Marine Environmental Depleted Uranium Survey Report Kirkcudbright Training Area - 2005

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Executive summary

Depleted uranium (DU) ammunition has been test fired at the Kirkcudbright Training Area (KTA) since 1982. Routine environmental monitoring has been carried out at KTA since 1980 to assess the environmental impact of the firings on the terrestrial and marine environments. Results of these surveys are published in regular reports.

This report describes and interprets the results of the marine environmental survey undertaken to monitor DU levels in the environment at KTA in 2005. Samples of intertidal sediments, seaweed and seafood were collected from the shore line in the local area. Underwater sediment samples were collected off shore from the KTA, and samples of locally fished, bottom dwelling crustaceans and molluscs were also obtained.

The results of the 2005 survey do not show evidence of the presence of DU in any of the environmental samples collected. There is no evidence to indicate that members of the public or site employees are, or have ever been, exposed to a radiological hazard from the marine environment as a result of test firing DU ammunition at the site.

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1 Introduction

This report presents and interprets the results of the 2005 routine monitoring programme for Depleted Uranium (DU) in the marine environment bordering and offshore from the Kirkcudbright Training Area (KTA). Intertidal shoreline samples and marine biota samples were collected, and environmental dose rate measurements were made in December 2005, but mechanical problems onboard the survey boat delayed the collection of underwater sediments and crustaceans until January 2006.

Routine marine environmental monitoring has been carried out annually since 1980 (i.e. before firing began) to measure uranium levels in the marine environment at the KTA and to identify the extent of any dispersion of DU that might result from operations at the site. The monitoring programme has consisted principally of the collection of seawater, marine sediment, seaweed and marine animals, which were subsequently analysed in the laboratory.

Over the years, the monitoring programme has changed to reflect increasing knowledge of the local conditions and changes in best practice, as may be expected over twenty-seven years, resulting in differences such as sample type and reporting units. However, sample type, analysis technique and reporting have been broadly consistent since 1996. In addition, there have been the following key changes.

Sea water and sediments were collected from 6 off-shore sites in the Solway Firth, each year from 1980 to 2001. In 2001, it was recognised that the immense dilution present in the Solway would never give rise to uranium levels in sea water that were detectable above background. Therefore, with the agreement of the Scottish Environmental Protection Agency (SEPA), the collection of seawater ceased. Instead, seabed sediment samples were collected from a greater number of sites (20). The number of underwater sediment samples for some sites was doubled to reduce analytical uncertainties. This protocol was kept in 2002 and 2003. In 2004 and 2005 the number of sampling sites was increased again and the area of underwater sediments sampling enlarged to improve the alignment of the sampling sites with changes in work practices during the most recent test firings undertaken in 2003. The number of sampling sites planned for 2005 was 25 and their locations more accurately reflect the splash down areas of the rounds. The number of samples collected at each site was reduced to one, in view of the improved geographical positioning.

The yearly sampling of sediments, seaweed and mussels, and the measurement of dose rates, from the intertidal shoreline of the Dumfries coast were added from 1994 onwards, as was the analysis of locally caught seafood. The total number of shore locations sampled in each survey is 11.

The monitoring for DU in the terrestrial environment undertaken at Kirkcudbright during 2005 is reported separately [1].

2 Background

The KTA range is located on the coast of Dumfries and Galloway, near Castle Douglas. Since April 2006, the range has been part of the Defence Training Estate.

DU has been released into the environment at Kirkcudbright as a consequence of the test firing of DU ammunition during design and accuracy assessment trials. The trials involve firing DU projectiles through soft vertical targets into the Solway Firth.

Testing of projectiles historically has taken place at five locations on the Kirkcudbright site. Each battery location had a designated target and line of fire, and hence a predictable impact area ranging from several hundred metres to about 7 km offshore. Although a small fraction of the total number of penetrators malfunctioned and impacted on land, the vast majority of malfunctioning penetrators, and all the functioning penetrators, entered the Solway Firth. The number of DU rounds fired each year at Kirkcudbright from the five firing locations and the cumulative mass of DU fired to date, are presented in Figure 1 and Figure 2 respectively.

3 Depleted uranium

There are three broad types of uranium: natural, enriched and depleted. Uranium is a naturally occurring radioactive material that is found as a mixture of three isotopes: uranium-238 (^{238}U), uranium-235 (^{235}U) and uranium-234 (^{234}U). It emits alpha and beta particles, gamma and X radiation. Uranium, in an 'enriched' form, is used as fuel in nuclear reactors. It is called enriched because, due to processing, it contains a higher concentration (> 0.72%) of ^{235}U than natural uranium. The by-product of the enrichment process is depleted uranium (DU), which has a reduced concentration of ^{235}U. Some ^{234}U is also removed in the depletion process. Because almost all (>99% by mass) of the uranium in natural and depleted uranium is ^{238}U, the ^{238}U concentration and therefore the ^{238}U activity remain almost the same whatever the depletion status[1]. The mass compositions of DU and of natural uranium are presented below.

	^{238}U	^{235}U	^{234}U
Natural uranium	99.274%	0.72%	0.00554%
The DU used at Kirkcudbright	99.8%	0.20%	0.0008%

[1] So changes in total uranium can be estimated with reasonable accuracy, by measuring changes in the concentration or activity of ^{238}U. (see Annex B)

In natural uranium, the ^{234}U activity is normally in equilibrium with the ^{238}U activity, but as a result of depletion, and the reduction in the concentration of ^{234}U, the $^{238}U/^{234}U$-activity ratio changes. The actual $^{238}U/^{234}U$-activity ratio varies per batch of DU, according to the degree of depletion achieved during processing, but usually lies between 7:1 and 8:1. Consequently, DU is slightly less radioactive than natural uranium.

For the remainder of this report isotopic ratios will be stated as a single value representing the ratio of a number of Becquerels (Bq) of ^{238}U to 1 Becquerel of ^{234}U (i.e. a $^{238}U/^{234}U$ ratio of 7 rather than 7:1).

When converting total uranium concentrations, reported in micro-grams of uranium per gram of sample (µg/g) to activity, two specific activities have been used as conversion factors, depending on the provenance of the uranium. The specific alpha activity of uranium in the DU alloy used at Kirkcudbright is approximately 14 kBq/g, whereas the specific alpha activity of natural uranium is generally 25 kBq/g [2]. The chemical toxicity of DU is about the same as lead.

4 Differentiating DU from natural uranium

A fundamental requirement of the DU environmental monitoring programme is to quantify the impact of DU firing. Historically, this was achieved partly by measuring the amount of total uranium in environmental materials and using this total uranium as an upper bound of DU contamination levels. This is a very cautious approach as natural uranium is present in most natural materials. More sophisticated approaches are now used. These involve specific measurements of the activities or masses of the ^{238}U and ^{234}U isotopes as described below. Although isotope measurements are used in this work, references to total uranium measurements are included for consistency with previous reports. The limitations of using total uranium concentrations are discussed further in Annex B.

As noted above in Section 3, a convenient fingerprint marker for DU is the $^{238}U/^{234}U$ activity ratio. The DU fired at KTA has a $^{238}U/^{234}U$ activity ratio of approximately 7, whereas natural uranium has an activity ratio close to 1. Environmental samples are therefore analysed for isotopes of ^{238}U and ^{234}U to determine activity ratios and hence identify the origin of the uranium.

Substantial deposition of DU, relative to an existing natural uranium inventory, is required before the $^{238}U/^{234}U$ activity ratio in the environment diverges significantly from its natural ratio. (An illustration of the impact of increasing the DU mass in a sample on the sample uranium isotopic ratio is given in Annex C). For the ratio to approach 7 in an analytical sample, the mass of DU would have to be about one hundred times the mass of the uranium that is naturally present. Hence, the lower the natural uranium background, the lower the levels of DU contamination that may be detected by isotopic analysis.

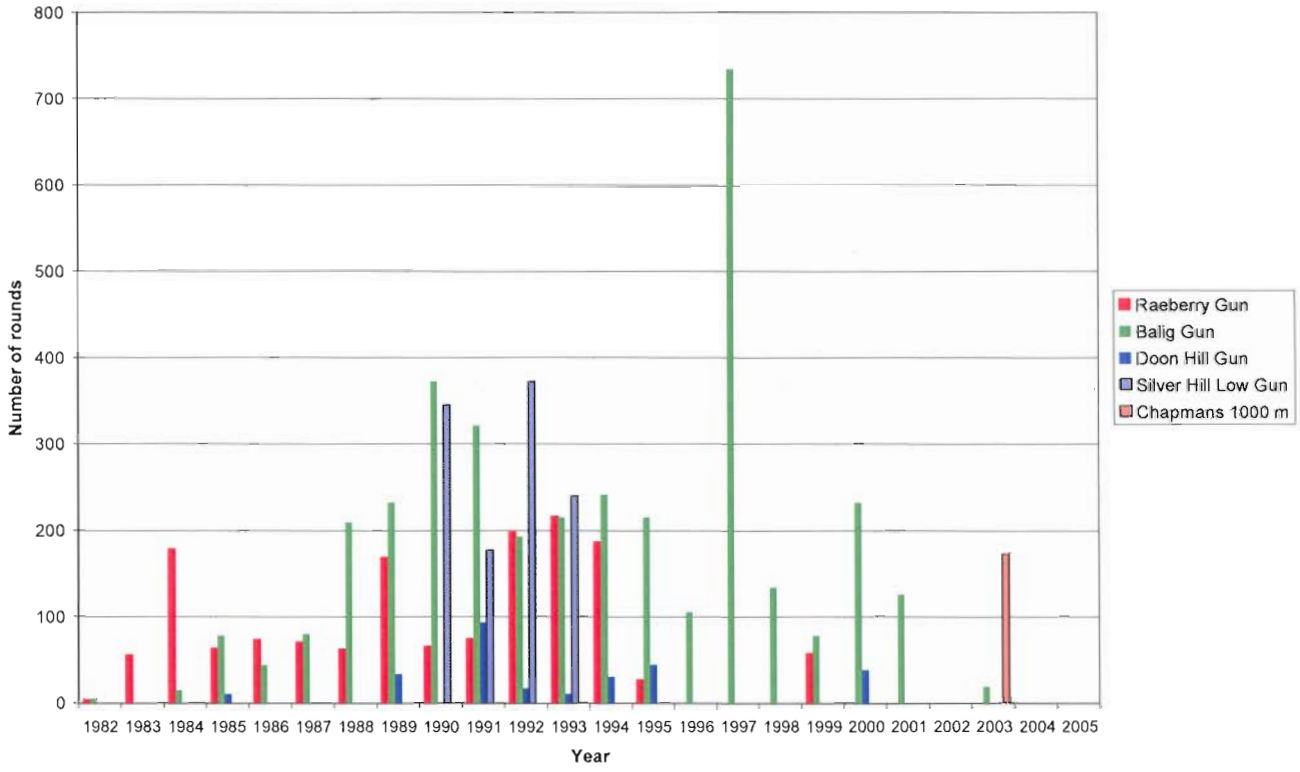

Figure 1: Numbers of DU projectiles fired from KTA, from 1982 to 2005

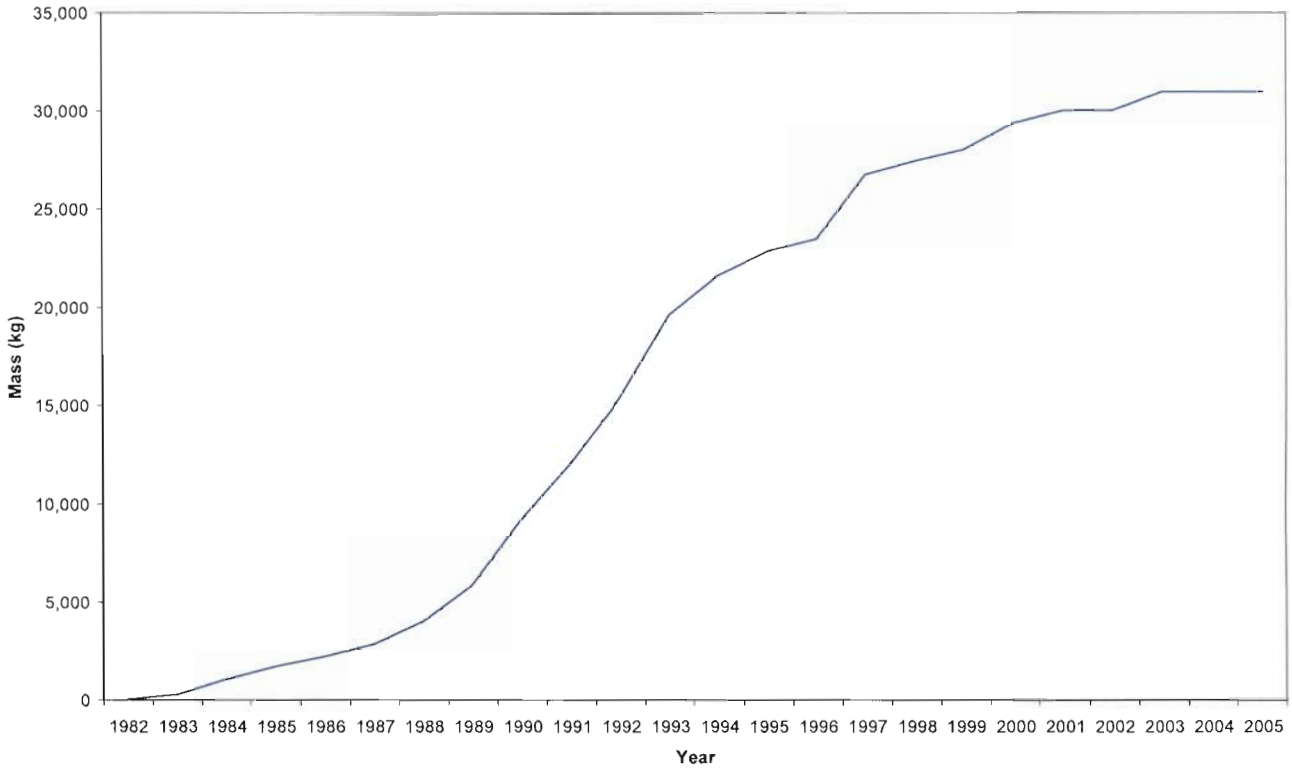

Figure 2: Approximate cumulative mass of DU projectiles fired from KTA, from 1982 to 2005.

Isotopic quantification is achieved by techniques such as Alpha Spectrometry (AS) or Mass Spectrometry (MS). AS can detect uranium to parts per billion, which is equivalent to mBq per kg, or to lower levels if count times are increased. MS is more sensitive than AS, but the lower levels detectable by MS are of no recognised health significance. Furthermore, because most MS measurements are not accredited by the UK Accreditation System (UKAS), the technique is not normally used in routine health and safety or environmental monitoring[2].

Gamma Spectrometry measurements of environmental samples can also yield isotopic information, but the technique is not sensitive enough to reach the limits of detection required for environmental analysis unless relatively large quantities of uranium are present.

5 Investigation/action levels

The MOD's Depleted Uranium Firing Environmental Review Committee (DUFERC) has not set investigation or action levels for DU in marine environmental samples. Instead, for sediments, reference may be made to both the Radioactive Substances Act 1993 (RSA 93) or to Generalised Derived Limit (GDL) for ^{238}U [3, 4]. For substances commonly used as foodstuff, the GDL is the appropriate reference quantity. For material such as seaweed, the reference level used is the European Council Food Intervention Level (CFIL) for other foodstuff of 1,250 mBq/g.[5]

GDL for uranium were last updated in 2000 by the National Radiological Protection Board, which is now part of the Health Protection Agency. GDL relate to an annual dose of 1 mSv for all pathways. This dose was also used as a benchmark by the Royal Society in their studies of the potential health effects of using DU munitions [6]. GDL exist for marine crustaceans and molluscs. GDL for foodstuffs are for the edible fraction and are expressed for fresh mass, whilst GDL for marine sediments are for dry mass.

	Generalised Derived Limit for ^{238}U (mBq/g)
Marine crustaceans	1,000
Marine molluscs	1,000
Marine sediments	100,000

6 Sample collection

6.1 Sampling locations and dates of sampling

In November 2005, samples of intertidal sediment, seaweed and mussels were collected from the 11 sampling locations shown in Table 1 and Figure 3. Scallops and lobsters were purchased locally.

[2] For academic studies, Inductively Coupled Plasma (ICP) MS has become the analytical technique of choice for the measurement of very low uranium content and /or the determination of the isotopic ratio in environmental and biological samples.

It was not possible to collect any underwater sediments samples because of mechanical problems on board the sampling vessel, and their collection was delayed until early January 2006. Similarly, the purchase of crabs was delayed until January 2006, because they are normally caught by, and bought from, the sampling vessel skipper. Whelks were not available for purchase in November 2005, but were in January 2006. Although the sampling/purchasing dates fell in the 2006 calendar year, this work was carried out as part of the 2005 survey, and accordingly, the results are reported here.

Samples of underwater sediments were collected from the sampling locations listed in Table 2 and illustrated in Figure 4.

Table 1: KTA intertidal sample station locations, November 2005.

Sample Station		
Number	Name	Grid reference
1.	South Carse	GR 997 593
2.	Sandyhills Bay	GR 895 551
3.	Port o' Warren Bay	GR 879 532
4.	Port Donnel/Rockliffe	GR 848 534
5	Rascarral Bay	GR 805 480
6.	Abbey Burn Foot	GR 743 444
7.	Mullock Bay	GR 711 437
8.	Lower Nunton Bay	GR 661 485
9.	Brighouse Bay	GR 635 454
10.	Carrick Point	GR 573 507
11.	Mossyard Bay	GR 554 518

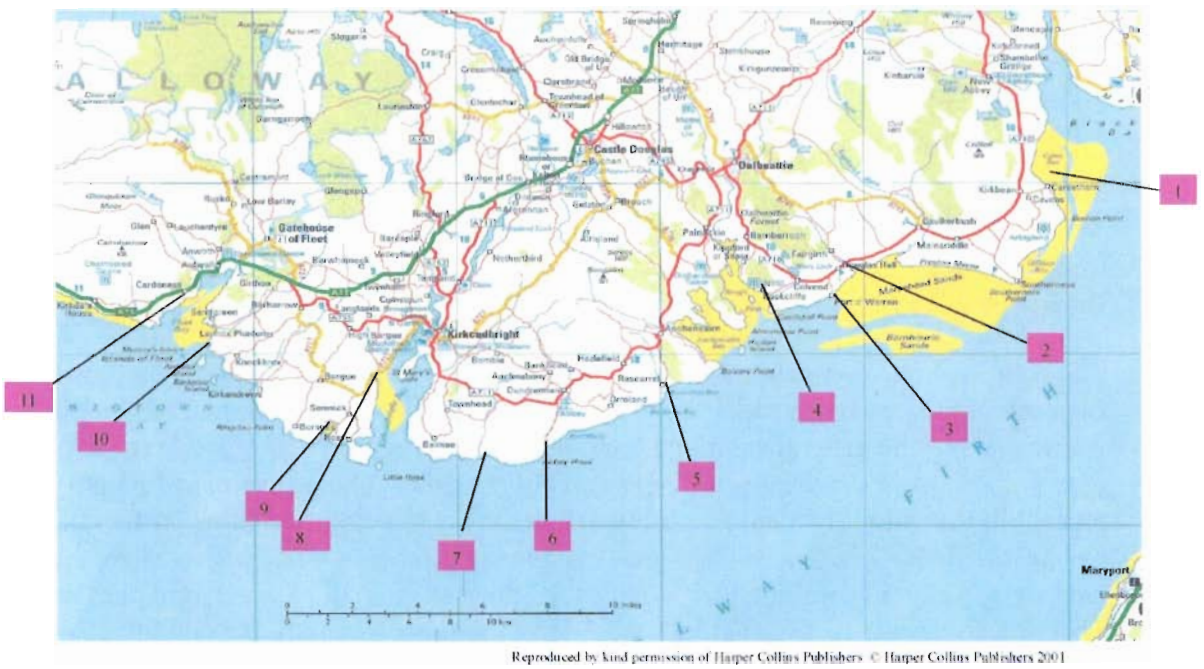

Figure 3: KTA intertidal sampling locations, November 2005.

Table 2: KTA underwater sampling locations, January 2006.

Sample Station Number	January 2006			
	Latitude (North)		Longitude (West)	
1	54°	45.078'	3°	59.782'
2	54°	45.600'	3°	59.808'
3	54°	44.527'	3°	59.746'
4	54°	45.569'	4°	2.577'
5	54°	45.032'	4°	2.541'
6	54°	44.479'	4°	2.522'
7	54°	45.211'	4°	1.174'
8	54°	45.083'	4°	1.243'
9	54°	44.827'	4°	1.391'
10	54°	44.684'	4°	1.477'
11	54°	44.991'	3°	58.807'
12	54°	44.856'	3°	58.758'
13	54°	45.484'	3°	57.969'
14	54°	45.375'	3°	57.744'
15	54°	45.673'	3°	57.125'
16	54°	45.136'	3°	57.108'
17	54°	44.575'	3°	57.081'
18	54°	42.874'	4°	1.311'
19	54°	42.747'	4°	1.347'
20	54°	43.235'	4°	2.493'
21	54°	42.732'	4°	2.476'
22	54°	42.225'	4°	2.451'
23	54°	43.292'	4°	0.243'
24	54°	42.785'	4°	0.226'
25	54°	42.258'	4°	0.192'

6.2 Intertidal sediments sampling methodology

The composition, abundance and availability of sediments vary with time and location. Because of this, each intertidal sediment sample was a composite from a number of points within each of the general sampling areas shown on Figure 3. Sediments were collected as near to the Low Water Mark (LWM) as safely possible. Where several sediment types were present in the sampling area, the finest sediment was selected, and large shells and stones were discarded. At each sampling location, the uppermost layer of sediment was collected using a trowel, and combined with other sub-samples in a 1 litre watertight plastic container. This procedure was repeated at intervals along the shore line until the container was full. Excess water was drained from the container prior to sealing. The overall general type of sediment, as recorded at the time of sampling is indicated in brackets.

The provenance, types and numbers of samples procured from the intertidal sampling locations are listed in Table 3.

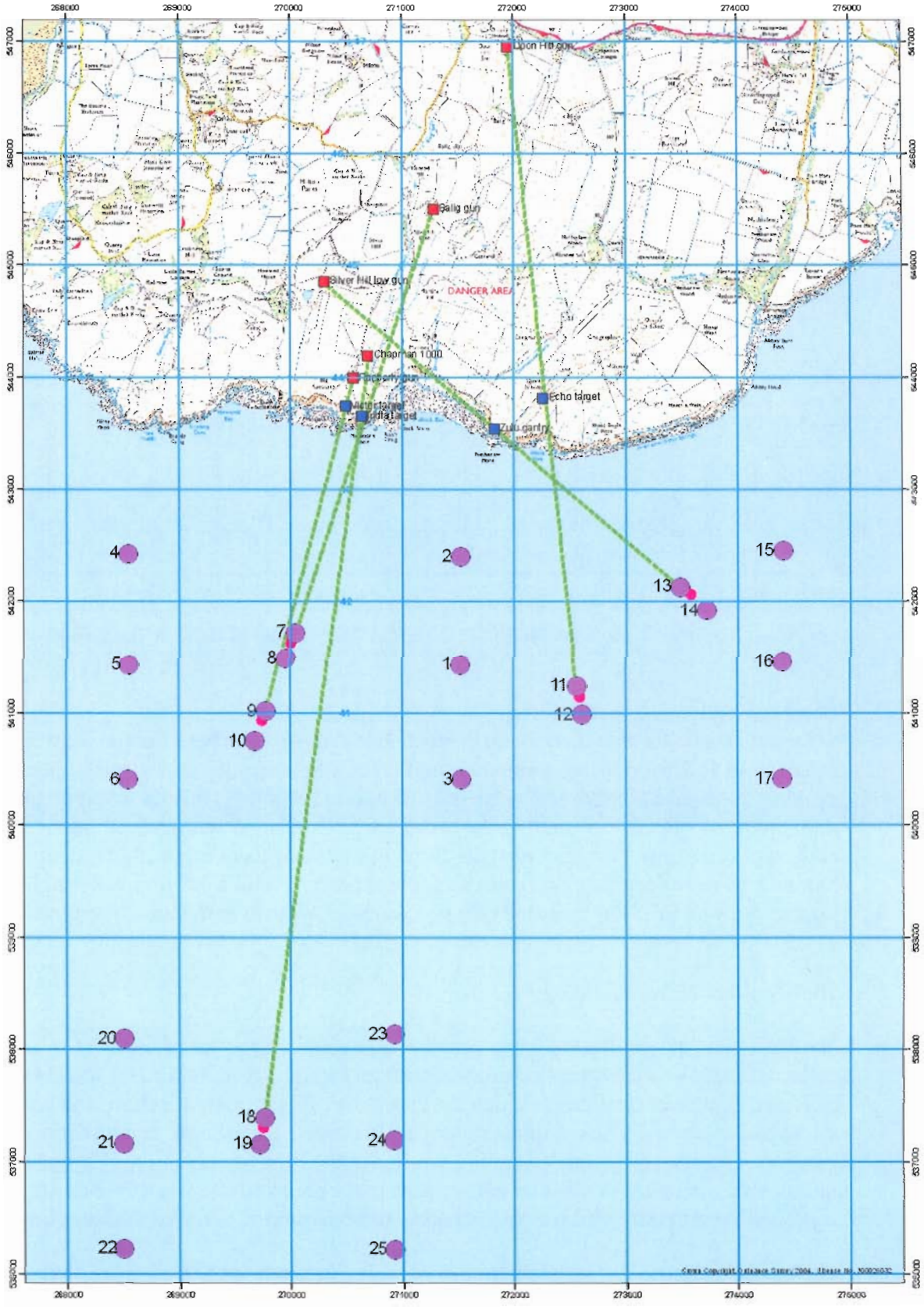

Figure 4: KTA underwater sampling locations, January 2006 (also showing line of fire for each battery).

Sample Area Number	Sample Type, and number of samples				Environmental Dose rates	
	Sediment	Seaweed	Mussel	Other	LWM @ 1m height	HWM @ 1m height
1	1 (sand)	×	1	×	✓	✓
2	1 (sand)	1	×	×	✓	✓
3	1 (sand)	1	1	×	✓	✓
4	1 (sand)	1	×	×	✓	✓
5	1 (gravel)	1	1	×	✓	✓
6	1 (gravel)	1	×	×	✓	✓
7	1 (sand)	1	×	×	✓	✓
8	1 (grit)	1	1	×	✓	✓
9	1 (sand)	1	×	×	✓	✓
10	1 (sand)	1	×	×	✓	✓
11	1 (grit)	1	×	×	✓	✓

Table 3: Summary of intertidal sample collection, November 2005.

Notes: LWM – Low Water Mark HWM – High Water Mark × Sample not collected.

6.3 Seaweed collection methodology

Seaweed was collected when found in sufficient abundance. There was no seaweed to collect at Location 1. The collected samples consisted of single species of seaweed, either *F. serratus* (Location 8 only) or *F. vesiculosus* (other locations). Only the most recent growth of seaweed was collected, by cutting the end 5 cm of fronds. As with the sediments samples, each seaweed sample was a composite from a number of locations within a sampling area. Samples were collected at intervals along the shore line until a 1.8 litre water tight plastic sample pot was full. The seaweed samples were not washed at the time of collection.

6.4 Biota collection methodology

Shellfish were only collected when found in abundance. Samples of mussel (*M. edulis*) were collected from 4 of the sampling areas shown on Figure 3 (see Table 3). Mussels were collected at a number of points along the shoreline of a sampling location, and combined into a composite sample. The composite sample filled two 1 litre sampling pots for each sampling location. The mussel samples were not washed at the time of collection. (Limpet or winkle samples are sometimes collected when mussels are not available. However during November 2005 neither mussels, winkles nor limpets could be found at seven of the sampling areas.)

The mussel samples for each location were boiled in water for two minutes, drained and prised from their shells within two days from collection. The shells were discarded, as were any mussels that remained closed after boiling. The flesh was placed in an individual self sealing plastic bag and frozen prior to transport to the laboratory.

6.5 Dose rate measurement methodology

Gamma radiation measurements were made using three Mini Instruments Type 6-80/81 rate meters fitted with compensated Geiger Muller tubes. The instruments were deployed so that the centres of the Geiger Muller tubes were at one metre above ground, at the High Water Mark (HWM) and the low water mark (LWM[3]) at each sampling location (away from granite boulders and sea walls). Three 100 second counts were taken at the LWM at distance intervals of 1m. The average of the three counts was used to calculate the dose rate. One 100 second count, taken at the HWM, was used to calculate the dose rate at that point.

6.6 Seafood purchase methodology

Four locally caught lobsters (*H.gammarus*) and 0.8 kg of locally fished queen scallops (*C. opercularis*) were purchased in Kirkcudbright, and boiled within a day of purchase. These were then bagged and frozen pending transport to the Dstl laboratory. Whelks (*B. undatum*) and crabs were not available for purchase at the time of the November 2005 visit. However, six crabs and 5 kg of whelks were purchased in January 2006. These were bagged and frozen pending transport to the Dstl laboratory.

6.7 Underwater sediment collection methodology

Underwater sediment samples were collected during January 2006, where there was enough sediment to fill sampling tubs. The sampling locations are shown in Figure 4 and are listed in Table 2. Locations 18, 19, 21-25 were not sampled due to adverse weather and sea conditions. As a result, only 18 samples were collected from 18 locations.

The dynamic nature of the sea makes the sampling of underwater sediment at precise locations difficult. A scooping device fitted with a rubberised bag was lowered in the water at each sampling location. This was dragged along the sea bed over a distance ranging between 50 and 100 m depending on tide and wind conditions. The location coordinates given in Table 2 should therefore be considered to be the approximate centre points of sampling areas of no more than 100 m radius.

7 Sample preparation, laboratory analysis and descriptors

7.1 Preparation and analysis

Sample preparation was either carried out in the field at the time of collection, as described in sections 6.3, 6.4 and 6.6, or at the Dstl Environmental Sciences Department UKAS accredited radiochemistry laboratory, on receipt of the sample. The samples were analysed by AS in the laboratory. An outline of the approach is given below.

One mixed sample of mussel flesh per location was submitted for analysis. Whelks were boiled and the meat was prised from each shell, and the shells were discarded. Cooked lobsters and crabs were opened up and the flesh removed for analysis, but the digestive tracts of the animals were not used. Both the white and the brown meat of crabs were used for

[3] The gamma monitors were placed as near as was safely possible to the LWM

analysis. Each lobster and crab was analysed as a discrete sample, which gave 4 individual lobster samples (each sample consisting of the flesh of one lobster only), 6 individual crab samples (each sample consisting of the flesh of one crab only). The scallop meat was divided into 3 scallop replicate samples by dividing the 0.8 kg of scallop meat into three samples of approximately similar mass. The whelk meat was divided, in a similar fashion, into 6 replicate samples of similar mass.

The sediments, seaweed and seafood samples thus obtained were weighed before drying to remove moisture, and then weighed again (with results being reported as dry weight). The samples were ashed to remove organic material, and homogenised by hand. During this process the quantity of uranium in each of the samples remains constant. The ashed samples were boiled in concentrated mineral acid (nitric acid and hydrochloric acid) to remove the 'loose' and leachable uranium from the sample. Recalcitrant matrices such as mineral grains are not broken down by the process and hence natural uranium bound up within them was not extracted. The samples were filtered to remove solids.

Uranium separation was carried out by extraction chromatography. Each eluted sample was electro-deposited onto a stainless steel planchette and the activity of each planchette was counted in a low background, silicon surface barrier, alpha spectrometer.

DU concentrations are reported in mBq/g (equivalent to Bq/kg) of dry weight.

7.2 Sample descriptor code

Each sample was given a unique sample descriptor. For intertidal and underwater sediments, seaweed and mussels the descriptor comprised of a descriptor of sample type, followed by a location descriptor.

The sample type descriptors are as follows:

- I/T sed - intertidal sediment;
- U/W sed - underwater sediment;
- Seaweed - *F. vesic/ F. ceranodies/ F serratus*;
- Mussels –*M. edulis*; and
- scallops, lobsters, crabs, or whelks.

In the case of purchased sea food no location descriptor has been included because the precise location were the sample was collected is unknown. For replicate samples of scallop and whelk, and for individual crabs and lobsters, A to F were added.

Five examples of the sample descriptors are given below:

I/T sed 3 – for intertidal sediments collected from intertidal location 3.

U/W sed 8 –underwater sediment collected at underwater location 8, (54°45.083'N and 4°1.243'W).

Mussels 10 – Mussel sample from intertidal sampling location 10, Carrick Point.

Scallops C – replicate sample from the purchased (0.8 kg) composite sample of scallop.

Crab E – fifth crab of six individual crabs purchased.

8 Results and interpretation

A summary of the results for all marine samples collected in 2005 and early 2006 is given in Table 4. The detailed marine monitoring results are given in Table 5 to Table 9.

Historical monitoring results for the KTA marine environment, presented in Table 10 at Annex A, provide an overview of the routine monitoring results through time, from 1996 to 2005.[7, 8, 9, 10, 11, 12, 13, 14]

To identify whether the DU firing activities at KTA have had a detectable impact on uranium concentrations in the local marine environment, it is possible to compare the results for the routine monitoring programme with uranium activities and ratios reported for similar sample types across the UK. Two studies have reported ambient dose rates and uranium activity in marine sediment, seaweed and biota samples from a range of coastal locations in the UK [15, 17]. One of the studies [17] included samples from a location in the South-West of Scotland, named Sandyhills, which is sufficiently near to KTA to be geochemically similar and sufficiently far to be considered unaffected by activities at the range (See Annex D).

Sample Type	Number of Samples	No. of samples containing detectable DU	Total uranium concentration (mBq/g)			
			Mean	Standard deviation of the mean	Minimum	Maximum
Shore Sediment	11	0	16.6	4.1	9.8	24.5
Seaweed	10	0	9.0	4.1	4.9	18.2
Mussels	4	0	8.2	2.3	5.1	10.7
Scallops (Queens)	3	0	2.4	0.9	1.6	3.3
Lobsters	4	0	0.7	0.2	0.4	0.9
Whelks	6	0	1.9	0.9	0.7	2.9
Crabs	6	0	0.8	0.5	0.4	1.6
Underwater Sediment	18	0	19.5	2.3	16.3	24.3

Table 4: Summary of results, all marine samples 2005 (values have been rounded to 1 decimal place for clarity).

8.1 Dose rates

Dose rate measurement results for 2005 are given in Table 5. The maximum recorded dose rate was 130 nGy/h. The results compare with four dose rate measurements averaging 72 nGy/h made by the Centre for Environment, Fisheries and Aquaculture Science (CEFAS) in 2005 over a salt marsh at Kirkcudbright [15] and with measurements made as part of the routine monitoring programme in previous years, and reflect background dose rates.

Table 5: Intertidal dose rate measurements, November 2005.

Station Number	Location	Average Dose rate (nGy/h)
1	South Carse	70
2	Sandyhills Bay	70
3	Port o' Warren Bay	90
4	Port Donnel	95
5	Rascarral Bay	130
6	Abbey Burn Foot	65
7	Mullock Bay	105
8	Lower Nunton Bay	85
9	Brighouse Bay	80
10	Carrick Point	75
11	Mossyard Bay	75

8.2 Intertidal sediments results

No sample was radioactive within the meaning of the RSA 93, or exceeded 0.03% of the GDL for uranium in sediments [3, 4].

As mentioned above, total uranium concentrations have also been used historically to provide an upper bound on DU contamination levels. Previous monitoring reports have compared sample results with the range of total uranium activity concentrations in UK coastal sediments[4] [13,16]. In 2005, the intertidal sediment samples all contained total U at activity concentrations (9.8 to 24.5 mBq/g of dry weight) that were within this range, (see Table 6). The range of total uranium activity concentrations for the samples was also well within the range of results reported for intertidal sediment collected since 1996 (see Annex A).

The uranium activity concentration for the sample collected from Rascarrel Bay was twice as high as the results reported by CEFAS. However, the difference is small in absolute terms and is of no health significance. Variations of this type are expected in the shore environment (see Annex B).

For sediment obtained from Sandyhills [17] in Southwest Scotland (see Annex D), only the ^{238}U activity concentration was reported (14 ± 0.4 mBq/g dry weight), as is common in environmental reports (see Annex B). It can be seen from Table 6, that all intertidal sediment samples collected at KTA in 2005 had a ^{238}U activity concentration below that value.

Within the range of the uncertainty for the analysis, the isotopic ratios for all but one of the intertidal sediment samples collected are in agreement with values reported in marine sediments in general (see Annex C). One sample (I/Tsed 2) has a uranium isotopic ratio

[4] Minerals containing uranium are widely distributed on the surface of the earth's crust and the concentration of natural uranium in the environment therefore varies between locations. Consequently there is no single, definitive, reference level for natural uranium in soils or the sediments that originate from these soils, but there is broad agreement in the range of values published in literature. Typical values in the UK range from 3.6 to 32.3 mBq/g. of dry weight but values of up to 100 times the typical range can be found locally.

(1.3 ± 0.4) greater than the typical isotopic ratio in marine sediments of 0.8. However, the increase in the isotopic ratio is not accompanied by a corresponding rise in the total uranium activity that would be expected for a natural medium contaminated by DU (see Annex E). The slightly elevated isotopic ratio may therefore be an artefact of counting statistics. The finding is also radiologically insignificant, as the total uranium activity concentration (17.6 mBq/g) was very much less than the GDL for sediments of 100,000 mBq/g.

8.3 Underwater sediments results

No sample was radioactive within the meaning of the RSA 93 or exceeded 0.03% of the GDL for uranium in sediments [3, 4].

The ^{238}U activity concentrations reported (in Table 7) for all 18 underwater sediments samples collected (8.1 to 12.0 mBq/g of dry weight) were all within the UK coastal sediments range of 3.6 to 32.3 mBq/g of dry weight (see Annex D). The range of total uranium activity concentrations agreed with ranges of values reported in previous years. This correlation is to be expected in such a well mixed environment.

The $^{238}U/^{234}U$ isotopic ratios reported for 18 underwater sediments samples were, within the range of the uncertainty for the analysis, in agreement with values generally reported in marine sediments (see Annex C). Three samples (U/W sed 3, 9, 17) have a uranium isotopic ratio (1.2 ± 0.3, 1.2 ± 0.3, 1.3 ± 0.3) greater than the typical isotopic ratio in marine sediments of 0.8. However, as above, the increases in the isotopic ratio are not accompanied by corresponding rises in the total uranium activity concentration that would be expected for a natural medium contaminated by DU (see Annex E). The slightly elevated isotopic ratios may therefore also be an artefact of counting statistics. The finding is also radiologically insignificant as the total uranium activity concentrations (18.4, 19.3 and 21.4 mBq/g) were very much less than the GDL for sediments of 100,000 mBq/g.

8.4 Intertidal seaweed results

No sample exceeded 1.6 % of the CFIL for uranium in other foodstuff [5].

The total uranium activity concentrations and the ^{238}U activity concentrations for the seaweed samples obtained from the 10 locations sampled were low (4.9 to 18.2 mBq/g of dry weight for total uranium, 2.2 to 8.2 mBq/g of dry weight for ^{238}U – See Table 8), and in agreement with the range of ^{238}U activity concentrations reported for seaweed sampled in the UK as a whole (3.8 to 18.6 mBq/g). Four samples had ^{238}U activity concentrations that were slightly higher than values reported for seaweed obtained from Sandyhills, but this may be due to the fact that the seaweed obtained during the KTA marine environmental survey was not washed to remove sediment that may have adhered to the weed (See Section 6.3). The range of total uranium results is in agreement with values reported in previous years (See Annex A).

There are no reported values in the literature of the $^{238}U/^{234}U$ isotopic ratios in seaweed. Values reported here are in agreement with those reported for the same locations from previous years, and do not indicate the presence of depleted uranium.

8.5 Biota results

No sample exceeded 1.1 % of the GDL for uranium in molluscs and crustaceans. [4]

The four mussel samples collected during November 2005 yielded total uranium activity concentrations from 5.1 to 10.7 mBq/g of dry weight, and ^{238}U activity concentrations of 2.4 to 4.9 mBq/g of dry weight (See Table 9). These values are three orders of magnitude below levels considered to be of significance to health (see Section 5). The ^{238}U activity concentrations are slightly above the values reported for mussels sampled at Sandyhills, but were at the lower end of the range of values reported for mussels sampled in the UK as a whole. However, the variation is not of any practical significance, at such low activity concentrations. In addition, higher activity concentrations may be due to sediment adhering the shells which were not washed prior to the cooking step (See section 6.4). Sediment from the shells may have therefore contaminated the mussel flesh that was supplied for analysis. The mean total uranium activity concentration is in agreement with mean values reported in previous years (see Annex A).

No ^{238}U could be detected in any of the lobster, crab and scallop, or in three of the six whelk replicate samples that were analysed. ^{238}U was detected in the other whelk samples but the $^{238}U/^{234}U$ ratio was indicative of natural uranium.

CEFAS have reported activity concentrations for crabs and lobsters in the UK[15]. (The total concentrations are reported per gram of wet weight). In contrast, the KTA 2005 survey results (for ^{238}U) reported here are per gram of dry weight. Since the drying step of the lobster and crab sample preparation removed an average moisture equal to 72 and 78 % of the sample wet weight (for lobster and crab samples respectively), the corresponding range of wet weight uranium activity concentrations for the lobster samples obtained in the survey would be 0.1 – 0.2 mBq/g, and for crabs would be 0.1 – 0.5 mBq/g. (Sample preparation is discussed in Section 6.3 and at Annex B). These values are higher than the mean ^{238}U activity concentrations for crabs and lobsters obtained from the UK as a whole, but are very much less than the GDL of 1,000 mBq/g for uranium activity concentrations in molluscs and crustaceans.

The total uranium activity concentrations for the duplicate scallop and whelk samples are in agreement with each other, and the ^{238}U activity concentrations detected in the whelk samples compare with ^{238}U activity concentrations for other molluscs from Sandyhills and from the UK at large. The mean total uranium activity concentration is in agreement with mean values reported in previous years (see Annex A).

8.6 Results summary

These findings are in agreement with those from previous years and do not indicate any impact from the DU firings.

9 Evaluation of potential exposure pathways.

The contamination of the marine environment with DU would result in four potential exposure pathways for humans. These are:

- external radiation exposure from contaminated seaweed or sediments;

- inhalation of DU contamination released into the air or re-suspended from sediments;
- ingestion of seaweed or food products contaminated with DU; and
- ingestion of DU contaminated sea water. Although ingestion of DU contaminated sea water is a theoretical possibility, seawater sampling was discontinued at the recommendation of SEPA in 2001. The technical assessment was that the immense dilution present in the Solway would never give rise to uranium levels detectable above background, or to any significant radiation dose from DU.

9.1 External radiation

Radiation levels on the shore are indistinguishable from natural background levels and hazards due to external radiation exposure from DU are therefore considered to be insignificant.

9.2 Inhalation of re-suspended DU

DU that has been deposited in sediment or in sediment attached to objects and plants such as fishermen's netting and seaweed may be re-suspended into the air and subsequently inhaled. Results reported here show that measured total uranium activity concentrations do not exceed background values for UK sediments and that there is no evidence of depletion in the isotopic ratios. Consequently, it is concluded that there are no risks from inhalation of DU to any member of the public.

9.3 Ingestion of DU Contaminated foodstuffs

Total uranium concentrations in the marine foodstuffs were in agreement with background levels for such sample types. There was no evidence of depletion in the isotopic ratio. It is concluded that there are no risks from ingestion of DU to any member of the public.

10 Conclusions

It is concluded that:

- the survey results show no evidence of DU being present in any marine environmental sample collected in the year 2005 and in January 2006; and,
- there is no evidence to suggest that members of the public are or have been subjected to any radiation hazard from the marine environment due to the firing of DU ammunition at Kirkcudbright.
- the 2004 survey report recommended that historical data should be subjected to a collective review. This has not been done, because protocols for the yearly environmental surveys and data obtained therein are already subjected to reviews in the quest for best practice, as part of the normal dynamic Dstl QA system.

11 Recommendations

It is recommended that seaweed and biota samples be rinsed at the point of sampling. This would minimise the amount of sediment that could be included with the biological samples and thus limit the potential for such sediment to contribute to uranium concentrations and would align the intertidal sampling protocol with that of other monitoring bodies. (See Annexes C and D.)

12 List of reference

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KTA 2005 survey results (Pages 26 to 29)

Sample Descriptor	Wet weight (g)	Dry weight (g)	Ashed weight (g)	Measured Activity of Dry Sample (mBq/g)														Ratio of $^{238}U/^{234}U$		
				^{238}U			^{235}U			^{234}U			Total U							
I/T sed 1	581.3	336.6	315.6	7.6	±	1.6		<	0.6	7.2	±	1.5	14.9	±	2.2	1.0	±	0.3
I/T sed 2	608.6	472.8	462.3	9.6	±	1.7	0.9	±	0.4	7.2	±	1.4	17.6	±	2.2	1.3	±	0.4
I/T sed 3	527.2	416.7	407.3	6.2	±	1.2		<	0.4	8.4	±	1.5	15.0	±	2.0	0.7	±	0.2
I/T sed 4	583.9	458.9	448.4	10.4	±	1.9		<	0.6	11.8	±	2.1	22.3	±	2.8	0.9	±	0.2
I/T sed 5	579.9	489.1	484.6	11.6	±	2.0	0.9	±	0.4	12.0	±	2.1	24.5	±	2.9	1.0	±	0.2
I/T sed 6	579.6	457.7	446.0	6.9	±	1.3		<	0.4	8.6	±	1.5	15.8	±	2.0	0.8	±	0.2
I/T sed 7	549.7	460.2	425.8	9.3	±	1.6		<	0.4	9.7	±	1.6	19.3	±	2.3	1.0	±	0.2
I/T sed 8	577.3	460.9	450.1	7.1	±	1.4		<	0.5	6.7	±	1.3	14.0	±	1.9	1.1	±	0.3
I/T sed 9	556.8	453.1	440.8	6.9	±	1.4		<	0.5	7.3	±	1.4	14.6	±	2.0	0.9	±	0.3
I/T sed 10	588.1	458.7	448.3	5.1	±	1.1		<	0.5	4.7	±	1.0	9.8	±	1.5	1.1	±	0.3
I/T sed 11	649.9	513.8	506.9	6.6	±	1.3		<	0.5	8.4	±	1.6	15.1	±	2.1	0.8	±	0.2

Table 6: Intertidal (Shoreline) sediment samples, total uranium and isotopic ratio results – November 2005.

Note: Activity results have been rounded to 1 decimal place. A specific activity for DU of 14.0 MBq/kg has been used. All uncertainties are stated at a 95% confidence level. Limits of Detection (LOD) are calculated by a 'modified Currie' formula[5] at 95%. The total activity is calculated from the sum of the actual activities for each isotope, regardless of the LOD quoted for that isotope, hence in Table 6, where activities are reported as less than LOD for any of the uranium isotopes, the total uranium value may not be equal to the sum of the individual isotopic values.

[5] Revisiting Currie - how low can you go? by Hurtgen C, Jerome S, Woods M. Applied Radiation and Isotopes 53 pp 45-50 (2000)

Sample descriptor	Wet weight (g)	Dry weight (g)	Ashed weight (g)	Measured Activity of Dry Sample (mBq/g)														Ratio of $^{238}U/^{234}U$		
				^{238}U			^{235}U			^{234}U			Total U							
U/W sed 1	649.5	459.8	437.9	10.1	±	1.7		<	0.4	9.0	±	1.6	19.3	±	2.3	1.1	±	0.3
U/W sed 2	588.3	392.9	370.4	9.5	±	1.7		<	0.5	9.2	±	1.6	19.1	±	2.3	1.1	±	0.3
U/W sed 3	563.2	414.5	396.2	8.1	±	1.5		<	0.5	9.9	±	1.7	18.4	±	2.3	1.2	±	0.3
U/W sed 4	534.3	353.3	334.3	9.8	±	1.7		<	0.4	10.0	±	1.7	20.2	±	2.4	0.9	±	0.2
U/W sed 5	524.8	399.4	383.6	9.6	±	1.6		<	0.4	9.6	±	1.6	19.5	±	2.3	1.1	±	0.3
U/W sed 6	566.7	404.0	386.4	9.0	±	1.5		<	0.4	10.2	±	1.7	19.6	±	2.3	1.0	±	0.3
U/W sed 7	561.3	404.5	389.0	9.6	±	1.7		<	0.5	8.7	±	1.6	18.8	±	2.4	1.1	±	0.3
U/W sed 8	564.8	416.7	401.2	9.0	±	1.5		<	0.4	8.5	±	1.5	17.8	±	2.1	1.1	±	0.3
U/W sed 9	563.3	410.7	396.7	10.3	±	1.8		<	0.5	8.7	±	1.6	19.3	±	2.4	1.2	±	0.3
U/W sed 10	533.7	388.3	373.5	9.5	±	1.7		<	0.5	10.6	±	1.8	20.4	±	2.5	0.9	±	0.2
U/W sed 11	543.8	401.3	386.0	8.4	±	1.5		<	0.5	7.6	±	1.4	16.3	±	2.1	1.1	±	0.3
U/W sed 12	495.1	371.3	358.7	10.9	±	2.1	2.1	±	0.7	11.3	±	2.1	24.3	±	3.0	1.0	±	0.3
U/W sed 13	535.5	379.5	360.9	11.5	±	2.0		<	0.5	11.5	±	2.0	23.2	±	2.8	1.0	±	0.2
U/W sed 14	564.5	410.8	393.5	8.9	±	1.6		<	0.5	8.2	±	1.5	17.4	±	2.2	1.1	±	0.3
U/W sed 15	486.9	362.9	350.3	8.3	±	1.5		<	0.4	8.7	±	1.5	17.2	±	2.2	1.0	±	0.2
U/W sed 16	511.7	384.1	370.3	8.3	±	1.4		<	0.4	7.9	±	1.4	16.4	±	2.0	1.1	±	0.3
U/W sed 17	488.8	370.4	323.1	11.8	±	1.9		<	0.4	9.3	±	1.6	21.4	±	2.5	1.3	±	0.3
U/W sed 20	476.1	339.3	323.1	12.0	±	2.0		<	0.5	10.9	±	1.9	23.2	±	2.7	1.1	±	0.3

Table 7: Underwater sediment samples, total uranium and isotopic ratio results – January 2006.

Note: Activity results have been rounded to 1 decimal place. A specific activity for DU of 14.0 MBq/kg has been used. All uncertainties are stated at a 95% confidence level. Limits of Detection (LOD) are calculated by a 'modified Currie' formula[6] at 95%. The total activity is calculated from the sum of the actual activities for each isotope, regardless of the LOD quoted for that isotope, hence in Table 7, where activities are reported as less than LOD for any of the uranium isotopes, the total uranium value may not be equal to the sum of the individual isotopic values.

[6] Revisiting Currie - how low can you go? by Hurtgen C, Jerome S, Woods M. Applied Radiation and Isotopes 53 pp 45-50 (2000)

Sample Description	Wet weight (g)	Dry weight (g)	Ashed weight (g)	Measured Activity of Dry Sample (mBq/g)												Ratio of $^{234}U/^{238}U$		
				^{238}U			^{235}U			^{234}U			Total U					
Seaweed 2	303.8	65.2	16.6	3.5	±	0.6		<	0.2	4.0	±	0.7	7.6	±	0.9	0.9	±	0.2
Seaweed 3	310.0	60.0	15.0	3.4	±	0.6	0.2	±	0.1	4.0	±	0.7	7.6	±	0.9	0.9	±	0.2
Seaweed 4	225.3	51.4	9.3	2.7	±	0.4	0.2	±	0.1	3.2	±	0.5	6.1	±	0.7	0.8	±	0.2
Seaweed 5	275.3	58.5	8.4	6.0	±	0.8	0.3	±	0.1	6.7	±	0.9	12.9	±	1.3	0.9	±	0.2
Seaweed 6	313.0	84.8	17.0	4.6	±	0.7	0.2	±	0.1	5.7	±	0.8	10.5	±	1.1	0.8	±	0.2
Seaweed 7	212.6	46.9	8.4	8.2	±	1.1	0.3	±	0.1	9.7	±	1.3	18.2	±	1.7	0.8	±	0.2
Seaweed 8	300.5	48.3	8.1	5.5	±	0.8	0.3	±	0.1	4.4	±	0.6	10.2	±	1.0	1.2	±	0.2
Seaweed 9	220.9	35.4	8.3	2.2	±	0.4	0.1	±	0.1	2.8	±	0.5	5.0	±	0.6	0.8	±	0.2
Seaweed 10	267.4	55.8	12.4	3.4	±	0.6	0.2	±	0.1	3.9	±	0.6	7.4	±	0.8	0.9	±	0.2
Seaweed 11	258.4	80.2	8.9	2.4	±	0.4	0.1	±	0.0	2.5	±	0.4	4.9	±	0.5	1.0	±	0.2

Table 8: Seaweed samples, total uranium and isotopic ratio results – November 2005.

Note: Activity results have been rounded to 1 decimal place. A specific activity for DU of 14.0 MBq/kg has been used. All uncertainties are stated at a 95% confidence level. Limits of Detection (LOD) are calculated by a 'modified Currie' formula[7] at 95%. The total activity is calculated from the sum of the actual activities for each isotope, regardless of the LOD quoted for that isotope, hence in Table 8, where activities are reported as less than LOD for any of the uranium isotopes, the total uranium value may not be equal to the sum of the individual isotopic values.

[7] Revisiting Currie - how low can you go? by Hurtgen C, Jerome S, Woods M. Applied Radiation and Isotopes 53 pp 45-50 (2000)

Sample Descriptor	Wet weight (g)	Dry weight (g)	Ashed weight (g)	Measured Activity of Dry Sample (mBq/g)														Ratio of ^{238}U / ^{234}U		
				^{238}U			^{235}U			^{234}U			Total U							
Mussels 1	104.4	26.5	4.7	4.9	±	0.7	0.2	±	0.1	5.6	±	0.8	10.7	±	1.1	0.9	±	0.2
Mussels 3	118.2	32.0	5.1	4.3	±	0.7	0.2	±	0.1	4.2	±	0.7	8.7	±	0.9	1.0	±	0.2
Mussels 5	88.9	19.1	2.0	4.1	±	0.7	0.2	±	0.1	4.1	±	0.7	8.3	±	0.9	1.0	±	0.2
Mussels 8	165.1	38.5	4.0	2.4	±	0.4		<	0.1	2.7	±	0.5	5.1	±	0.6	0.9	±	0.2
Scallops A	273.0	62.1	N/A		<	3.7		<	2.2	2.5	±	1.5	3.3	±	1.7		N/A	
Scallops B	294.6	67.7	N/A		<	1.6		<	0.9		<	0.9	1.6	±	0.7		N/A	
Scallops C	207.3	47.7	N/A		<	2.2		<	1.3	1.5	±	0.8	2.1	±	1.0		N/A	
Lobster A	99.8	22.5	N/A		<	2.3		<	1.4		<	1.4	0.7	±	0.6		N/A	
Lobster B	53.0	11.3	N/A		<	1.4		<	0.8		<	0.8	0.4	±	0.4		N/A	
Lobster C	59.1	11.6	N/A		<	2.9		<	1.7		<	1.7	0.9	±	0.7		N/A	
Lobster D	85.3	19.3	N/A		<	2.3		<	1.4		<	1.4	0.7	±	0.6		N/A	
Welks A	371.9	91.8	N/A		<	2.0		<	1.6		±	1.6	0.9	±	0.7		N/A	
Welks B	409.3	103.0	N/A		<	1.7		<	1.4		<	1.4	0.7	±	0.6		N/A	
Welks C	434.1	110.6	N/A		<	1.6		<	1.2		±	1.2	1.7	±	0.9		N/A	
Welks D	424.5	103.0	N/A	1.1	±	0.6		<	0.8	1.0	±	0.5	2.3	±	0.8	1.1	±	0.8
Welks E	406.4	97.6	N/A	1.1	±	0.6		<	0.8	1.4	±	0.6	2.6	±	0.9	0.8	±	0.5
Welks F	462.7	115.3	N/A	1.7	±	0.7		<	0.8	1.1	±	0.6	2.9	±	1.0	1.6	±	1.1
Crab A	81.5	23.3	N/A		<	1.3		<	1.0		<	1.0	0.5	±	0.4		N/A	
Crab B	115.5	34.0	N/A		<	1.1		<	0.9		<	0.9	0.5	±	0.4		N/A	
Crab C	67.7	19.7	N/A		<	0.9		<	0.7		<	0.7	0.4	±	0.3		N/A	
Crab D	81.2	22.0	N/A		<	2.0		<	1.6		<	1.6	0.8	±	0.7		N/A	
Crab E	73.7	19.9	N/A		<	1.6		<	1.2		<	1.2	0.7	±	0.5		N/A	
Crab F	82.8	23.4	N/A		<	1.0		<	0.8	1.0	±	0.6	1.6	±	0.7		N/A	

Table 9: Marine biota samples, total uranium and isotopic ratio results – November 2005 (scallops, mussels, lobsters) and January 2006 (crab and whelks)

Note: Activity results have been rounded to 1 decimal place. A specific activity for DU of 14.0 MBq/kg has been used. All uncertainties are stated at a 95% confidence level. Limits of Detection (LOD) are calculated by a 'modified Currie' formula[8] at 95%. The total activity is calculated from the sum of the actual activities for each isotope, regardless of the LOD quoted for that isotope, hence in Table 9, where activities are reported as less than LOD for any of the uranium isotopes, the total uranium value may not be equal to the sum of the individual isotopic values.

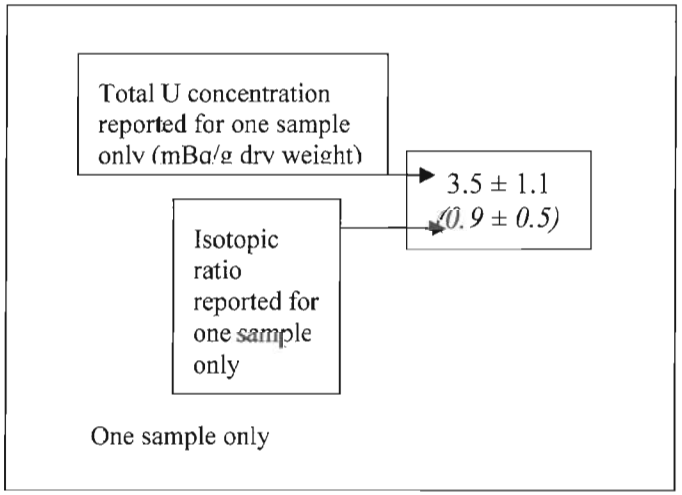
[8] Revisiting Currie - how low can you go? by Hurtgen C, Jerome S, Woods M. Applied Radiation and Isotopes 53 pp 45-50 (2000)

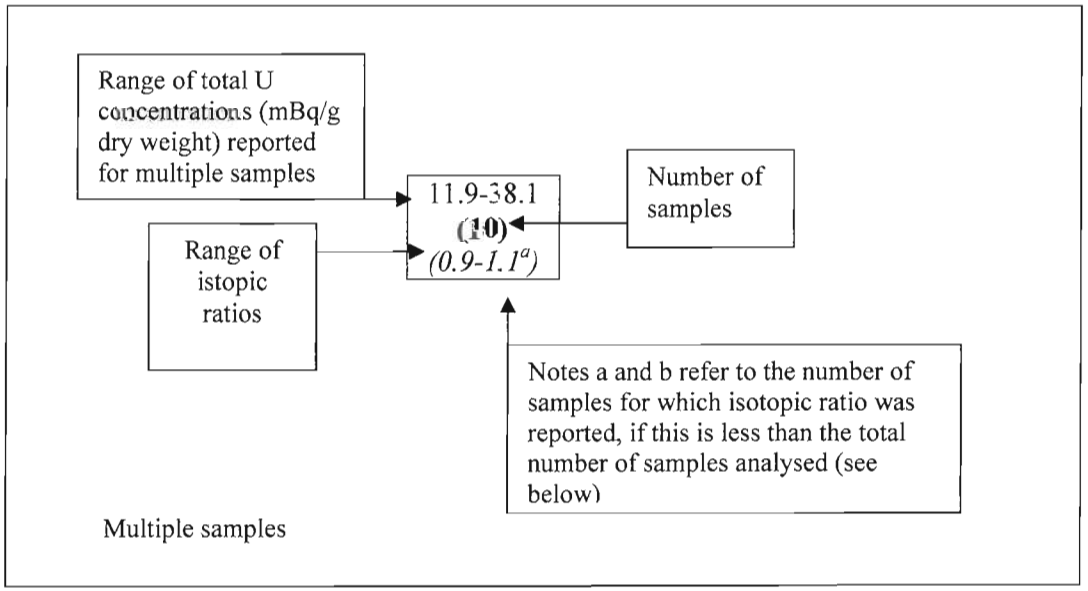
ANNEX A Historical data

Table 10: Summary of marine results reported for 1996-2005.

Sample Type	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Shore Sediment	11.9-38.1 **(10)** *(0.9-1.1)*	14.3-19.0 **(10)** *(0.8-1.2)*	12.2-31.8 **(10)** *(0.6-1.2)*	11.0-32.6 **(10)** *(0.9-1.1)*	2.2- 44.7 **(10)** *(0.8-1.2)*	10.7-31.1 **(9)** *(0.9-1.4)*	12.0-31.9 **(10)** *(0.9-1.2)*	9.8-714.4 **(10)** *(0.8-1.3)*	4.2-35.0 **(11)** *(0.9-1.3)*	9.8-24.5 **(11)** *(0.8-1.3)*
Seaweed	8.7-26.7 **(9)** *(0.7-0.9)*	9.8-22.7 **(9)** *(0.8-1.1)*	5.0-19.6 **(9)** *(0.8-1.2)*	7.8-14.1 **(9)** *(0.8-1.3)*	1.3 -5.1 **(9)** *(0.9 -1.1)*	6.8-12.1 **(7)** *(0.8-1.3)*	0.2-16.7 **(9)** *(0.7-1.2)*	7.2-34.2 **(9)** *(0.7-0.9)*	9.2-22.6 **(9)** *(0.8-1.0)*	4.9-18.2 **(10)** *(0.8-1.2)*
Mussels	6.8-14.9 **(8)** *(0.8-1.1)*	4.3-11.4 **(8)** *(0.8-1.1)*	7.6-12.2 **(5)** *(0.8-1.1)*	6.6-11.1 **(5)** *(0.8-1.1)*	3.6 -7.2 **(5)** *(0.8-0.9)*	2.9-11.6 **(7)** *(0.3-0.9)*	6.4-10.5 **(5)** *(0.9-1.0)*	5.6-7.4 **(5)** *(0.7-0.9)*	4.0-7.8 **(4)** *(0.8-0.9)*	5.1-10.7 **(4)** *(0.9-1.0)*
Whelks	×	×	2.5-3.0 **(3)** *(0.8-1.0)*	0.2-0.4 **(4)** *(0.3-1.3)*	0.6-1.3 **(5)** *(0.5-1.6)*	2.0-3.7 **(8)** *(0.5-1.4)*	1.1-39.4 **(8)** *(0.9 ± 0.5[a])*	×	×	0.7-2.9 **(6)** *(0.8-1.6[b])*
Scallops	×	2.8-3.5 **(4)** *(0.8-1.0)*	2.8-3.2 **(3)** *(0.8-1.1)*	0.7-1.1 **(4)** *(1.3-1.5)*		2.8-13.3 **(8)** *(0.7-1.6)*	×	1.5-33.0 **(6)** *(n/a)*	2.8-3.1 **(2)** *(0.9-9.9)*	2.8-13.3 **(3)** *(n/a)*
Crabs	×	0.6 ± 0.3 **(1)** *(0.9 ± 0.5)*	×	0.3-0.4 **(2)** *(1.5-1.7)*	0.2-1.4 **(8)** *(0.1-3.7)*	1.2-2.3 **(2)** *(0.9-1.0)*	0.8-23.8 **(4)** *(0.9 ± 0.5[a])*	0.1-2.3 **(3)** *(n/a)*	×	0.4-1.6 **(6)** *(n/a)*
Lobsters	×	×	0.2 ± 0.1 **(1)** *(n/a)*	0.3-0.3 **(2)** *(0.1-0.5)*		2.4-3.2 **(4)** *(0.7-1.6)*	1.2-1.8 **(2)** *(n/a)*	0.3-3.4 **(3)** *(n/a)*	0.4 ± 0.1 **(3)** *(1.0 ± 0.4[a])*	0.4-0.9 **(4)** *(n/a)*
Underwater Sediment	22.3-30.0 **(6)** *(0.9-1.2*	22.3-27.2 **(6)** *(0.9-1.1)*	21.1-28.0 **(6)** *(0.9-1.1)*	15.4-32.1 **(6)** *(0.5-1.1)*	0.2-12.8 **(6)** *(0.8-1.3)*	19.3-31.0 **(34)** *(0.8-1.3)*	15.2-30.4 **(33)** *(0.9-1.2)*	20.2-30.9 **(36)** *(0.9-1.2)*	17.0-33.2 **(64)** *(0.8-1.2)*	16.3-24.3 **(18)** *(0.9-1.3)*

Notes:
For simplicity, results for both the total uranium activity and the isotopic $^{238}U/^{234}U$ ratio are reported as follows:





n/a denotes not applicable.
× denotes that no sample of this type was collected.
a: value reported for one sample only,
b: values reported for 3 samples only

- When there are multiple results for a sample type, the range (min – max) is reported. The analytical error for these results, which is reported in the yearly reports [6 -13, 15] varies per sample type. It is typically below 20% for sediments, but can be as high as 95% for biological samples. The number of samples is indicated in bold in brackets.
- Where the isotopic activities are below the Limit of Detection (LOD) for more than one isotope, the isotopic ratios could not be calculated and are reported as n/a.
- Notes a and b refer to the number of samples for which isotopic ratio was reported, if this is less than the total number of samples.

ANNEX B Issues to be considered when interpreting or comparing uranium data

A number of issues may give rise to uncertainties when interpreting or comparing uranium data:

- analytical technique and reporting conventions
- statistical variation;
- spatial variability;
- temporal variability; and
- species variation (for plant and animal samples).

Analytical approaches and reporting conventions

Sediment sample results may be reported as either dry weight or wet weight depending on whether the masses of the samples were obtained prior to or after drying. This will have implications for comparison of results between the surveys at Kirkcudbright, which are reported as dry weight and other UK uranium in sediments data, which may be reported as wet weight. Sediment samples reported as dry weight will appear to have concentrations of uranium at least 20% higher than those reported as wet weight (although this will depend on the moisture content).

For analysis techniques such as ICP-MS or alpha spectrometry, uranium may be extracted from samples and into solution by either leaching the sediment samples or totally dissolving them. Total dissolution will give rise to higher uranium results, because the analysis will include all uranium, including that which is contained within the mineral grains, whereas leached samples will only contain uranium that is either easily dissolved or is sorbed to the surfaces of mineral grains. Total sample analysis techniques such as gamma spectrometry will give results similar to those for total dissolution. Given the differences between the results for total analysis and leached analysis care should be taken when comparing sets of data to ensure that either the same approach has been used or that differences are appropriately discussed.

Uranium concentrations in seaweed may be affected by contamination of surfaces with sediment particles. Preparation of seaweed for analysis may or may not involve a washing stage. Hence, it is important to be aware of the preparation approaches that have been applied when comparing the results of seaweed analyses.

Seaweed and marine biota sample results may be reported as either dry weight or wet weight depending on whether the masses of the samples were obtained prior to or after drying. For marine biota, samples reported as dry weight will appear to have concentrations of uranium higher than those reported as wet weight. The relative increase in uranium concentration will depend on the moisture content, and can be as much as 300% for molluscs (whelks, scallops) [A1].

In addition, uranium activity (or concentration) results are expressed either as total uranium activity (or concentration), or as a breakdown of the ^{238}U, ^{235}U, and ^{234}U isotopic activities, or are expressed in terms of the ^{238}U activity (or concentration) only[9].

Statistical variations

There will be minor variations between the true uranium content of a sample and that reported by an analytical technique. This variation is highlighted in the counting statistics for the technique. The statistical uncertainties of laboratory results are likely to be small in comparison with the true variation in activity between samples.

Spatial variations

Any DU contamination will not be uniformly distributed within a sampling area, but will depend on the surface water flow, and the movement of tides and sediments. Hence, the repeat sampling and analysis of sediment from within an area may give rise to a significant degree of variation.

In addition to DU contamination due to firing at the KTA, there may be variations in uranium concentrations due to local anthropogenic or natural discharges. For example, natural uranium concentrations may be enhanced by the local application of phosphate based fertilisers to agricultural land. Veins rich in uranium minerals occur naturally along the coast of the Solway Firth, such as uraninite found at Needle's Eye on the north coast of the estuary. These features are thought to be present on a more regional basis, although this has not been studied [A2, A3].

Temporal variations

There will be natural temporal variations in the uranium concentration and abundance of the various isotopes in the seawater, for example, due to seasonal variations in rainfall. Rainfall can impact on the dissolution of uranium, and its migration in surface waters, which eventually enter the sea in coastal regions.

The activities of samples from any particular sampling site may vary from year to year. This may relate to temporal changes in uranium concentration, but will also be affected by spatial variation (see above).

Marine plant uptake of radionuclides is affected by the period in the plant growing cycle. This is also mirrored in the animal uptake of radionuclides within their life cycle.

Species variations

Plant uptake of radionuclides is affected by the substrate characteristics (uranium concentration and speciation as well as other physico-chemical characteristics) and varies with plant species.

[9] As seen in section 3, there are three broad types of uranium (natural, enriched and depleted). An increase or decrease in ^{238}U value (either concentration or activity) reflects a corresponding increase or decrease in the total uranium value. Within each type of uranium (i.e. for samples of identical uranium isotopic mix), the change in ^{238}U is proportional to the change in total U. Across all types however, the change in ^{238}U is not proportional to the change in total U, because each type has a different isotopic concentration and activity breakdown. Instead such a change is merely a qualitative indicator of a change in total U.

Animal uptake of uranium is affected by their life habits, feeding patterns, physiology and the uranium concentration in their foodstuffs and the environment. The marine animals that have been sampled were chosen for their importance in the human food chain. Animals that live in intertidal fringes have been selected to provide samples representative of a diet of sea food gathered by the consumer. In addition, animals that live on the sea bed were selected for their relevance to a diet of local sea food purchased by the consumer in the local Kirkcudbright area.

ANNEX C Variability of uranium concentration and uranium isotopic ratios in marine environmental samples

There are few specific examples that demonstrate the variability of uranium concentrations and isotopic ratios within marine environmental samples. To understand these, it is useful to consider the flux of uranium between the various components of the marine environment.

Uranium occurs naturally in seawater and its concentration generally varies in proportion to salinity. It is present in open seawater at an average concentration of 82.5 ± 5 Bq/m^3, with a $^{238}U/^{234}U$ activity ratio of 0.88 ± 0.03 at a salinity of 35 ‰. The isotopic ratio of seawater is below unity, as ^{234}U is preferentially mobilised from the soil during chemical weathering, thus enhancing its presence in seawater [A6, A4]. The physical mixing of low U- river water [10] and high U- sea water in estuary environments generally results in a dilution of uranium in the estuarine waters. As a result uranium concentrations in estuarine water are mostly lower than open ocean values, and increase linearly with salinity. [A6].

Concentrations of uranium in marine sediments are variable (from 32.5 to 1,625 mBq/g dry weight) and vary primarily as a function of the geology of the area.

Sample Type	Uranium concentration	Typical $^{238}U/^{234}U$ activity ratio
Ocean water	82.5 Bq/m^3	0.88
Estuarine water	<82.5 Bq/m^3	<0.88
Marine sediments (^{238}U only)	32.5-1,625 mBq/g	0.81

Table 11: Typical uranium concentrations and isotopic ratios in marine samples [A5, A6, A7].

Levels of uranium also vary depending on the precise location and depth of the sediment sample, as these directly affect its chemical characteristics, and hence its potential for concentrating uranium from the marine waters [A6, A7]. Both low oxygenation[11] and low salinity[12] provide favourable conditions for uranium scavenging from the water column by a variety of processes. These include the precipitation of U-rich colloids into the low salinity zone and the reduction of uranium into insoluble forms. [A7, A8]

[9] Concentrations of uranium in rivers vary considerably, with carbonate and dissolved solids concentrations, with an average of 7.5-15 Bq/m3, and an isotopic $^{238}U/^{234}U$ activity ratio of 0.77-0.83 as ^{234}U is preferentially mobilised during chemical weathering.

[11] Low oxygenation is found with increasing depth and increasing organic content

[12] Low salinity is dependent on river rate of flow, proximity to river outflow and depth of water.

ANNEX D Reference values for uranium in the Solway Firth and the UK

For UK coastal sediments, ^{238}U concentrations generally range from 3.6 and 32.3 mBq/g of dry weight [A8]. Values reported for shore sediment uranium concentrations in the Solway Firth vary with location and the specific uranium rich geology of the area:

Values for ^{238}U, ^{235}U and ^{234}U for sediment collected at Rascarrel Bay in 2005 were reported by CEFAS as 4.7, 0.39 and 5.4 mBq/g of dry weight respectively.

Values reported for ^{238}U concentrations at Sandyhills Bay, in South West Scotland, were lower than worldwide average values for seawater and sediments reported in Annex C. Seawater and sediments from Sandyhills had ^{238}U concentrations of 40 ± 3 Bq/m^3 and 14 ± 0.4 mBq/g dry weight[13], respectively [A9]. No other uranium isotope was measured, so no indication could be provided for the isotopic ratio.

However, semi-quantitative analysis of shore sediments samples obtained from the mudflats offshore from the uraninite vein, located a few miles east from Kirkcudbright at Needle's Eye, yielded uranium values of up to several hundred parts per million. These activities were measured both around open, oxygenated root channels and in near surface peat material. This analysis has suggested that uranium disperses seawards from the vein and ground waters, and is retarded in this by organic material in the sediments. [A3]

Sample	Location	Activity per wet weight (mBq/g)	Activity per dry weight (mBq/g)
Seaweed	[A9] Sandyhills [A9] UK		3.8 ± 0.1 3.8 – 18.6
Mussel	[A9] Sandyhills [A9] UK		1.1 ± 0.1 1.01 – 37.1
Mollusc	[A9] Sandyhills (winkle) [A9] UK [A15] UK (mollusc & winkle) [A15] Parton (winkle)	 0.89 1.3	2.72 ± 0.01 1.36 - 18.9
Crab	[A15] UK [A15] Parton	0.046 0.063	
Lobster	[A15] UK [A15] Parton	0.035 0.034	

Table 12: Reported ^{238}U concentrations in seaweed and marine biota in the UK

Literature values reported for ^{238}U in seaweed and marine biota samples are shown in Table 12 for the Solway Firth area; also reported are estimated values of ^{238}U from natural sources in aquatic foodstuff for the UK given by CEFAS [A15, A9].

[13] Sample preparation for the sediment sample included drying of the wet sample, grinding, homogenising, followed by acid leaching, electroplating and ICP-MS analysis for U.

Uranium-238 values reported for seaweed and marine biota at Sandyhills Bay were: ^{238}U concentrations of 3.8 ± 0.1 mBq/g dry weight for seaweed (*F. vesiculosus*) and 1.1 ± 0.1 mBq/g dry weight for mussel (*M. edulis*) [A9]. In 2005, ^{238}U concentrations for marine samples procured in Parton, near Whitehaven, ranged from 0.034 mBq/g wet weight for lobster, to 0.063 mBq/g wet weight for crab [A15]. However, it should be noted that the uranium inventory in Parton is dominated by historical anthropogenic input of natural uranium from a local phosphate processing plant.

ANNEX E Change in the $^{238}U/^{234}U$ activity ratio of a medium containing natural uranium with the addition of depleted uranium

Mass of depleted uranium added [14]	Activity [15] (mBq/kg)			Ratio of Total Activity Natural Uranium to Total Activity	$^{238}U/^{234}U$ Activity Ratio
	^{238}U	^{235}U	^{234}U		
0	3.7×10^4	1.7×10^3	3.8×10^4	1.0	1.0
1	7.4×10^4	2.2×10^3	4.3×10^4	1.6	1.7
2	1.1×10^5	2.7×10^3	4.8×10^4	2.1	2.3
3	1.5×10^5	3.2×10^3	5.3×10^4	2.6	2.8
4	1.9×10^5	3.7×10^3	5.8×10^4	3.2	3.2
5	2.2×10^5	4.1×10^3	6.3×10^4	3.8	3.6
6	2.6×10^5	4.6×10^3	6.8×10^4	4.3	3.9
7	3.0×10^5	5.1×10^3	7.2×10^4	4.9	4.1
8	3.4×10^5	5.6×10^3	7.7×10^4	5.4	4.3
9	3.7×10^5	6.0×10^3	8.2×10^4	6.0	4.5
10	4.1×10^5	6.5×10^3	8.7×10^4	6.5	4.7
20	7.8×10^5	1.1×10^4	1.4×10^5	12.0	5.8
60	2.3×10^6	3.1×10^4	3.3×10^5	34.1	6.9
80	3.0×10^6	4.0×10^4	4.3×10^5	45.1	7.1
100	3.8×10^6	5.0×10^4	5.2×10^5	56.1	7.2
200	7.5×10^6	9.8×10^4	1.0×10^6	111.0	7.4
600	2.2×10^7	2.9×10^5	3.0×10^6	332.0	7.6
800	3.0×10^7	$3..9 \times 10^5$	3.9×10^6	442.0	7.6
1000	$3.73\ 10^7$	4.8×10^5	4.9×10^6	552.0	7.6

Table from Volume 2 - Appendices, WS Atkins Environmental Assessment on DU Firings

[13] The value represents the additional mass of depleted uranium added (all radionuclides) relative to the original mass of natural uranium present (3 mg U/kg soil).

[14] Table assumes 3 mg/kg of natural uranium in a medium, prior to addition of DU, and present in the following proportion: ^{238}U (2.978 mg U/kg); ^{235}U (0.022 mg U/kg); ^{234}U (2e-04 mg U/kg).

ANNEX F List of references to annexes

[A1] Correspondance from Mr G Hunt (CEFAS) to Mr Phillips (MoD Directorate of Safety, Environment and Fire Policy), 23 February 2001.

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13. Abstract (A brief (approximately 150 words) factual summary of the report)

This report describes and interprets the results of the marine environmental survey undertaken to monitor DU levels in the environment at KTA in 2005. Samples of intertidal sediments, seaweed and seafood were collected from the shore line in the local area. Underwater sediment samples were collected offshore from the KTA, and samples of locally fished, bottom dwelling crustaceans and molluscs were also obtained. The results of the 2005 survey do not show evidence of the presence of DU in any of the environmental samples collected. There is no evidence to indicate that members of the public or site employees are, or have ever been, exposed to a radiological hazard from the marine environment as a result of test firing DU ammunition at the site.

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