reviewed as part of regular AQC performance review. Precision and bias shall not statistically exceed 15% precision and 30% bias.

Laboratories shall have documented procedures that define loss of statistical control and specify actions to be taken (control rules) when control limits are breached. All breaches shall be investigated, and the findings and actions recorded and made available to the Environment Agency, if requested. Samples in an analytical batch where laboratory control samples breach the defined control rules shall be reanalysed.

The investigation shall include but shall not be restricted to the following checks:

- changes in concentration of stock standard solutions and reagents and that expiry date has not been exceeded
- calibration of instruments used in the analytical process
- documented methods were strictly adhered to
- that system suitability check data meet requirements
- significant drift does not occur for automated determinations
- service/fault records
- recent proficiency testing scheme results.

Records shall include:

- identification of control sample and all associated sample results
- control rules in force at time of breach and breach result
- investigation details, conclusions and actions taken
- action taken with respect to affected sample results (i.e. analysis repeated or results reported – see Annex C3).

5.10 Reporting the results

5.10.1 – 5.10.3 No additional requirements to ISO/IEC 17025.

5.10.3.1 For data submitted to the Environment Agency for regulatory purposes, appropriate information shall be included in the report that clearly identifies and locates the sample relating to the results. This information shall require the recording of all data necessary to allow a complete audit trail to be made. Relevant information includes:

- location of sample, including depth where necessary
- unique sample code or reference
- date/time sample taken
- name of laboratory
- name of any sub-contracting laboratories, if used
- date sample analysis completed
- parameter analysed, including whether sample was preserved or stabilised at sampling site
- whether analysis carried out on dried, air-dried or “as submitted” basis
- result of analysis on dry-weight basis
- other relevant comments, for example, visual characteristics of sample.

Note: It is recognised that some of this information may only be available from, or be able to be provided by, the procurer of the analytical service and not the laboratory.

However results are determined all values shall be calculated and reported on a dry-weight basis. The laboratory shall report the procedures used, how air-dried and dry-weight have been determined and defined, and the drying temperature.

Results that are submitted to the Environment Agency shall be accompanied with a statement indicating whether the results have been recovery corrected or not, as the case may be, and if so, the criteria used, including the manner of calculation.

Whenever possible and where appropriate, individual compounds should be analysed and individual results reported. Where a group of similar compounds is analysed and the combined concentrations of these compounds are expressed as the sum of individual concentrations, the laboratory shall record the number and identity of each compound analysed. This information shall be reported with the results. If this approach is not possible or appropriate, the laboratory shall define the analysis undertaken and the calculated result. This information shall be reported with the result.

5.10.3.2 - 5.10.9 No additional requirements to ISO/IEC 17025.

6. Status of this document

- 6.1** Version 4 of the MCERTS performance standard replaces version 3, which is withdrawn and should now be discarded. Version 4 will be subject to review and amendment following publication. The latest version of the standard is available on the Environment Agency’s web site at: www.mcerts.net
- 6.2** If you have any questions regarding the accreditation process, including how to make an application, please contact UKAS at the address given in the foreword to this document.

Annex A (normative): Performance Characteristics

Table 1 - Metals and organometallics

Parameter ¹	Precision ²	Bias ³
antimony	7.5	15
arsenic	7.5	15
barium	7.5	10
beryllium	7.5	10
boron (water soluble)	10	20
cadmium	7.5	10
cobalt	7.5	10
copper	7.5	10
chromium	7.5	10
hexavalent chromium	7.5	10
iron	7.5	10
lead	7.5	10
manganese	7.5	10
mercury	7.5	15
molybdenum	7.5	10
nickel	7.5	10
organolead compounds	15	30
organotin compounds	15	30
selenium	7.5	15
thallium	7.5	10
vanadium	7.5	10
zinc	7.5	10

Notes

1. Whilst no limit of detection has been specified, it shall be fit for purpose, especially, for example when compared to “soil guideline values” or critical levels of interest.
2. Precision expressed as percent relative standard deviation.
3. Bias expressed in percentage terms.

Table 2 - Inorganics

Parameter ¹	Precision ²	Bias ³
ammonia	10	20
chloride	10	20
easily liberatable cyanide	15	30
complex cyanide	15	30
total cyanide	15	30
loss on ignition	7.5	15
pH	0.2	0.2
sulfide	15	30
sulfate	10	20
sulfur	10	20
thiocyanate	15	30

Notes

1. Whilst no limit of detection has been specified, it shall be fit for purpose, especially, for example when compared to “soil guideline values” or critical levels of interest.
2. Precision expressed as percent relative standard deviation except for pH, which is in terms of pH units.
3. Bias expressed in percentage terms except for pH, which is in terms of pH units.

Table 3 - Organics

Parameter ¹	Precision ²	Bias ³
benzene	15	30
benzo[a]pyrene	15	30
chlorobenzene	15	30
chloromethane	15	30
chlorophenol	15	30
chlorotoluene	15	30
dichloroethane	15	30
1,2-dichloroethene	15	30
dichloromethane	15	30
dioxins⁵	15	30
explosive substances⁴	15	30
ethylbenzene	15	30
furans⁵	15	30
hexachloro-1, 3-butadiene	15	30
petroleum hydrocarbons⁶	15	30
nitroaromatics⁵	15	30
pentachlorophenol	15	30
phenols⁵	15	30
phthalate esters⁵	15	30
polyaromatic hydrocarbons⁵	15	30
polychlorinated biphenyls⁵	15	30
tetrachloroethane	15	30
tetrachloroethene	15	30
tetrachloromethane (carbon tetrachloride)	15	30
total organic carbon (soil organic matter estimation)	10	20
toluene	15	30
trichloroethane	15	30
trichloroethene	15	30
trichloromethane (chloroform)	15	30
vinyl chloride	15	30
xylene	15	30

Notes

1. Whilst no limit of detection has been specified, it shall be fit for purpose, especially, for example when compared to “soil guideline values” or critical levels of interest.
2. Precision expressed as percent relative standard deviation.
3. Bias expressed in percentage terms.
4. Covers organic explosive compounds as listed in Environment Agency guidance.
5. Performance targets are for individual compounds within these groups. If a total (e.g. total PAH) result is requested, then each individual component should be determined and reported with the total. See section 5.10.3.1.
6. Requirements for validation of hydrocarbon banding methods can be found in appendix E

Table 4 – Additional parameters

Method/Parameter	Precision (%)	Bias (%)
extractable phosphate content (Olsen)	10	15
extraction of the exchangeable cations in soil: potassium	10	15
extraction of the exchangeable cations in soil: magnesium	10	15
extraction of the exchangeable cations in soil: sodium	10	15
organic carbon content % modified Walkley Black	5	10
determination of electrical conductivity	5	15
metals by nitric acid extraction and microwave digestion	7.5	10
total nitrogen	5	15
determination of carbonate content – volumetric method	6	10
fluoride	10	20

Notes

1. Whilst no limit of detection has been specified, it shall be fit for purpose, especially, for example when compared to “soil guideline values” or critical levels of interest.
2. Precision expressed as percent relative standard deviation.
3. Bias expressed in percentage terms.

Annex B (normative): Statistical Analysis

B1 Limits of detection and reporting

B1.1 Introduction

The limit of detection (LOD) is widely but inappropriately used as the primary performance measure of an analytical system, but does not indicate whether a method is fit for purpose. For example, a very low LOD value does not mean that the method is suitable for a particular purpose (See also section 5.4.5.3), as precision and bias could be unacceptable at the critical level of interest. The LOD is not specified in this performance standard. However, a common approach to the estimation of LOD is desirable in order to allow a laboratory's performance to be evaluated in a consistent and comparable way. If data reported to the Environment Agency are to include results reported as less than values, the LOD shall be estimated using the following protocol.

For further guidance on estimation of LOD for hydrocarbon banding methods see Appendix E2.3

B1.2 Choice of sample and sample pre-treatment

The blank sample used to estimate LOD shall be a soil containing a small but measurable amount of parameter(s) of interest. If it can be demonstrated that a suitable soil cannot be obtained then a sand containing a negligible amount of parameter of interest shall be used.

Note1: The sample used for estimating LOD should be as similar as possible to the matrix being analysed. Using a single sample for the determination of LOD for a given method will not take into account different matrix effects.

Ideally analysis of the blank sample will produce normally distributed results scattered around zero; i.e. both negative and positive results will be generated. It is usually possible for the blank sample to have a sufficiently small background concentration of the parameter to fulfil this requirement. However, this may not always be possible because in some analytical systems negative or low results cannot be obtained. In these cases the blank sample should be spiked with a small amount of the parameter, sufficient to produce a small but significant response from the analytical system, i.e. close to the expected LOD. This concentration shall not exceed 5 times the LOD.

The blank or spiked sample shall be put through the entire analytical process (including, as necessary, drying, grinding, extraction, cleanup and measurement). The extraction and measurement of blank solutions based only on solvent or reagent blanks is not sufficient for estimating LODs for the purpose of satisfying MCERTS requirements. The blank samples or spikes shall be processed in the same manner and using the same equipment and reagents as other samples in a batch.

Note2: For commonly occurring substances such as iron, zinc, chloride and sulfate etc., where soils may contain a significant amount of these substances, the method used to determine an LOD for that substance using a pure sand or silica blank can give an optimistic (lower concentration) LOD. Alternatively if an "uncontaminated natural" soil is used to determine the LOD and it contains a significant amount of these substances then a pessimistic (higher concentration) LOD will be obtained.

Note3: It is important that users of results should appreciate that the LOD for these common substances obtained by all MCERTS accredited laboratories should be adequate for all these commonly (naturally) occurring substances. However, it is unlikely that LOD will be an issue with these substances, as adequate precision and bias at the level of interest is more pertinent.

Note4: For commonly occurring substances the variation in blank values should be consistent and within acceptable limits. Ideally all blank values for these substances should be less than 10% of the critical level of interest.

B1.3 Calculation

For the purpose of this performance standard, LOD is defined by the equation:

$$\text{LOD} = 2\sqrt{2} \cdot t_{(df, \alpha=0.05)} \cdot s_w$$

where:

df is the number of degrees of freedom (minimum 10)

t is the one-sided Student's t-test statistic (95% confidence level)

s_w is the within-batch standard deviation of results from samples ideally containing negligible concentration of the parameter of interest.

An estimate of the LOD can be made when initial validation studies are undertaken. Pairs of sample blanks shall be analysed in at least 10 different analytical runs or batches. Ideally these blanks should contain a negligible amount of the parameter being determined and should be consistent with and similar to the matrices of the samples being analysed. These sample blanks shall not be used as a calibration blank, and if the analytical procedure requires samples to be blank corrected, then the sample blanks used to estimate LOD should also be blank corrected.

Results shall not be rounded before being used for the estimation of LOD.

In the most general case, where **m** batches of different numbers of replicates **n_i** give a series of within-batch standard deviations **s_i**:

The pooled value of **s_w** is given by:

$$s_w \text{ (pooled)} = \sqrt{\frac{\sum s_i^2 \cdot (n_i - 1)}{\sum (n_i - 1)}}$$

where:

s_i = individual batch standard deviation,

n_i = number of results in the batch.

Where the batches all contain the same number of results, this equation simplifies to:

$$s_w \text{ (pooled)} = \sqrt{\frac{\sum s_i^2}{m}} \text{ with } m(n-1) \text{ degrees of freedom}$$

for example for 10 batches of 2 blanks:

$$s_w \text{ (pooled)} = \sqrt{\frac{\sum s_i^2}{10}} \text{ with 10 degrees of freedom}$$

Since **t_(α=0.05)** for a one sided t-test with 10 degrees of freedom is 1.812

Then $LOD = 2\sqrt{2} \cdot t \cdot s_w = 5.13s_w$

If a different number of batches and replicates is used a minimum of 10 degrees of freedom shall be obtained. Where more than 10 batches of replicates are determined, all valid results shall be used in calculating the LOD.

As an ongoing check, an estimate of LOD can be obtained by analysing 11 blank samples in the same batch, here s_t (total standard deviation) equates to s_w , with 10 degrees of freedom. This procedure should be used when a matrix is analysed by a method that has not been fully validated for that matrix.

B1.4 Form of expression

For a multi-parameter method such as PAH, each individual PAH will need to have its own LOD estimated.

For TPH and similar parameters, it would not be appropriate to estimate the LOD using just one of the hydrocarbons within the analytical range. Blank sample data shall be generated in the same way as normal sample data to obtain the results used in estimating LOD.

LOD values shall always be reported in the same units as the parameters they represent. The calculated value may be rounded up for convenience and ease of use.

B1.5 Reporting limit

For the purposes of this MCERTS performance standard the reporting limit will be the limit of detection calculated as above. However, a laboratory may use higher reporting limits than calculated LODs, for example a laboratory calculated LOD for a method as <0.2 mg/kg but prefer to report <1 mg/kg due to issues with reporting software and customer requirements. This is considered wholly acceptable by the Agency, as long as LOD is calculated in the correct way.

If samples are diluted before analysis then the LOD must be scaled up, i.e. if a sample is diluted 1:5, and the analytical result is <5, then <25 should be reported.

B2 The use of statistical significance tests in the interpretation of method performance.

B2.1 Introduction

Method validation aims to produce data on the precision of analysis and to provide an indication of any susceptibility to systematic error or bias.

Assuming that validation has been carried out as described in section 5.4.5 and that Analysis of Variance (ANOVA) has been applied to the results, there should be sufficient data to assess whether method performance complies with Annex A criteria (see section 5.4.5.3).

B2.2 Assessment of precision

The convention in analysis has been to consider precision to be satisfactory if the measured standard deviation is found not to be statistically significantly larger than the target standard deviation.

This implies there is uncertainty about the measured standard deviation value, although this uncertainty is minimised by specifying its calculation with at least 10 degrees of freedom.

Assessment of precision is in three stages:

1. Determine the target standard deviation at the concentration of interest, in accordance with section 5.4.5.3.
2. If the measured standard deviation is less than the target standard deviation, the target has been achieved.
3. If, however, the measured standard deviation is greater than the target it is still possible to comply with the requirements of this standard if it is not significantly greater. To assess this significance a statistical test is required.

B2.3 F-Test of standard deviation.

The F-test or variance ratio test is a way of determining whether or not differences between two standard deviations are statistically significant (at a chosen probability level). The procedure is to calculate the F ratio as shown below:

$$F = \frac{s_t^2}{Z^2}$$

where s_t is the measured total standard deviation, estimated using between batch and within batch mean squares in ANOVA, and Z is the target standard deviation.

The calculated value of F is then compared with a reference value obtained from statistical tables. The reference value of F is obtained using the correct probability (5% for this performance standard) and using the relevant degrees of freedom for s_t and Z .

Z is a target standard deviation and therefore has infinite degrees of freedom. In the case of s_t , the number of degrees of freedom is calculated during the analysis of variance. If a complete 11x2 validation is performed, the equation can be simplified to:

$$df = \frac{110[M_1 + M_0]^2}{11M_1^2 + 10M_0^2}$$

where M_0 and M_1 are the within batch and between batch mean squares respectively, each obtained from ANOVA.

If the F ratio is less than the tabulated reference F value then the measured standard deviation is not significantly greater than the target value i.e. performance passes

If the F ratio is greater than the tabulated reference F value then the measured standard deviation is significantly greater than the target value i.e. performance is not satisfactory.

B2.4 Assessment of systematic error or bias

This assessment is only relevant and should only be carried out if the assessment of precision is acceptable.

The assessment of bias depends on independent knowledge of a “true” value with which to compare the average of measured data. This is accomplished by the use of reference materials or by spiking recovery experiments.

To assess bias and its associated uncertainty the procedure is to calculate the mean recovery for each batch and to use the batch mean recoveries to estimate the overall recovery and its standard deviation (strictly its standard error).

Significance is assessed by means of calculating the confidence interval about the mean and checking to see if this overlaps the limits of tolerable bias.

$$\text{Overall Mean Recovery} = \frac{\sum R_i}{m} = \mathbf{M}$$

$$\text{Standard Error of Recovery} = \frac{S_R}{\sqrt{m}} = S_e$$

$$90\% \text{ Confidence Interval of Recovery} = \mathbf{M} \pm S_e \times t_{(0.05, m-1)}$$

where

m = number of batches

R_i = %Recovery of the i th batch

S_R = standard deviation of batch recoveries

$t_{(0.05, m-1)}$ = single-sided Student's t value at 5% probability level
and $(m-1)$ degrees of freedom

If there is an overlap (i.e. one or both of the target recovery limits is within the confidence interval), the recovery is not significantly worse than required and should be regarded as acceptable.

Note: When a bias is estimated it is either positive or negative, therefore a one sided t-test at the 95% confidence level is used to assess if observed bias is greater than permitted bias. However, by definition, a confidence interval is two sided, therefore the significance test is at the 95% confidence level but the resulting confidence interval is 90%.

Examples

Two examples are presented below to illustrate the application of the statistical tests mentioned above. The first considers a spiking exercise for cadmium, the second a CRM used to validate a method for benzo(b)fluoranthene.

Example 1: cadmium mg/kg in soil – spiked samples

	Test sample	Recovered low spike sample	Recovered High spike sample	Batch mean values	Batch mean %recovery	Batch mean values	Batch mean %recovery
	spiked Concentration	4 mg/kg	40 mg/kg	for low sample	For low sample	for high sample	for high sample
Batch	Replicate						
1	1	3.60	47.0	3.705	92.625	47.75	119.38
1	2	3.81	48.5				
2	1	3.83	42.6	3.895	97.375	42.85	107.13
2	2	3.96	43.1				
3	1	4.02	47.5	4.06	101.5	48.4	121
3	2	4.10	49.3				
4	1	4.12	44.0	4.21	105.25	45.05	112.63
4	2	4.30	46.1				
5	1	4.05	47.02	3.945	98.625	46.695	116.74
5	2	3.84	46.37				
6	1	3.70	40.12	3.805	95.125	40.405	101.01
6	2	3.91	40.69				
7	1	3.44	41.93	3.39	84.75	41.625	104.06
7	2	3.34	41.32				
8	1	3.68	43.4	3.755	93.875	44.135	110.34
8	2	3.83	44.87				
9	1	3.85	42.19	3.825	95.625	42.57	106.42
9	2	3.80	42.95				
10	1	3.52	43.46	3.42	85.5	43.23	108.08
10	2	3.32	43.0				
11	1	4.02	43.81	3.96	99.0	44.075	110.09
11	2	3.90	44.34				
	Overall mean			3.815		44.25	
	Overall mean recovery				95.39		110.63

Precision test		Low sample	High sample
Mean		3.815	44.25
Within-Batch sd		0.112	0.812
Between-Batch sd		0.234	2.46
Total sd		0.26	2.58
Relative sd %		6.8%	5.9%
Target sd	5% of mean	0.19	2.21
F 0.05 from tables		1.75	1.79
F-Value calculated		1.86	1.37
Estimate degrees freedom		12	11
Assessment		FAIL	PASS
Overall mean recovery		95.39%	110.63%

This is obtained from statistical tables for the estimated degrees of freedom at the 5% probability level ($p=0.05$)

This value is calculated as $(\text{total sd} / \text{target sd})^2$.

In this example the observed standard deviation of the low concentration sample is greater than the target standard deviation, so an F test is performed. F calculated is greater than the tabulated reference F value so the standard deviation of the low sample is significantly different than 5% and therefore fails to meet the MCERTS target. In the case of the high concentration sample the measured total sd is larger than target but the F test shows that this is not significantly larger – hence this is judged to meet MCERTS requirements.

Recovery for high sample	
Reference concentration	40
Mean measured value	44.25
Overall mean recovery	110.63%
sd of mean recovery	6.306
Standard error of mean recovery	1.901
90 % Confidence interval of recovery	+/-3.44
Recovery range	107.2% - 114.1%
Assessment	PASS

This value is the average of the mean recovery for each batch

This value is the relative sd of overall mean recovery divided by the square root of the number of batches

This value is the standard error of mean recovery multiplied by the Student's t value ($p=0.05$ single sided) for degrees of freedom equal to number of batch - 1 ($t=1.812$ for 11 batches)

The bias target for cadmium is 10% so the tolerable range of recovery in this example is 90-110%. In the case of the high sample this overlap of confidence interval with the tolerable range means that although recovery is nominally outside this range it is not significantly so and is therefore acceptable. **Note that the precision must be acceptable before this test can be applied, so it would not be appropriate to test the low sample.**

Example 2: CRM for Benzo(b)fluoranthene

	Test sample	CRM		
	Certified Concentration	26 µg/kg	Batch mean values	Batch mean %recovery
Batch	Replicate			
1	1	18.4	18.75	72.12
1	2	19.1		
2	1	17.2	18.3	70.38
2	2	19.4		
3	1	21.6	20.5	78.85
3	2	19.4		
4	1	15.8	16.25	62.5
4	2	16.7		
5	1	17.4	19.4	74.62
5	2	21.4		
6	1	18.0	18.45	70.96
6	2	18.9		
7	1	16.8	17.1	65.77
7	2	17.4		
8	1	17.6	18.15	69.81
8	2	18.7		
9	1	16.2	16.0	61.54
9	2	15.8		
10	1	16.2	17.25	66.35
10	2	18.3		
11	1	17.4	16.8	64.62
11	2	16.2		
Overall mean			17.91	
Overall mean recovery				68.86

Precision test		
Mean		17.91
Within-Batch sd		1.27
Between-Batch sd		1.04
Total sd		1.64
Relative sd %		9.16%
Target sd	15% of mean	2.7
F 0.05 from tables		
F-Value calculated		
Estimate degrees freedom		
Assessment		PASS
Estimated Bias		-31.14%

In this example the requirements for precision has been met without the need for significance testing. However, the bias appears to be outside of the 30% target. As precision is acceptable the significance test for bias can be carried out.

Bias test	
Reference concentration	26
Mean measured value	17.91
Overall mean recovery	68.9%
sd of mean recovery	5.2823
Standard error of mean recovery	1.5927
90% Confidence interval of recovery	+/-2.9
Recovery range	66.0% - 71.8%
Assessment	PASS

The calculated recovery range overlaps with the required range of 70-130% so the bias is not significantly different to the MCERTS requirement, at the 95% confidence level ($p=0.05$).

Annex C (normative): Production and use of Control Charts

C1 Introduction

The interpretation of results from the analysis of internal laboratory control samples is usually carried out using control charts. These charts are used to compare current results against limits set after estimating the variability of an analytical system operating under statistical control. A method is said to be in statistical control when the variability within the analytical system arises from a stable sources of random analytical variability. Various forms of control chart may be appropriate for use, for example:

- Shewhart charts (individual result) the most common in use
- cusum charts – more sensitive to bias detection than Shewhart charts
- zone control chart (J-chart) – combines Shewhart and cusum charts capabilities.

As a minimum a Shewhart chart should be used, as described below (the use of other charts is described in references b), m) and n)).

C2 Setting up and updating Shewhart control charts

Control charts should be set up using estimates of mean (**M**) and standard deviation (**sd**) obtained from results of at least 20 control samples that are analysed when the analytical system is under statistical control. This data shall initially have been obtained during method validation procedures.

The properties of the normal distribution allow the prediction that for on-going analysis, 95% of results will fall within $M \pm 2sd$ and that 99.7% of results will fall within $M \pm 3sd$ given no deterioration in method performance.

An example is shown in Fig.1. The chart is constructed as follows:

The y-axis is concentration, the x-axis time (that is date of analysis)
The mean laboratory control standard value **M** is plotted as a line (“mean”)
Two warning limits are plotted as lines at $M \pm 2sd$ (“2SD”)
Two Action limits are plotted as lines at $M \pm 3sd$ (“3SD”)
The laboratory control standard nominal value may be plotted as a line (“value”).

As laboratory control sample results become available, they are plotted individually and consecutively against date of analysis. They shall not be averaged before plotting.

As they are plotted, control rules to indicate a system failure include:

- one laboratory control Standard result outside the control chart action limit
- two consecutive laboratory control Standard results outside the control chart warning limit.

In addition, nine successive laboratory control Standard results on the same side of the chart mean could indicate a change in the bias of the analytical system, and should be investigated. However, this may be due to a small insignificant change and laboratories should use other methods of identifying significant changes in bias. For example, if a number of control standards are measured in each analytical batch of a high throughput method, then the average value can be plotted to monitor bias only. However, a separate chart of individual values shall still be required for control of precision.

C3 Reporting

Results associated with failed AQC samples shall not be reported as MCERTS accredited results. In some circumstances customers may request that results be released. Whenever results associated with failed AQC are required by the customer they should only be issued under the direct authority of an appropriate manager.

Any report issued, which contains results associated with failed AQC samples shall include a printed disclaimer as to this effect.

Annex D (informative): Estimation of soil organic matter (SOM)

The Environment Agency has published soil guideline values (SGVs) to aid the risk assessment of contaminated land. SGVs for contaminant organic compounds such as toluene, ethylbenzene and phenol address the fact that these compounds tend to adsorb onto soil organic matter (SOM), and so plant uptake and volatilisation to air are reduced. As SOM increases less contaminant is available for these exposure pathways. The SGV calculated for these compounds therefore varies with the SOM, the higher the SOM the higher the SGV. It follows therefore that SOM needs to be estimated.

The Environment Agency considers that the most appropriate way to proceed is to measure the fraction of organic carbon (foc) and then calculate an estimate of the SOM.

The definition of SOM used in the CLEA model is:

$$\%SOM = foc \times (100/0.58)$$

This assumes that SOM has a carbon content of 58%.

A method of estimation of foc is via the determination of total organic carbon (TOC) after prior removal of inorganic carbon with acid by dry combustion at 900°C, and measurement of released carbon dioxide. This does not preclude other appropriate analytical methods.

$$foc = TOC \times 10^{-6} \quad \text{where TOC is expressed in units of mg/kg}$$

$$\therefore \%SOM = TOC / 0.58 \quad \text{where TOC is expressed as a percentage}$$

An estimate of SOM can also be obtained from the loss on ignition method. However, laboratories will need to develop a conversion factor that will depend on the matrix of the soil being analysed and the temperature of ignition.

Annex E (normative): Accreditation of hydrocarbon banding

E1 Introduction

Many laboratories have obtained accreditation for TPH (total petroleum hydrocarbons) or EPH (extractable petroleum hydrocarbons). However, reporting of petroleum hydrocarbon concentrations in soils using bands that are defined by carbon number is becoming increasingly important for contaminated land assessment.

Note 1: The term EPH is often used to describe the fraction C₁₀ to C₄₀, and VPH (volatile petroleum hydrocarbons) the fraction C₆ to <C₁₀ and the sum of the two fractions being reported as TPH. This convention is not strictly adhered to and some laboratories quote different ranges. It is not the purpose of this document to define these terms or fractions or how these fractions are divided into specific bands. Laboratory methods and their associated scopes should do this. The requirements of the Environment Agency will be stated elsewhere.

Note 2: TPH analysis is usually performed on an as received sample, rather than a dried and crushed sample, to minimise losses of the volatile fraction.

E2 Validation

E2.1 Hydrocarbon banding (without separation into aliphatic and aromatic fractions)

A laboratory requiring accreditation for hydrocarbon banding without separation into aliphatic and aromatic fractions shall use the following procedure:

- The bands that are reported will contain both aliphatic and aromatic hydrocarbons.
- The beginning and end for integration of each band shall be defined by running a mixture that contains n-alkanes (straight chain, saturated hydrocarbons) with carbon numbers whose range covers the bands defined in the method. [Note: It is not acceptable to report bands to odd carbon numbers using solely even numbered hydrocarbons]
- A minimum of three soil matrices shall be used in the validation.
- If available and appropriate, matrix CRMs shall be used.
- The performance characteristics shall be determined with a minimum of 10 degrees of freedom by analysing batches of duplicates. If CRMs are not available, each matrix should be spiked at two significantly different but appropriate concentration levels, for example, at 20 % and 80 % of the range of the method, using a spike composed predominantly of a mixture of petroleum hydrocarbon fractions to ensure that there are adequate amounts of appropriate hydrocarbons in each of the bands for which accreditation is sought. This oil mixture may be fortified with n-alkanes or PAHs if necessary but use of a spike containing individual n-alkanes or PAHs alone or containing predominantly individual n-alkanes or PAHs alone is inappropriate and shall not be used.
- Soil matrices should be extracted, analysed and the resulting chromatogram interpreted using the method for which accreditation is sought.

Performance targets:

- The precision of the method for the sum of the bands shall not significantly exceed 15% and the bias of the method for the sum of the bands shall not significantly exceed 30%.
- The precision of the method for each individual band shall not significantly exceed 15%.

E2.2 Aliphatic and aromatic fractions and subsequent banding

The validation protocol is exactly the same as that outlined in E2.1 above and the performance targets are as follows:

- If total TPH is derived from summing the aliphatic and aromatic fractions, or the bands of those fractions, then the precision of the method for the sum of the bands shall not significantly exceed 15% and the bias of the method for the sum of the bands shall not significantly exceed 30%.
- The precision for the aromatic and aliphatic fractions shall not significantly exceed 15%.
- The precision of the method for each individual band shall not significantly exceed 15%.

In this case the precision obtained by laboratories for each band will be reviewed and changed if deemed necessary.

Accreditation for banding and splitting into aliphatic and aromatic fractions and subsequent banding will only be granted if each and all of the bands within the range defined by 'total' EPH/TPH can be shown to meet the specified targets.

E2.3 Limit of Detection (see also Annex B1)

A soil containing a small but detectable amount of the parameter of interest shall be analysed for determining the limit of detection for TPH [total, aromatic/aliphatic split and banded] in soils. This sample can be prepared by spiking.

The low spiking solution for LoD estimation should use a spike composed predominantly of a mixture of petroleum hydrocarbon fractions ensuring that there are adequate [but not greater than 5 times the resulting LOD] amounts of appropriate hydrocarbons in each of the bands for which accreditation is sought. This oil mixture should be the same as that used for the spiking experiments and may be fortified with n-alkanes or PAHs if necessary. Use of a spike containing individual n-alkanes or PAHs alone or containing predominantly individual n-alkanes or PAHs alone is inappropriate and shall not be used.

The low spiked sample shall be put through the entire analytical process (including, as necessary, drying, grinding, extraction, cleanup and measurement). The extraction and measurement of blank solutions based only on solvent or reagent blanks is not sufficient for estimating LODs for the purpose of satisfying MCERTS requirements. The low spiked samples shall be processed in the same manner and using the same equipment as other samples.

The limit of detection shall be calculated as described in Annex B1 of the Performance Standard for Laboratories Undertaking Chemical Testing of Soil, version 3, 2006.

E3 Quality assurance for hydrocarbon banding; aliphatic and aromatic split and subsequent banding

E3.1 Quality control

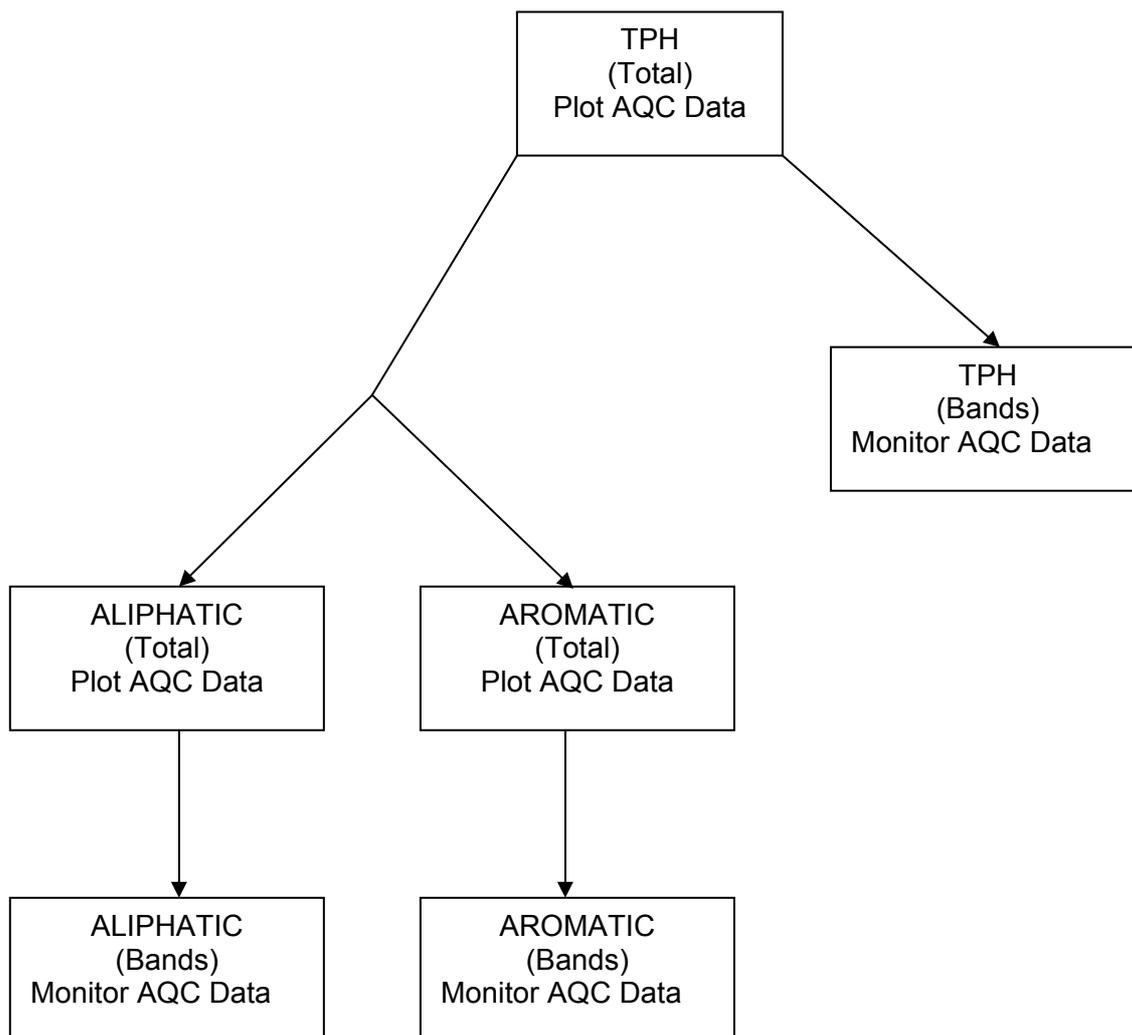
AQC samples shall be representative soils spiked at an appropriate level with an oil mixture as described in the validation protocol.

This shall be extracted and analysed, a minimum of 1 sample in every 20 samples shall be an AQC sample. AQC performance targets are 15% for precision and 30% for bias.

With respect to the evaluation of AQC results obtained the minimum that is expected is outlined in Diagram 1.

Note 3: The frequency at which AQC data is monitored (where plotting is not mandatory) is at the discretion of the laboratory. However, if at the monitoring/review stage laboratories find that the targets have not been met any associated results shall be treated as non-conforming work that may necessitate the re-issuing of reports. If a laboratory wishes to plot for all AQC's analysed, this is acceptable.

Diagram 1



For AQC data required to be plotted it shall be plotted on a chart and the statistical acceptance rules as detailed in Annex C shall be followed. For AQC data required to be monitored, AQC samples shall include the parameters of interest and AQC results shall be recorded. Performance of monitored AQC shall be reviewed as part of regular AQC performance review and precision and bias shall not statistically exceed the targets given in this Annex.

E3.2 System suitability

Where splitting into aliphatic and aromatic fractions is performed there are additional requirements with respect to system suitability checks to ensure adequate separation column efficiency.

With every batch of new column material a synthetic aliphatic and aromatic mixture shall be separated and analysed. This mixture shall cover the range of hydrocarbons analysed and shall be composed of aliphatic and aromatic components. A full range of hydrocarbons shall be present in the mixture and shall include critical compounds in each fraction (such as naphthalene, decane and corresponding compounds in the C30- C40 range). Recoveries for the aliphatic and aromatic fractions shall be estimated concurrently with the method validation, and used to set limits for subsequent analysis. As a minimum, records of the recovery of the critical compounds in each fraction shall be maintained.