Depleted Uranium Survey Report: Kirkcudbright Training Area 2008 Part 1 Terrestrial Environment

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Executive Summary

Depleted uranium (DU) ammunition has been test fired at the Kirkcudbright Training Area (KTA) since 1982. Routine environmental monitoring has been carried out at KTA since 1980 to assess the extent of any environmental impact of the firings on the terrestrial and marine environments and any associated radiological risk.

This report presents the findings of the terrestrial survey undertaken at KTA during 2008; the marine survey is reported separately in Part 2. The survey was undertaken to monitor the levels of uranium in the terrestrial environment resulting from operations on the site and to identify the extent of any environmental transfer processes.

None of the samples analysed were radioactive within the meaning of the Radioactive Substances Act 1993 (RSA93) nor did they exceed the Generalised Derived Limits (GDLs) advised by the Health Protection Agency (formerly the National Radiological Protection Board). Isotopic ratios indicate that low-level DU contamination was detected in 2 grass samples from the known areas of contamination at locations K3 (Raeberry Gun) and K4 (India Target), both of which are surrounded by a perimeter fence. Soil samples collected from location K5 (mid-point between Balig Gun and India Target) also indicated the presence of low-level DU contamination, although it is stressed that the levels present are very low.

Potential exposure pathways and doses to critical groups, site personnel and members of the public have been assessed and are deemed to be insignificant compared to the local natural background.

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1 Introduction

1.1 Depleted uranium (DU) ammunition has been test fired at the Kirkcudbright Training Area (KTA) since 1982. Routine environmental monitoring has been carried out at KTA since 1980 to assess the extent of any environmental impact of the firings on the terrestrial and marine environments and any associated radiological risk [1-11].

1.2 This report presents the findings of the terrestrial survey undertaken at KTA during 2008; the marine survey is reported separately in Part 2 [12]. The survey was undertaken to monitor the levels of uranium in the terrestrial environment resulting from operations on the site and to identify the extent of any environmental transfer processes.

1.3 The findings of a review of historic environmental monitoring procedures were reported in an independent assessment of the firing of DU projectiles at the KTA (and Eskmeals) ranges in July 1995 [13]. During 1996, the environmental monitoring programme for KTA was revised in line with the recommendations of that report and the 1996 DU Baseline Survey report [14] was published. Environmental monitoring was undertaken in a consistent manner between 1996 and 2007.

1.4 Due to altered work activities at KTA (with only one battery-target combination now available for use and greatly reduced proof firing of DU munitions) the MOD and the Scottish Environmental Protection Agency (SEPA) jointly agreed in 2007 that there should be greater emphasis on stream sediment sampling rather than soil sampling. The current terrestrial monitoring programme reflects this agreement and consists principally of the collection and analysis of grass and stream sediment/water samples, along with soil sampling around the active battery-target combination. Further details are given in the 2007 terrestrial environmental survey report [11]. Animal indicator samples are also collected when available.

2 Background

2.1 The KTA range is located on the coast of Dumfries and Galloway, near Castle Douglas. In April 2006, the range became part of the Defence Training Estate.

2.2 DU has been released into the environment at KTA as a consequence of the test firing of DU ammunition during design and accuracy assessment trials. DU projectiles are fired through soft vertical targets and continue their trajectory coming to rest in the Solway Firth.

2.3 Testing of projectiles historically has taken place at five locations on the Kirkcudbright range. Strength of design trials were initially conducted at the Raeberry range using prototype ammunition. Functionality and accuracy trials

were conducted until 2001 at the Balig and Doon Hill ranges, whilst confirmatory proving trials were carried out at the Silver Hill Low range. In 2003, rounds were fired from Challenger tanks placed at the Chapman 1000 metre firing point towards India Target.

2.4 The current firing policy is to use the Balig Gun and India Target combination. There was no DU firing at KTA between 2003 and 2007. A total of 20 rounds were fired in 2008 as part of a routine surveillance programme to ensure the safety and serviceability of the ammunition.

2.5 The number of DU rounds fired each year at KTA from the five firing locations and the cumulative mass of DU fired to date, are presented in Figure 1 and Figure 2 respectively.

2.6 DU penetrators do not fragment in air under normal circumstances. However, it was inevitable that some malfunctions would occur during the early test firings. Although the majority of malfunctioning penetrators still entered the Solway Firth, a small proportion impacted on land and some fragmented on exiting the gun barrel or in the early stage of flight. The majority of these fragments still entered the sea, but it is known that small quantities of particulate DU were deposited at a few discrete locations in the local terrestrial environment. Most of these locations are now fenced off, although the levels of DU present are below regulatory concern. The recovery of misfired DU penetrators has been attempted although in most cases penetrators are suspected to be buried at depth in the soil and therefore it has not been possible to locate or recover them.

2.7 An extensive radiological survey was carried out in 2002 using sensitive large area radiation monitoring equipment. Whilst identifying the expected isolated areas of low level DU contamination, the survey showed levels of radioactivity in most areas to be consistent with background levels. Other slight elevations in radiation levels were found to be due to naturally occurring radioactive material in construction materials such as the granite chippings used on the roads [15].

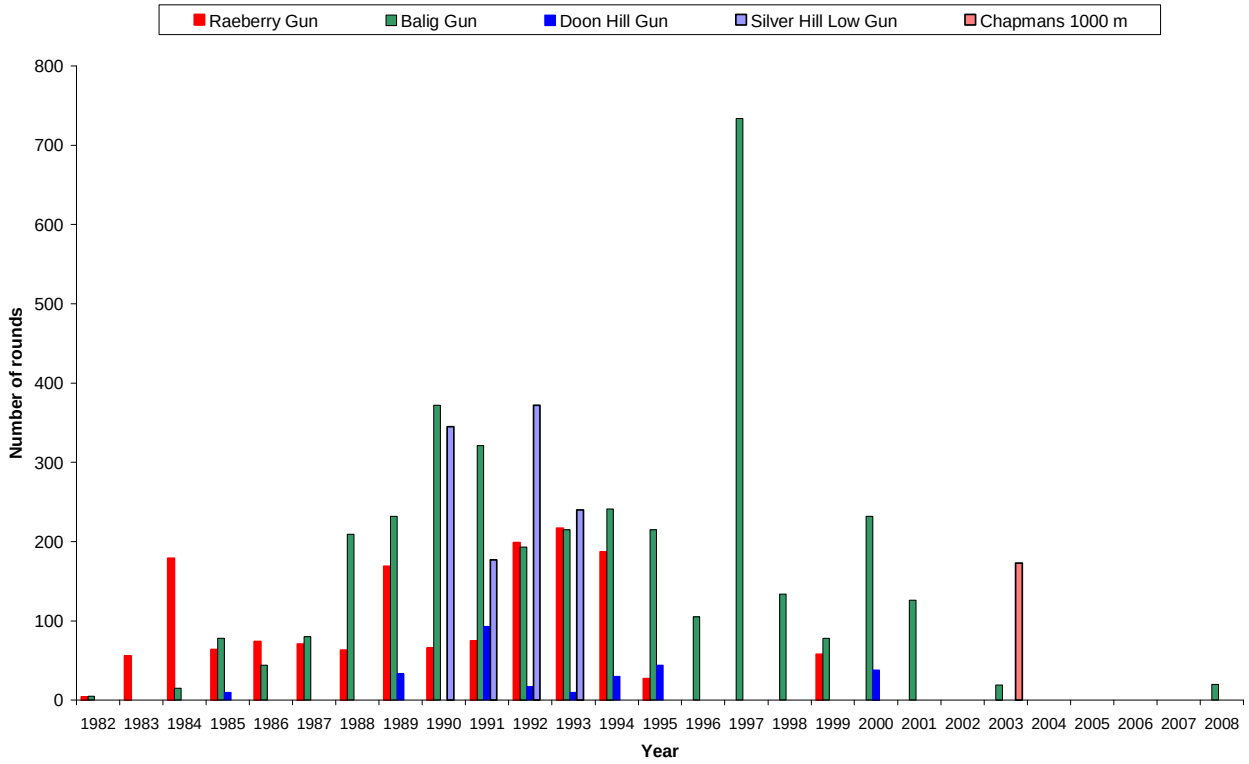

Figure 1. Number of DU projectiles fired from KTA between 1982 and 2008.

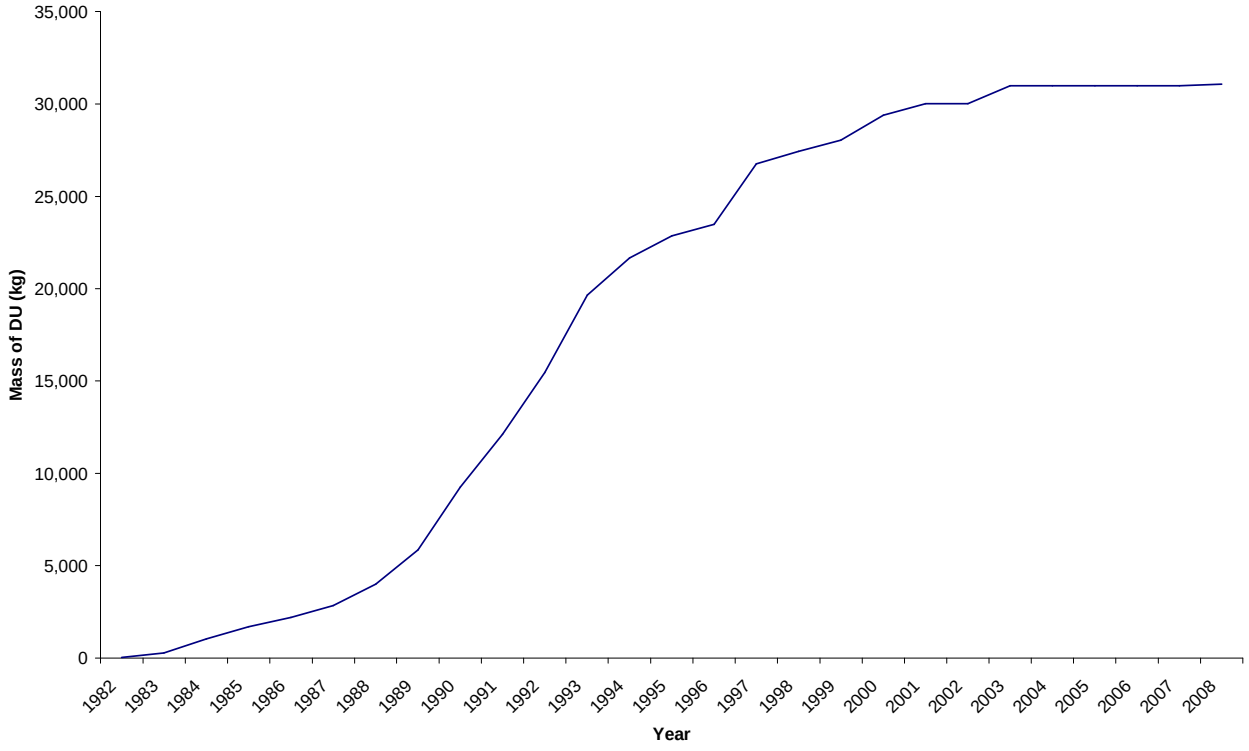

Figure 2. Approximate cumulative mass of DU projectiles fired from KTA between 1982 and 2008.

3 Depleted Uranium (DU)

3.1 Uranium is a naturally occurring radioactive material which exists mainly as three isotopes: uranium-238 (^{238}U), uranium-235 (^{235}U) and uranium-234 (^{234}U). The approximate mass composition of these isotopes is shown in Table 1. In the environment, natural uranium normally exists in approximate equilibrium with the daughter products of the ^{238}U and ^{235}U decay series[1] in terms of radioactivity. Together these isotopes emit a range of alpha and beta particles along with gamma radiation. The chemical toxicity of uranium is approximately equal to that of lead.

3.2 Uranium in an 'enriched' form is used as fuel in nuclear reactors. The enrichment process amplifies the concentration of ^{235}U (above 0.72%) in comparison to the natural form. The by-product of this process is 'depleted' uranium (DU), which has a reduced concentration of ^{235}U. Uranium-234 is also removed in the depletion process; DU is consequently less radioactive than natural uranium (the specific alpha activity of the DU fired at Kirkcudbright being approximately 1.4×10^7 milli becquerels per gram (mBq/g) compared to 2.5×10^7 mBq/g for natural uranium [16]). The mass compositions of DU and natural uranium are presented in Table 1 below.

Form of Uranium	^{238}U	^{235}U	^{234}U
Natural uranium	99.274%	0.72%	0.00554%
The DU used at Kirkcudbright	99.8%	0.20%	0.0008%

Table 1. Approximate mass compositions of uranium isotopes in natural and depleted uranium.

3.3 As discussed in paragraph 3.1, ^{234}U normally exists in approximate equilibrium with ^{238}U in the natural environment. In comparison, DU exhibits a ^{238}U/^{234}U activity ratio of between 7:1 and 8:1, dependant on the degree of depletion achievable by different methods of processing. This distinction is important in differentiating DU contamination from naturally occurring uranium in the terrestrial environment (see Section 4). For the remainder of this report, isotopic ratios will be stated in terms of activity rather than mass and as a single value representing the ratio of becquerels of ^{238}U to 1 becquerel of ^{234}U (i.e. a ^{238}U/^{234}U ratio of 7 rather than 7:1).

[1] A radioactive decay series occurs when a heavy radionuclide decays into successively lighter radionuclides. For example, ^{238}U decays to ^{234}Th, then ^{234m}Pa, then ^{234}U and so on until a stable element is reached (^{206}Pb).

4 Differentiating DU From Natural Uranium

4.1 The fundamental requirement of the DU environmental monitoring programme is to quantify the impact of DU firing. This is achieved partly by measuring the amount of total uranium in environmental materials and using this figure as an upper bound of DU contamination levels. However, as uranium is present at detectable levels in most environmental media, this overestimates the risk. More sophisticated analyses involve the specific measurement of ^{238}U and ^{234}U isotopes (by activity and/or mass). Although isotope measurements are used in this survey, references to total uranium measurements are included for consistency with historic reports. The limitations of using total uranium concentrations are discussed further in Annex A.

4.2 A convenient fingerprint marker for DU contamination is the $^{238}U/^{234}U$ activity ratio. The DU fired at KTA has a $^{238}U/^{234}U$ activity ratio of approximately 7, whereas natural uranium in the environment typically has an activity ratio close to unity. Environmental samples are therefore analysed for isotopes of ^{238}U and ^{234}U to determine activity ratios and hence identify the origin of the uranium.

4.3 Substantial deposition of DU in the terrestrial environment (in addition to an existing natural uranium background) is required before the $^{238}U/^{234}U$ activity ratio diverges significantly from its natural ratio. An illustration of the impact of DU contamination on the isotopic ratio is given in Annex B. For the ratio to approach 7 in an analytical sample, the mass of DU would have to be approximately one hundred times the mass of the uranium that is naturally present. Hence, the lower the natural uranium background, the lower the levels of DU contamination that may be detected by isotopic analysis.

4.4 Isotopic quantification is achieved by techniques such as alpha spectrometry and mass spectrometry. Alpha spectrometry can detect uranium to parts per billion, which is equivalent to mBq per kg, or to lower levels if count times are increased. Mass spectrometry is more sensitive, but the lower levels detectable are of no recognised health significance. Isotopic information can also be yielded from gamma spectrometry analyses, although limits of detection are not generally sufficient for measurement of environmental levels.

5 Reference Levels

5.1 The Depleted Uranium Firing Environmental Review Committee (DUFERC), on which the MOD is represented, has agreed investigation/action levels for levels of DU in soil[2]. These are currently based on, and represent a small fraction of, the Generalised Derived Limits (GDLs) advised by the Health Protection Agency (formerly the National Radiological Protection Board)[17] and the Schedule 1 activity concentration for uranium laid down in the Radioactive Substances Act 1993 (RSA93) [18]. GDLs for uranium were last updated in 2000 and were referred to by the Royal Society in their studies of the potential health effects of using DU munitions [19]. The investigation/action levels are set intentionally low to ensure that any DU released into the environment is identified before it can accumulate to significant levels. In particular, action levels are set to less than 10% of the level at which control would be required under RSA 93.

Source	Reference Level	Activity concentration (mBq/g dry weight)
DUFERC	Investigation Level	300
DUFERC	Action Level	1,110
RSA 93	Level at which regulatory control is required.	11,100
NRPB (2000)	Generalised Derived Limit: ^{238}U in well-mixed soil	20,000
NRPB (2000)	Generalised Derived Limit: ^{238}U in freshwater sediment	400,000

Table 2. DUFERC investigation/action levels, RSA93 Schedule 1 activity concentrations and Generalised Derived Limits for uranium in soil.

5.2 There are no DUFERC agreed investigation/action levels for uranium in water samples. The alpha spectrometry system deployed by Dstl can detect uranium in water at levels down to about 10% of the World Health Organisation's (WHO) most restrictive recommendation for uranium levels in drinking water (2 μg per litre). This 10% level is commonly used as a 'trigger' in occupational health monitoring.

[2] Soil in areas of contamination above the DUFERC investigation or action level shall be managed in accordance with the draft KTA Depleted Uranium Management and Remediation Plan [20].

6 Methodology

6.1 As discussed in Section One, the current survey methodology consists primarily of the collection and analysis of grass and stream sediment/water samples, along with soil sampling around the active battery-target combination. Animal indicator samples are also collected when available and environmental gamma dose rates are recorded at each sampling location. Stream sediment/water sample points are located at areas where any potential contamination is likely to leach from the surrounding soil.

6.2 Full details of the methodology are provided in the following paragraphs. A list of the soil, grass and faecal samples collected is given in Table 3. A list of stream sediment and water samples collected is given in Table 4.

Terrestrial sampling sites

6.3 For each gun-target combination, sampling sites were chosen in the vicinity of the gun position; at the soft target stand; and at the mid-point of the range. In addition, samples collected from the Gypsy Point location were used for the purpose of background comparison. Grass samples were collected from all locations whereas soil samples were only collected from locations 4, 5 and 6 (Balig-India combination) and the background location (13). The 13 sampling points are shown in Figure 3; the location names and Ordnance Survey of Great Britain grid references are provided below:

1.	Raeberry Target	NX 70449 43744
2.	Raeberry Bunker Midpoint	NX 70472 43836
3.	Raeberry Gun	NX 70521 43980
4.	India Target	NX 70631 43658
5.	Balig Gun/Target waypoint	NX 70964 44498
6.	Balig Gun	NX 71238 45452
7.	Zulu Gantry	NX 71848 43548
8.	Mullock Farm	NX 71188 44295
9.	Silver Hill (Low) Gun	NX 70355 44851
10.	Echo Target (Doon Hill)	NX 72291 43807
11.	Doonhill/Target waypoint	NX 72096 45429
12.	Doon Hill Gun	NX 71904 46947
13.	Gypsy Point	NX 68676 43789

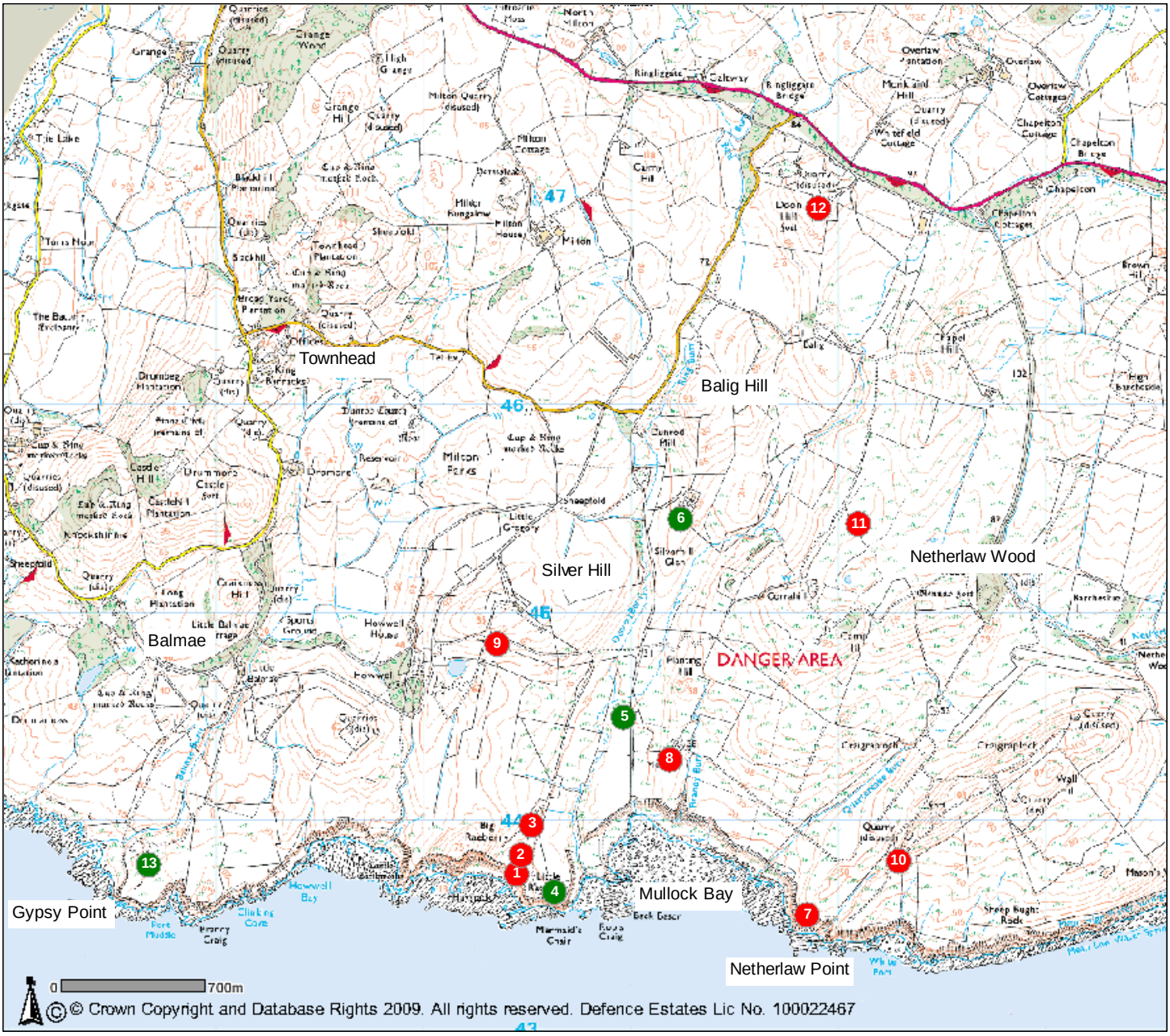

Figure 3. Terrestrial sampling locations at KTA 2008. Note: red markers indicate grass sampling; green markers indicate grass plus soil sampling.

6.4 At each sampling site, grass samples were collected from three areas of approximately 1 m2 in size located within 5 metres of each other. Where the grass was scarce or short; the sampling area was increased until the samples obtained were of the requisite mass for laboratory analysis (greater than 200 g). The grass was cut at a height of at least 2 cm above the ground to avoid including soil in the sample. The type of grass collected and the content of other plant species varied from site to site. The 3 unwashed grass samples from each site were combined into one composite sample and analysed by alpha spectrometry.

6.5 Soil samples were collected as undisturbed cores of 4.8 cm diameter and up to 30 cm in depth from the centre of each of the three 1m2 grass sampling areas. Each soil core was divided into sub-samples of two depth intervals (0-2 cm and 2-5 cm)

in the laboratory and portions of the three sub-samples from each point combined to create one composite sample for each depth interval (0-2 cm and 2-5 cm). This system yielded a total of 8 composite samples from the 4 soil sample locations. These were analysed by alpha spectrometry to provide information on distribution with depth and hence give an indication of the mobility of any surface deposited DU.

6.6 At each grass sampling area, gamma dose rates were derived from measurements taken using a Mini-Instruments 6-80 (compensated Geiger-Muller tube) set up at 1 m above the ground and allowed to record over a period of 300 seconds. A mean gamma dose rate was calculated from three replicate measurements.

6.7 Where available, samples of animal faeces (e.g. deer, rabbit, fox, badger, sheep and cattle) were collected close to the sampling sites. Only fresh samples were collected, taking care to minimise the amount of soil or grass that could inadvertently be sampled at the same time. Rabbit droppings were collected until enough material had been sampled; making one composite sample for that location. These samples were analysed by alpha spectrometry.

6.8 Deer culling takes place regularly on the range to manage the deer population. Following recommendations to extend the monitoring to include biological samples from wild animals that roam on the KTA, kidney samples (and liver samples in 2008) were collected from two culled deer in 2008. Both animals were less than 2 years old. The samples were analysed by alpha spectrometry.

Stream sediment and water sampling

6.9 Stream sediment was collected at fifteen locations from streams that collect run-off water from the soil surface of the range. Stream water samples were collected at five locations. The Ordnance Survey of Great Britain grid references for the fifteen stream sediment and water sampling sites used in 2008 are given below. The exact position of the sampling sites changes slightly from year to year due to changing stream conditions and access. The locations are shown on a map of the site in Figure 4.

Stream sediments sampling grid references:

S1.	Burnfoot Bridge	NX 74205 44583
S2.	Netherlaw Wood	NX 74165 44638
S3.	Netherlaw Burn,	NX 73438 44779
S4.	Quatercake Burn	NX 72326 44363
S5.	Quatercake Burn	NX 71804 43917
S6.	Brandy Burn	NX 71264 44145
S7.	Dunrod & Overlaw Burns	NX 70820 43732
S8.	Dunrod Burn	NX 70958 44825
S9.	Dunrod Burn	NX 71062 45536
S10.	Ring Burn	NX 71137 45948
S11.	Overlaw Burn	NX 71167 44843
S12.	Overlaw Burn	NX 72124 46277
S13.	Overlaw Burn	NX 72514 46942

S14. Balmae Burn NX 69138 44651
S15. Balmae Burn NX 68556 43890

Stream water sampling grid references:

W1. Abbey Burn NX 74209 44611
W2. Netherlaw Burn NX 74202 44594
W3. Balmae Burn(Gypsy Point)NX 68556 43890
W4. Dunrod & Overlaw Burns NX 70820 43732
W5. Quartercake Burn NX 71804 43917

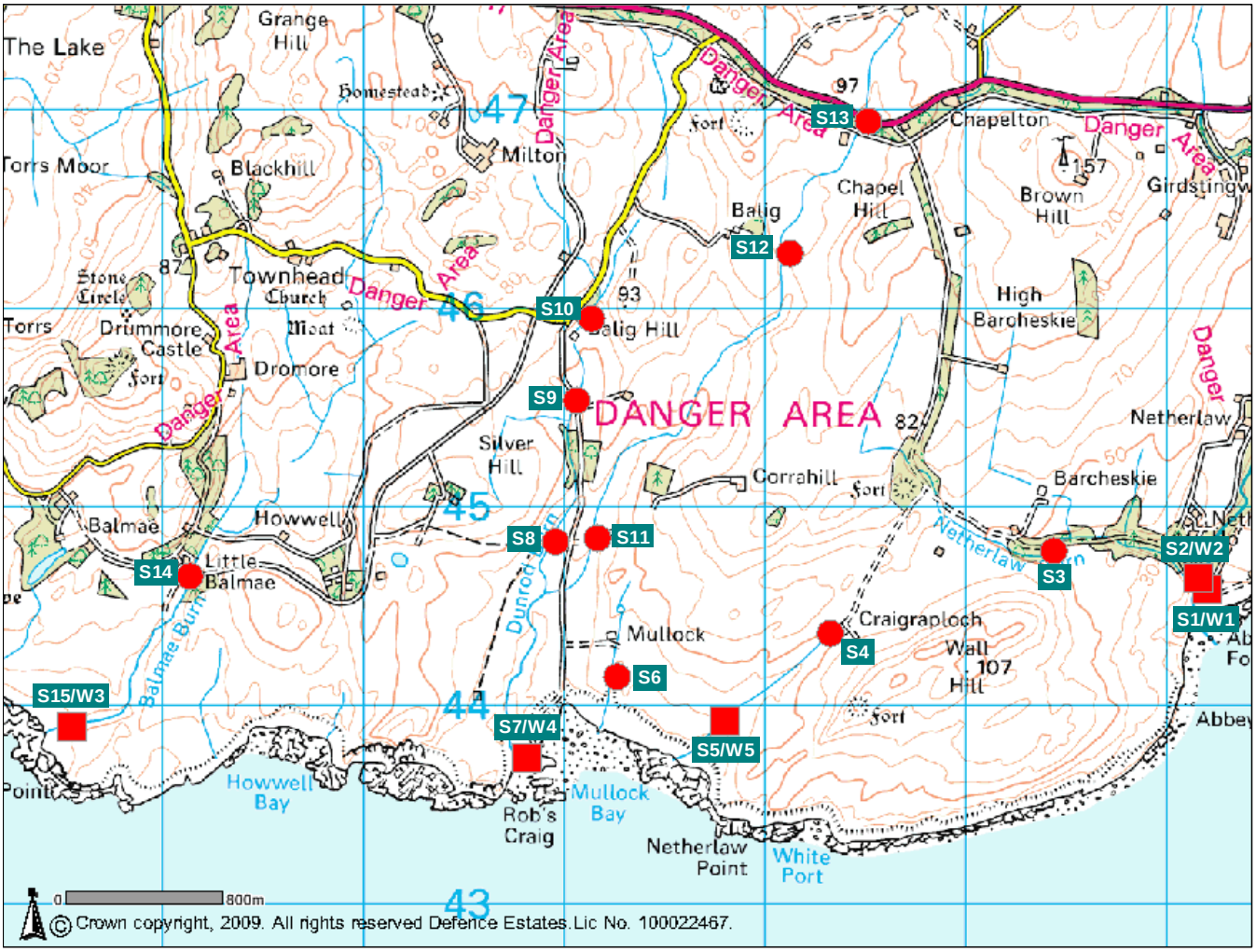

Figure 4. Stream sampling locations - KTA 2008. Note: circular markings denote sediment sampling; square markings denote both sediment *and* water sampling.

6.10 Stream sediment was collected from below the water where possible. Care was taken to sample undisturbed sediments which had not been disturbed by cattle, for example. The top layer of the sediment (up to 5 cm depth) was collected and any large stones were removed. Sample size ranged from 0.4 to 0.8 litres. The sediment samples were analysed by alpha spectrometry.

6.11 Five water samples (0.5 litre) were collected from watercourses that run through the range. Samples were collected at accessible locations along the stream (e.g.

from bridges and easily accessible banks). To remove any suspended sediment, the water was filtered through a Whatman filter paper using a funnel previously rinsed in stream water into a rinsed sample bottle. Both the water and the filter paper were analysed by alpha spectrometry.

Sample Descriptors

6.12 Each sample was given a unique sample descriptor. Those samples collected from specific sites (e.g. grass samples) were given a prefix such as K5 ('K' for Kirkcudbright followed by the location number). This prefix was followed by 'S' for soil sample; a 'G' for grass samples or the type of faecal sample. Soil samples were also denoted with a 'U' or a 'L' for the upper and lower soil horizon composite samples respectively. Deer kidney (and liver samples this year) were simply named 'Deer kidney (a) or (b)' and water/sediment samples were given descriptors such as W3 and S13 respectively. Some examples of sample descriptors are provided below:

K5/S/U : Composite soil sample, upper horizon, collected from location 5.

K8/G : Composite grass sample collected from location 8

K3 rabbit : Rabbit faecal samples collected from location 3

W2 : Water sample collected from water sample location 2

S12 : Stream sediment sample collected from stream location 12

Sample preparation and laboratory analysis

6.13 The samples were prepared and analysed by alpha spectrometry in the Dstl UKAS accredited radiochemistry laboratory following the procedure adopted for the analysis of the terrestrial samples during the 1996 Baseline Survey [14]. An outline of the approach is given below.

6.14 The solid samples (soil, grass, animal faeces, deer kidneys (and deer livers for this year's survey)) were dried to remove moisture and weighed (with results being reported as dry weight). The samples were ashed to remove organic material and homogenised. During these processes, the loss of uranium from the sample is deemed to be insignificant. The ashed samples were boiled in concentrated mineral acid (nitric acid and hydrochloric acid) to remove the 'loose' and leachable uranium from the sample. Recalcitrant matrices such as mineral grains were not broken down by the process and hence natural uranium bound up within them was not removed. The samples were filtered to remove solids. Water samples of 250 ml were boiled down to approximately 100 ml and acidified.

6.15 Uranium separation was carried out by extraction chromatography. Each eluted sample was electro-deposited onto a stainless steel planchette and the activity of each planchette was counted in a low background alpha spectrometer with a silicon surface barrier.

6.16 Uranium activity concentrations are reported in mBq/g (equivalent to Bq/kg) of dry weight for soils, grass and biological indicators and mBq/l for water samples.

7 Results and Interpretation

7.1 A summary of the results for all terrestrial samples collected in 2008 is given below in Table 3. The full terrestrial monitoring results are provided in Tables 6 to 13 in Section 12. Historical monitoring results for the KTA for the years 1996 to 2008 are presented in Appendix A.

7.2 It should be noted that the summary below is provided as an overview of the data collected. Given that the samples were generally selected from specific areas of known contamination, the mean values should not be viewed as an indicator of average uranium concentrations across the KTA as a whole.

Sample type	Number of Samples	Number of samples containing detectable DU[3]	Total uranium concentration (mBq/g or mBq/l)			
			Mean	Standard deviation of the mean	Minimum	Maximum
Soil upper	4	1	76	52.9	35.3	152.2
Soil lower	4	1	108	141.1	34.0	319.6
Grass	13	2	0.4	0.4	0.1	1.5
Water	5	0	4.7	1.6	2.7	6.4
Stream sediment	15	0	29.3	5.8	16.4	38.5
Faeces (Sheep)	1	0	7.4	-	7.4	7.4
Faeces (Rabbit)	3	0	1.1	0.4	0.7	1.4
Faeces (Cow)	4	0	1.1	0.3	0.9	1.5
Faeces (Deer)	1	0	0.5	-	0.5	0.5
Kidney (Deer)	2	0	0.1	0.02	0.1	0.1
Liver (Deer)	2	0	0.1	0.02	0.1	0.1

Table 3. Summary of sample analyses - KTA 2008.

Grass sample analysis

7.3 Alpha spectrometry results for grass samples are shown in Table 6. The activity concentration of total uranium ranges from 0.1 ± 0.1 mBq/g to 1.5 ± 0.2 mBq/g. This range lies within the typical values stated in literature (0.2 to 3.8 mBq/g) [20].

[3] Samples are reported as containing detectable DU if the ratio of ^{238}U to ^{234}U (after subtraction of the associated uncertainty to give the 95% confidence level) is greater than 1.0 for soil, grass and biota samples.

7.4 The activity concentration found at the background site (Gypsy Point - K13) was 0.2 ± 0.1 mBq/g. Grass samples from six locations exhibited activity concentrations in excess of at least twice this background level:

- Raeberry Gun (K3) : 0.6 ± 0.1 mBq/g
- India Target (K4) : 1.5 ± 0.2 mBq/g
- Balig Gun (K6) : 0.7 ± 0.2 mBq/g
- Zulu Gantry (K7) : 0.4 ± 0.1 mBq/g
- Mid-point between 7 & 9 (K8) : 1.0 ± 0.2 mBq/g
- Mid-point between 10 & 12 (K11) : 0.5 ± 0.1 mBq/g

7.5 Only two of these samples showed isotopic ratios indicative of DU contamination: K3 (Raeberry Gun) and K4 (India Target). Isotopic values of grass samples are discussed further in Section 8.

Soil sample analysis

7.6 Alpha spectrometry results for soil samples are shown in Table 7. The levels of total uranium found at the background location (K13) were 34.0 ± 3.4 mBq/g (lower) and 70.1 ± 6.7 mBq/g (upper)[4]. The quantities of uranium found in the soil from K4 and K6 were consistent with this background level. However, sample location K5 (mid-point between Balig Gun and India Target) showed levels of total uranium of 152.2 ± 15.0 mBq/g (upper) and 319.6 ± 31.7 mBq/g (lower). Isotopic ratios indicate that the contamination is likely to be due to DU, as discussed further in Section 8. The highest level was found at the lower soil horizon suggesting that the contamination is due to DU firing in the past, rather than more recent operations on site.

7.7 Although the level of total uranium in the lower soil fraction at K5 exceeds the DUFERC Investigation level of 300 mBq/g, it is emphasised that the material is below the level of regulatory concern (RSA93 Schedule 1 limit of 11,100 mBq/g) and that the amount of ^{238}U equates to approximately 1% of the relevant GDL. The area will be managed in accordance with the KTA DU Management and Remediation Plan [20].

Animal indicator sample analysis

7.8 Alpha spectrometry results for faecal samples and deer kidney/liver samples are shown in Tables 8 and 9 respectively. No DU was detected. The highest total uranium level was 7.4 ± 0.8 mBq/g in one sheep faecal sample. This result is

[4] Soil minerals containing uranium are widely distributed on the surface of the Earth's crust and the concentrations of natural uranium in the terrestrial environment can vary significantly between locations. Consequently, there is no single definitive reference level for natural uranium in soils. However, there is broad agreement in the range of values published in the literature: typical values in the UK range from 2 to 50 mBq/g wet weight, but concentrations of up to 100 times the typical range can be found in some locations.

similar to that found in a sheep faecal sample in the 2007 survey. The total uranium content of all other samples ranged from 0.5 to 1.5 mBq/g with isotopic ratios consistent with natural uranium.

7.9 No isotopes of uranium were detected in the deer kidney/liver samples above the limit of detection of 0.1 mBq/g for alpha spectrometry. In comparison, the GDL for offal is 1000 mBq/g for ^{238}U.

Stream water sample analysis

7.10 Alpha spectrometry results for stream water and filter paper samples are presented in Table 10 and 11 respectively. The levels of total uranium in water ranged from 2.7 ± 2.2 mBq/l to 6.4 ± 4.7 mBq/l. These levels are well below the previously mentioned WHO limit of 2 μg/l for uranium which equates to approximately 30 mBq/l. Uranium isotopes were not detected above the limits of detection; meaning that isotopic ratios could not subsequently be calculated. Total uranium levels in the associated filter papers analysed did not exceed 1 mBq/g.

Stream sediment sample analysis

7.11 Alpha spectrometry results for stream sediments are presented in Table 12. Total uranium activity concentrations ranged from 16.4 ± 2.0 to 38.5 ± 4.0 mBq/g, which is consistent with the levels found in previous years. Isotopic ratios are indicative of natural uranium and the total uranium content represents less than 0.01% of the GDL for freshwater sediment of 400,000 mBq/g [17].

Environmental dose rate measurements

7.12 Environmental gamma dose rate measurements recorded during 2008 are shown in Table 13. The results are consistent with the background radiation measurement recorded at Gypsy Point as well as measurements recorded in previous surveys.

8 Interpretation of Soil and Grass Isotopic Ratios

8.1 This section provides an assessment of the $^{238}U/^{234}U$ isotopic ratios within samples to ascertain the degree of depletion and hence the extent of DU contamination. Further discussion on the use of isotopic ratios to differentiate DU from natural uranium is given in Section 4. An illustration of the impact of increasing the DU mass in a sample on the isotopic ratio is provided in Annex B.

8.2 Provided below are graphical interpretations of grass and soil isotopic ratios together with an explanation of the findings. As no DU was detected, i.e. no isotopic ratios significantly above unity were identified for other sampling media, graphs for animal indicators and stream sediment/water samples are omitted.

Isotopic ratios in grass samples

8.3 A graphical summary of the isotopic ratios for the grass samples collected in 2008 is provided in Figure 5. Where no isotopic value is shown in the graph, uranium isotopes have not been detected above the limits of detection.

8.4 Grass samples with $^{238}U/^{234}U$ isotopic ratios of greater than 1 (within the analytical limits of uncertainty) were found at sample locations K3 (Raeberry Gun position) and K4 (India Target area). These findings are consistent with those of previous surveys. The total levels of uranium present still lie within the typical natural background range of 0.2 to 3.8 mBq/g stated in the literature [21]. They are also well below the GDL of 20 Bq/g for sea-washed pasture [17].

Isotopic ratios in soil samples

8.5 A graphical summary of the isotopic ratios for the soil samples analysed in 2008 is shown in Figure 6. Soil samples recovered from location K5 (mid-point between Balig Gun and India Target) exhibited elevated levels of total uranium with isotopic ratios which indicate that the contamination is due to DU. The isotopic ratio of the upper and lower soil fractions were 4.1 ± 0.7 and 5.3 ± 0.9 respectively.

8.6 All other isotopic ratios found in soil samples range from 0.9 to 1.1 and are therefore indicative of natural uranium.

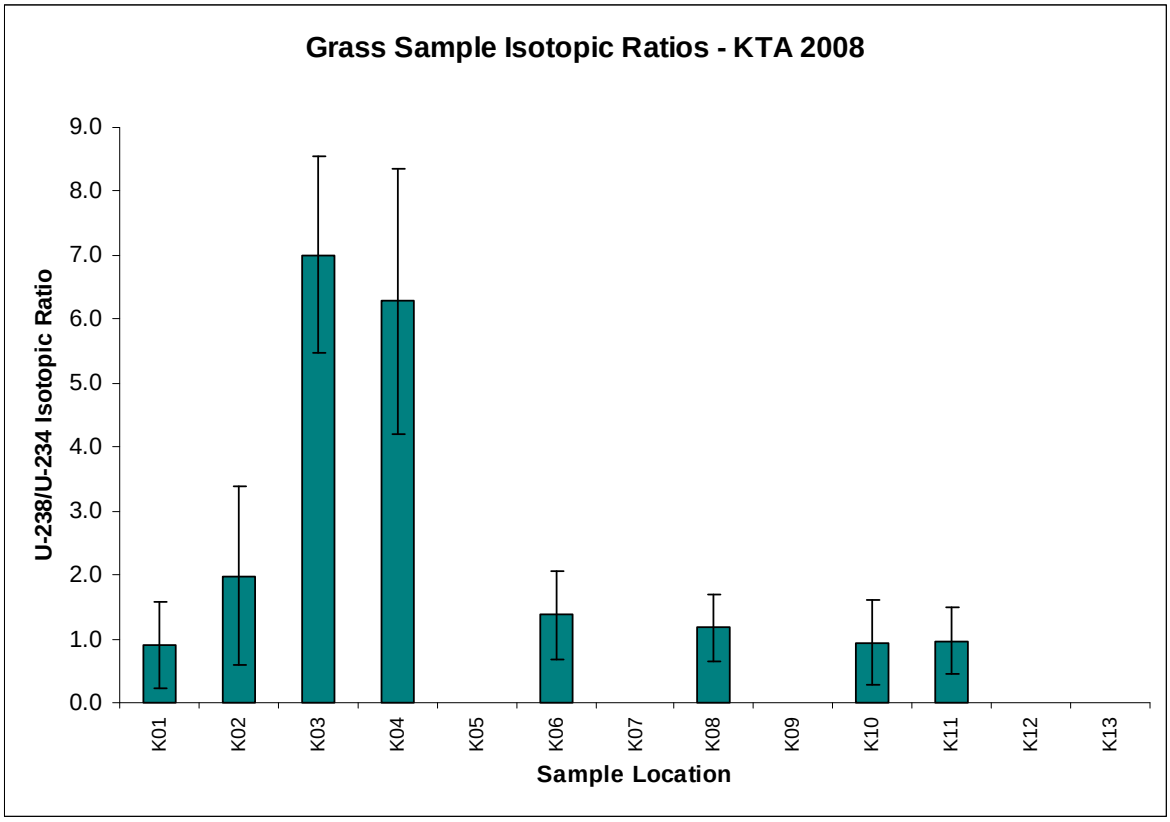

Figure 5. Grass samples: isotopic ratios from alpha spectrometry analysis - KTA 2008.

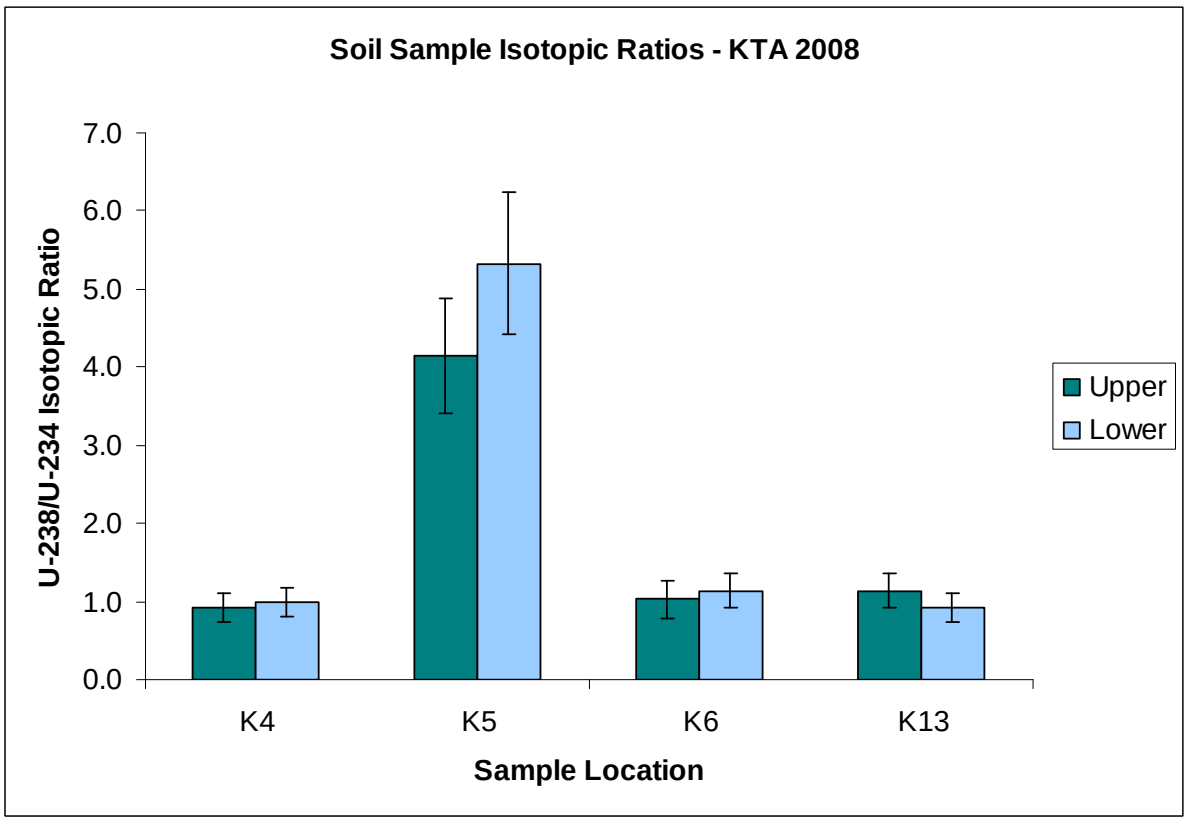

Figure 6. Soil samples: isotopic ratios from alpha spectrometry analysis - KTA 2008.

9 Evaluation of Potential Exposure Pathways

9.1 Any contamination of the terrestrial environment with DU results in five potential exposure pathways for humans, as described below:

- External radiation exposure from contaminated vegetation and soil or from DU fragments;
- Inhalation of DU contamination released into the air or re-suspended from vegetation, soil or sediment;
- Ingestion of crops or animal products from DU contaminated pasture or soil;
- Ingestion of DU contaminated water from streams;
- (Inadvertent) ingestion of DU contaminated soil or sediment.

9.2 As discussed in Section 7, the total uranium activity results for all samples represented a fraction of the relevant GDLs and WHO limit for drinking water. Therefore, any associated radiation doses are deemed to be insignificant. For completeness, the potential doses from the highest level of uranium found in grass and soil samples are considered below, together with a more general dose assessment for the range as a whole.

External radiation exposure

9.3 Environmental gamma dose rates recorded across the KTA during 2008 are consistent with natural background radiation. This is also the case where low levels of DU contamination have been found (e.g. at location K5) and indicates that there is no increased risk from external exposure related to DU at KTA.

Inhalation of re-suspended DU

9.4 DU that has been deposited on soil, river sediment and vegetation may be re-suspended into the air and subsequently inhaled by range staff or members of the public. The levels of DU found in soil during 2008 are well below the GDL for well-mixed soil (20,000 mBq/g) which itself relates to a Committed Effective Dose of 1 mSv. This is the current UK annual dose limit for a member of the public. The potential doses relating to this contamination are not, therefore, deemed to be significant.

9.5 The level of total uranium found in soil at K5 was 319.6 ± 31.7 mBq/g (at 2 to 5 cm depth). The hazard associated with re-suspended DU would be greatest during intrusive work which may take place on site (e.g. excavation). Using pessimistic assumptions, approximate doses due to inhalation are calculated to be between 20 and 50 μSv Committed Effective Dose. This is well below the 1 mSv annual dose

limit mentioned above. Furthermore, no intrusive work is known to take place in the area so exposure through this pathway does not presently exist.

Ingestion of DU contaminated foodstuffs

9.6 No agricultural crops intended for human consumption are grown on the range [13] so the potential exposure route involving the ingestion of contaminated crops need not be considered. However, as a small number of cattle, sheep and deer graze on the range, consideration must be given to the exposure pathway of ingestion of animal products derived from contaminated soil or pasture. It should be noted, however, that the known areas of contamination on site are enclosed within fenced areas so any potential dose estimates represent a worst case scenario.

9.7 Reference [22] states that the equilibrium soil to plant transfer ratio for pasture is $1x10^{-3}$ for uranium. Given that the highest level of uranium found in soil during the survey was 319.6 ± 31.7 mBq/g, this would indicate an approximate maximum uranium level in grass of 0.3 mBq/g (wet weight). Higher levels of uranium of up to 1.5 ± 0.2 mBq/g were found in grass collected during this survey. However, it is possible that the uranium detected could be due to external contamination or small amounts of soil attached to the grass rather than uranium taken up from the soil itself. The important point is that even the highest contamination levels identified represent only a small fraction of the relevant GDLs (which consider doses due to ingestion). Furthermore, DU contamination is only present in small areas which are widely separated. The degree of transfer to animal products is therefore deemed to be negligible and the fact that no DU was detected in the deer kidneys/liver provides further support for this conclusion.

9.8 Wild rabbits, pheasants and deer that feed and graze on the range may occasionally be consumed. As mentioned above, no DU has been detected in deer kidney/liver samples above the limit of detection (the GDL for offal is 1Bq/g). Analysis of faecal samples has been used historically to give an indication of the potential contamination of these wild animals. However, it should be noted that this can only be an approximation as the DU may have been transferred from soil or grass on the surface on the faeces rather than being a constituent of the faecal material. However, the uranium detected in faecal samples in this survey was found to be of natural origin rather than due to DU. This is consistent with the findings of previous years in which it was concluded that any potential exposures would be negligible.

Ingestion of DU contaminated water from streams

9.9 The surface water samples collected in this survey contained no detectable DU contamination. In all cases the total uranium concentration was towards the lower end of what is found naturally and the highest result was approximately six times lower than the WHO limit for drinking water mentioned previously. Furthermore, the Burns are not known as a drinking water source and regular consumption by a single individual or group is not considered plausible. It is concluded that any radiation dose received via the theoretical consumption of DU contaminated surface water is insignificant.

Ingestion of DU contaminated soil or stream sediment

9.10 While this survey has identified one area of DU contamination in soil which exceeds the DUFERC investigation level, this is of no health significance as the level found is well below the GDL for well-mixed soil. Furthermore, the likelihood of this exposure pathway occurring is considered to be very low as there are no routine intrusive ground-works which may inadvertently lead to the ingestion of contaminated soil. No DU was detected in any stream sediment.

Radiation exposure to critical group

9.11 Using local knowledge of the range, two groups of people are deemed to constitute critical groups for the potential radiological doses associated with DU released into the KTA terrestrial environment.

9.12 Workers employed to inspect and maintain the boundary fence on the firing ranges are likely to be at most risk because of their regular presence on the site. They form a critical group whose main exposure pathway is likely to be via inhalation of re-suspended DU contamination from soil or vegetation. Based on the findings of this report, their potential worst case dose is considered to be minor.

9.13 Local inhabitants that have access to venison or meat from the wild animals that roam on the range constitute the other critical group. The results of this year's survey are consistent with those of previous years in which it was concluded that any potential exposures to either group would be insignificant.

10 Conclusions

10.1 The 2008 Kirkcudbright terrestrial monitoring programme was undertaken to assess the levels of DU in the environment resulting from firing of DU munitions. The monitoring programme consisted principally of the collection and analysis of grass and stream sediment/water samples, along with soil and grass sampling around the active battery-target combination. Animal indicator samples were also collected.

10.2 Stream water and sediment samples were found to contain levels of uranium consistent with natural background levels. Isotopic ratios also indicate that this is due to natural uranium rather than DU resulting from firing on the site.

10.3 Isotopic ratios indicate that DU contamination was detected in 2 grass samples from locations K3 (Raeberry Gun) and K4 (India Target). The levels of total uranium were 0.6 ± 0.1 and 1.5 ± 0.2 mBq/g respectively. This is elevated above the background level recorded at Gypsy Point, but lies within the range of common UK values reported in the literature (0.2 to 3.8 mBq/g)[20]. These results are not unexpected as the samples were collected from known fenced areas of low level DU contamination, as reported in previous survey reports [1-11]. While some other grass samples showed levels of uranium in excess of the background level, isotopic ratios indicate that the uranium is natural in origin.

10.4 Soil samples collected from location K5 (mid-point between Balig Gun and India Target) showed levels of total uranium of 152.2 ± 15.0 mBq/g (upper) and 319.6 ± 31.7 mBq/g (lower). This level of uranium is significantly higher than the level recorded at the background location and isotopic ratios indicate that the contamination is likely to be due to DU. However, it is emphasised that the levels represent an insignificant health risk and are far below the GDL for well-mixed soil and the RSA93 Schedule 1 limit.

10.5 Potential exposure pathways and doses to critical groups, site personnel and members of the public have been assessed and are deemed to be insignificant compared to the local natural background and considering the known ongoing use of the site.

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12 KTA Terrestrial Survey Results

Sample station number	Sample type and number of samples			Environmental dose rates
	Soil	Grass	Animal indicators (faeces)	
K1	×	1	1	✓
K2	×	1	×	✓
K3	×	1	1	✓
K4	1	1	1	✓
K5	1	1	×	✓
K6	1	1	1	✓
K7	×	1	1	✓
K8	×	1	1	✓
K9	×	1	×	✓
K10	×	1	1	✓
K11	×	1	1	✓
K12	×	1	×	✓
K13	1	1	1	✓

Table 4. Summary of soil, grass and animal indicator samples collected - KTA 2008.
Note: × denotes sample not collected.

Sediment sample number	Water sample number	Location name	Burn
S1	W1 & W2	Burnfoot Bridge	Abbey and Netherlaw
S2	×	Netherlaw Wood	Netherlaw
S3	×	Cross roads	Netherlaw
S4	×	Craigrapploch	Quatercake
S5	W5	Downstream	Quatercake
S6	×	Mullock Farm	Brandy
S7	W4	Mullock Bay	Dunrod and Overlaw
S8	×	Upstream	Dunrod
S9	×	Balig Gun turn off	Dunrod
S10	×	Dunrod Mill gate	Ring
S11	×	Overlaw Burn	Overlaw
S12	×	Bailey Bridge	Overlaw
S13	×	EM Gun	Overlaw
S14	×	Gypsy Barrier	Balmae
S15	W3	Gypsy Point	Balmae

Table 5. Summary of stream sediment and water samples collected - KTA 2008.

Sample descriptor	Fresh weight (g)	Dry weight (g)	Ashed weight (g)	Measured activity of dry sample (mBq/g)				$^{238}U/^{234}U$ ratio
				^{238}U	^{235}U	^{234}U	Total U	
K01	129.7	50.0	3.5	0.1 ± 0.0	< 0.1	0.1 ± 0.0	0.2 ± 0.1	0.9 ± 0.7
K02	169.6	65.5	4.2	0.1 ± 0.1	< 0.1	0.1 ± 0.0	0.2 ± 0.1	2.0 ± 1.4
K03	168.6	62.1	3.7	0.5 ± 0.1	0.1 ± 0.0	0.1 ± 0.0	0.6 ± 0.1	7.0 ± 1.5
K04	126.5	61.6	4.6	1.3 ± 0.2	< 0.0	0.2 ± 0.1	1.5 ± 0.2	6.3 ± 2.1
K05	182.6	77.8	4.7	< 0.0	< 0.0	0.1 ± 0.0	0.1 ± 0.0	N/A
K06	296.1	47.8	6.0	0.4 ± 0.1	< 0.1	0.3 ± 0.1	0.7 ± 0.2	1.4 ± 0.7
K07	121.8	60.4	4.9	< 0.1	< 0.1	0.3 ± 0.1	0.4 ± 0.1	N/A
K08	121.4	50.8	5.1	0.5 ± 0.2	< 0.1	0.4 ± 0.1	1.0 ± 0.2	1.2 ± 0.5
K09	122.6	65.5	3.7	< 0.1	< 0.1	< 0.1	0.1 ± 0.1	N/A
K10	115.1	73.6	3.0	0.1 ± 0.0	< 0.0	0.1 ± 0.0	0.1 ± 0.0	0.9 ± 0.7
K11	133.3	61.2	5.7	0.2 ± 0.1	< 0.1	0.2 ± 0.1	0.5 ± 0.1	1.0 ± 0.5
K12	283.3	49.4	4.0	< 0.1	< 0.1	< 0.1	0.1 ± 0.1	N/A
K13	230.4	55.3	4.7	< 0.1	< 0.1	< 0.1	0.2 ± 0.1	N/A

Table 6. Grass samples: alpha spectrometry results showing total uranium and isotopic ratios - KTA 2008.

Sample descriptor	Fresh weight (g)	Dry weight (g)	Ashed weight (g)	Measured activity of dry sample (mBq/g)				$^{238}U/^{234}U$ ratio
				^{238}U	^{235}U	^{234}U	Total U	
K4/S/U	133.0	34.5	28.4	16.6 ± 2.3	0.7 ± 0.3	18.0 ± 2.5	35.3 ± 3.4	0.9 ± 0.2
K/4/S/L	266.5	40.6	34.3	20.0 ± 2.7	0.7 ± 0.3	20.3 ± 2.8	41.0 ± 3.9	1.0 ± 0.2
K5/S/U	123.5	30.8	24.6	121.0 ± 14.5	2.0 ± 0.5	29.2 ± 3.8	152.2 ± 15.0	4.1 ± 0.7
K5/S/L	212.5	46.0	39.0	266.2 ± 31.0	3.5 ± 0.7	50.0 ± 6.2	319.6 ± 31.7	5.3 ± 0.9
K6/S/U	159.7	31.4	26.9	23.1 ± 3.7	< 1.0	22.6 ± 3.7	46.3 ± 5.3	1.0 ± 0.2
K6/S/L	244.4	45.4	40.4	19.4 ± 2.6	0.7 ± 0.3	17.1 ± 2.4	37.2 ± 3.6	1.1 ± 0.2
K13/S/U	117.0	23.5	16.5	36.3 ± 4.9	1.8 ± 0.6	32.1 ± 4.4	70.1 ± 6.7	1.1 ± 0.2
K13/S/L	207.6	40.1	32.5	15.9 ± 2.3	0.8 ± 0.3	17.3 ± 2.5	34.0 ± 3.4	0.9 ± 0.2

Table 7. Soil samples: alpha spectrometry results showing total uranium and isotopic ratios – KTA 2008.

Sample descriptor	Fresh weight (g)	Dry weight (g)	Ashed weight (g)	Measured activity of dry sample (mBq/g)				$^{238}U/^{234}U$ ratio
				^{238}U	^{235}U	^{234}U	Total U	
K6 Cow	221.1	41.2	3.2	0.3 ± 0.1	< 0.1	0.5 ± 0.2	0.9 ± 0.2	0.6 ± 0.3
K7 Cow	158.2	26.2	3.9	0.6 ± 0.2	< 0.1	0.6 ± 0.2	1.2 ± 0.2	0.9 ± 0.4
K8 Cow	186.0	32.2	5.1	0.4 ± 0.1	< 0.1	0.6 ± 0.1	1.0 ± 0.2	0.8 ± 0.3
K10 Cow	191.7	32.2	4.1	0.6 ± 0.4	< 0.5	0.8 ± 0.4	1.5 ± 0.6	0.8 ± 0.6
K1 Rabbit	37.2	8.8	0.7	0.3 ± 0.2	< 0.2	0.3 ± 0.2	0.7 ± 0.2	0.9 ± 0.7
K4 Rabbit	31.7	9.9	0.8	0.7 ± 0.3	< 0.2	0.6 ± 0.2	1.4 ± 0.4	1.2 ± 0.6
K13 Rabbit	82.1	11.0	1.2	0.6 ± 0.3	< 0.3	0.6 ± 0.3	1.2 ± 0.4	1.0 ± 0.6
K3 Deer	47.6	7.1	0.7	< 0.3	< 0.3	< 0.3	0.5 ± 0.2	N/A
K11 Sheep	150.2	46.0	17.4	3.6 ± 0.6	< 0.1	3.7 ± 0.6	7.4 ± 0.8	1.0 ± 0.2

Table 8. Animal indicators (faeces): alpha spectrometry results showing total uranium and isotopic ratios – KTA 2008.

Sample descriptor	Fresh weight (g)	Dry weight (g)	Ashed weight (g)	Measured activity of dry sample (mBq/g)				$^{238}U/^{234}U$ ratio
				^{238}U	^{235}U	^{234}U	Total U	
Deer Kidney (a)	48.3	12.5	0.6	< 0.3	< 0.2	< 0.2	0.1 ± 0.1	N/A
Deer Kidney (b)	46.7	12.7	0.6	< 0.4	< 0.2	< 0.3	0.1 ± 0.1	N/A
Deer Liver (a)	250.8	55.7	3.6	< 0.1	< 0.1	< 0.1	0.1 ± 0.0	N/A
Deer Liver (b)	246.7	45.7	2.1	< 0.3	< 0.2	< 0.2	0.1 ± 0.1	N/A

Table 9. Animal indicators (deer kidneys and livers): alpha spectrometry results showing total uranium and isotopic ratios – KTA 2008.

Sample descriptor	Measured activity of dry sample (mBq/l)				$^{238}U/^{234}U$ ratio
	^{238}U	^{235}U	^{234}U	Total U	
W1	< 10.0	< 10.0	< 10.0	6.0 ± 4.5	N/A
W2	< 7.3	< 7.3	< 7.3	3.8 ± 3.1	N/A
W3	< 5.5	< 5.5	< 5.5	4.4 ± 2.9	N/A
W4	< 5.1	< 5.1	< 5.1	2.7 ± 2.2	N/A
W5	< 10.0	< 10.0	< 10.0	6.4 ± 4.7	N/A

Table 10. Water samples: alpha spectrometry results showing total uranium and isotopic ratios – KTA 2008.

Sample descriptor	Measured activity of dry sample (mBq/g)				$^{238}U/^{234}U$ ratio
	^{238}U	^{235}U	^{234}U	Total U	
W1	< 1.0	< 1.0	< 1.0	0.9 ± 0.5	N/A
W2	< 1.0	< 1.0	< 1.0	0.5 ± 0.4	N/A
W3	< 1.0	< 1.0	< 1.0	0.8 ± 0.5	N/A
W4	< 1.0	< 1.0	< 1.0	0.5 ± 0.4	N/A
W5	< 0.9	< 0.9	< 0.9	0.9 ± 0.5	N/A

Table 11. Water samples (filter paper analysis): alpha spectrometry results showing total uranium and isotopic ratios - KTA 2008.

Sample descriptor	Fresh weight (g)	Dry weight (g)	Ashed weight (g)	Measured activity of dry sample (mBq/g)				$^{238}U/^{234}U$ ratio
				^{238}U	^{235}U	^{234}U	Total U	
S1	306.7	87.0	85.0	11.6 ± 1.7	0.5 ± 0.2	13.5 ± 1.9	25.6 ± 2.6	0.9 ± 0.2
S2	360.2	85.4	82.9	16.2 ± 3.0	< 1.1	13.5 ± 2.6	30.0 ± 4.0	1.2 ± 0.3
S3	359.2	88.1	84.8	15.2 ± 2.3	0.7 ± 0.3	16.5 ± 2.5	32.4 ± 3.4	0.9 ± 0.2
S4	350.9	56.3	52.1	14.7 ± 2.2	0.8 ± 0.3	16.4 ± 2.4	31.9 ± 3.2	0.9 ± 0.2
S5	233.1	40.7	36.8	16.3 ± 2.5	0.9 ± 0.4	18.5 ± 2.7	35.6 ± 3.7	0.9 ± 0.2
S6	193.6	28.6	24.7	14.2 ± 2.1	0.5 ± 0.3	15.5 ± 2.2	30.2 ± 3.1	0.9 ± 0.2
S7	340.4	78.4	74.9	7.4 ± 1.3	0.4 ± 0.2	8.6 ± 1.5	16.4 ± 2.0	0.9 ± 0.2
S8	367.7	82.1	79.2	15.3 ± 3.5	< 1.5	12.3 ± 3.0	28.1 ± 4.6	1.3 ± 0.4
S9	287.4	50.7	47.4	12.4 ± 1.9	0.7 ± 0.3	14.5 ± 2.1	27.7 ± 2.9	0.9 ± 0.2
S10	373.2	70.8	67.1	10.5 ± 1.6	< 0.3	11.5 ± 1.7	22.2 ± 2.4	0.9 ± 0.2
S11	273.0	48.1	44.1	12.3 ± 1.9	0.6 ± 0.3	14.9 ± 2.2	27.8 ± 2.9	0.8 ± 0.2
S12	292.7	41.6	37.9	18.7 ± 2.8	0.8 ± 0.4	19.1 ± 2.9	38.5 ± 4.0	1.0 ± 0.2
S13	232.9	47.9	42.5	15.5 ± 2.2	0.9 ± 0.3	19.7 ± 2.7	36.1 ± 3.5	0.8 ± 0.2
S14	144.5	29.9	24.9	13.6 ± 2.1	0.8 ± 0.3	17.0 ± 2.5	31.4 ± 3.3	0.8 ± 0.2
S15	261.5	49.7	45.1	11.9 ± 1.7	0.5 ± 0.2	12.8 ± 1.8	25.2 ± 2.5	0.9 ± 0.2

Table 12. Stream sediments: alpha spectrometry results showing total uranium and isotopic ratios – KTA 2008.

Note (for Tables 5 to 11): Activity results have been rounded to 1 decimal place. All uncertainties are stated at a 95% confidence level. Limits of Detection (LOD) are calculated by a 'modified Currie' formula[5] at 95%. The total activity is calculated from the sum of the actual activities for each isotope, regardless of the LOD quoted for that isotope. Therefore, where activities are reported as less than LOD for any of the uranium isotopes, the total uranium value may not be equal to the sum of the individual isotopic values. Where the isotopic activities are below the LOD for more than one isotope, the total activity and the isotopic ratios could not be calculated and are reported as n/a.

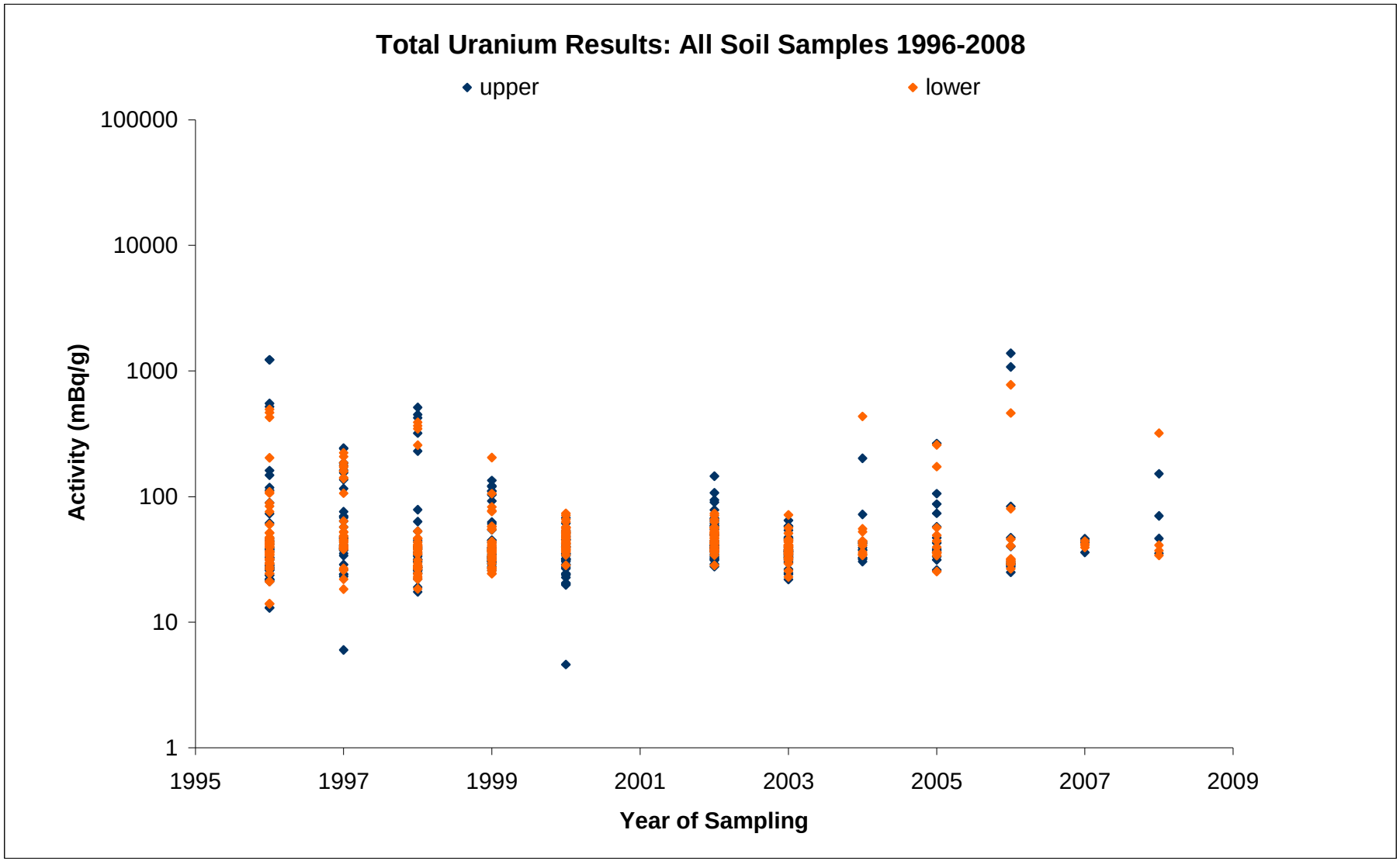
[5] Hurtgen C, Jerome S, Woods M. (2000) 'Revisiting Currie - how low can you go?' *Applied Radiation and Isotopes* 53 pp 45-50

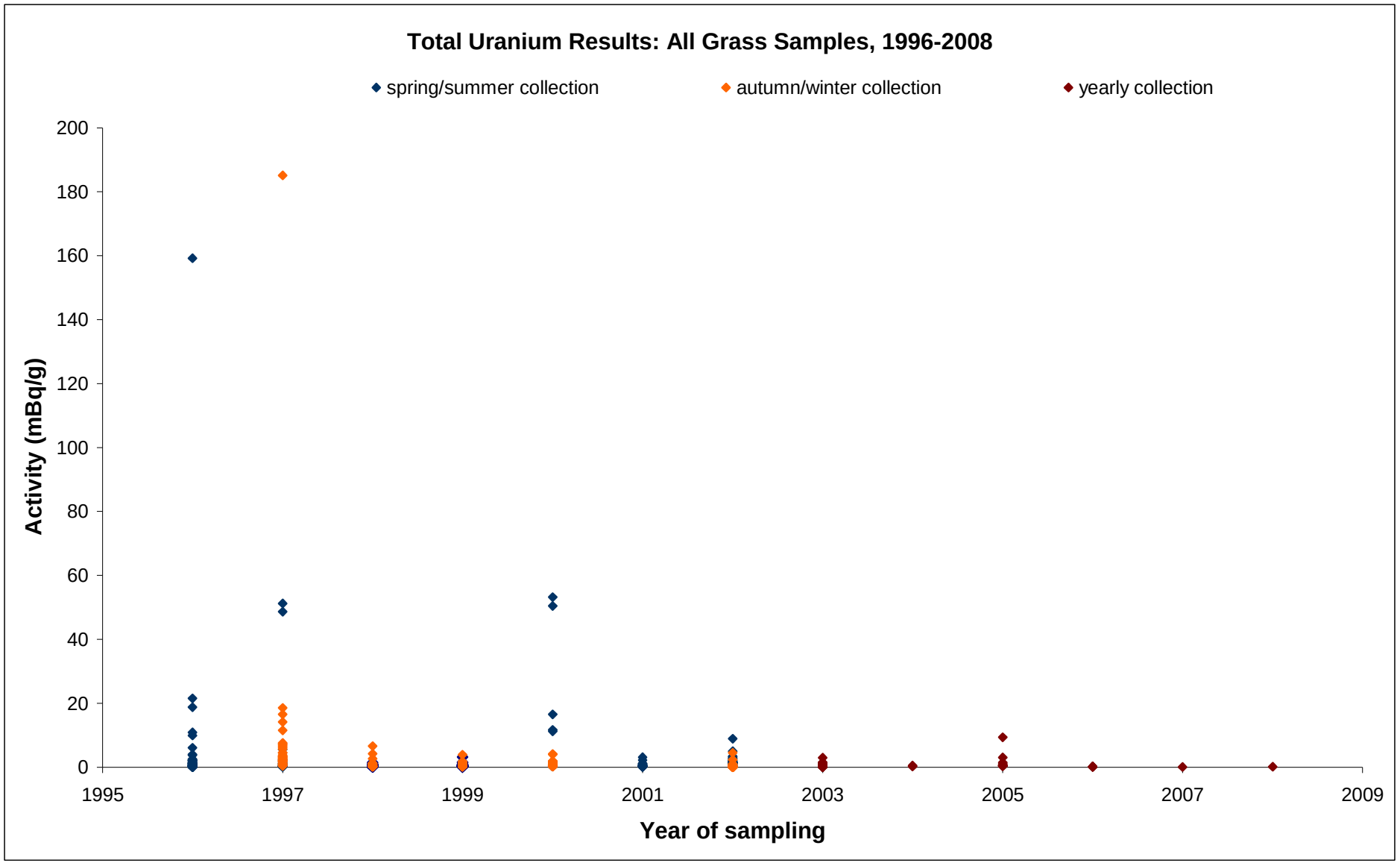
Station number	Location	Average dose rate (nGy/h) (n = 3)
1	Raeberry Target	91
2	Raeberry Bunker	116
3	Raeberry Gun	102
4	India Target	90
5	Balig Gun/Target waypoint	91
6	Balig Gun	91
7	Zulu Gantry	86
8	Mullock Farm	87
9	Silver Hill (Low) Gun	113
10	Echo Target (Doon Hill)	81
11	Doonhill/Target waypoint	89
12	Doon Hill Gun	85
13	Gypsy Point	81

Table 13. Dose rates measurements - KTA 2008.

Historical Data

To allow year-to-year comparison, data from the 1996 to 2008 KTA environmental surveys is presented on the following pages. Data for soil, grass and animal indicator analyses are presented separately.


Total Uranium Results: All Soil Samples 1996-2008
upper
lower
100000
10000
1000
100
10
1
Activity (mBq/g)
1995
1997
1999
2001
2003
2005
2007
2009
Year of Sampling


Total Uranium Results: All Grass Samples, 1996-2008
spring/summer collection
autumn/winter collection
yearly collection
Activity (mBq/g)
200
180
160
140
120
100
80
60
40
20
0
1995
1997
1999
2001
2003
2005
2007
2009
Year of sampling

Sample Type	1996	1997	1998	1999	2000		2002	2003	2004	2005	2006	2007	2008
Cow Faeces	2.1-5.6 *(0.8-1.3)*	0.8-1.3 *(0.8-1.0)*	1.4 *(0.9)*	0.8 *(0.7)*	31.5 *(1.1)*	*No sampling in 2001*	0.2 *(0.8)*	1.6 *(0.5)*	0.5-1.5 *(0.6-0.8)*	0.4-7.7 *(0.7-1.5)*	1.1-1.5 *(0.9-1.0)*	1.0-2.6 *(0.7-1.2)*	0.9-1.5 (0.6-0.9)
Sheep Faeces	<0.3	1.1 *(1.1)*	3.1 *(0.6)*	0.7-7.8 *(0.9-1.7)*	3.9 *(1.3)*		5.2 *(0.7)*	0.9 *(1.5)*	0.5 *(1.5)*	×	0.3 *(1.0)*	1.2-14.5 *(0.8-0.9)*	7.4 (1.0)
Fox Faeces	×	×	11.4 *(0.7)*	6.0 *(0.8)*	46.7 *(6.7)*		8.7 *(0.6)*	5.2 *(2.2)*	×	×	12.4 *(1.8)*	×	×
Deer Faeces	×	2.1 *(n/a)*	7.8 *(2.7)*	0.4-0.6 *(1.7 ±1.5)*	×		1.0 *(0.7)*	0.5 *(0.7)*	×	1.4 *(1.0)*	0.5 *(2.7)*	×	0.5 (n/a)
Rabbit Faeces	×	×	0.3 *(0.9)*	×	3.9 *(2.0)*		42.6 *(7.2)*	0.5 *(1.0)*	0.5 *(1.5)*	1.2 *(0.9)*	1.3 - 5.8 *(1.2 – 1.4)*	1.2-13.2 *(1.1-2.3)*	0.7-1.4 (0.9-1.2)
Badger Faeces	×	×	×	×	×		×	×	×	×	×	0.1-1.2 *(1.3 ± 0.5)*	×
Deer Kidneys	×	×	×	×	×		×	×	×	0.1 ± 0.0 (n/a)	×	0.0-0.1 (n/a)	0.1 (n/a)
Mushroom	×	×	×	×	×		×	0.5 *(n/a)*	×	×	×	×	×
Black-berries	×	×	×	×	×		0.2 (n/a)	×	×	×	×	×	×
Water samples	4.4-8.1 (1.0-1.1)	<9.4 (n/a)	3.8-16.6 (1.1-1.4)	9.0-14.6 (n/a)	12.3-13.0 (n/a)		2.7-7.1 (n/a)	1.7-4.7 (n/a)	<11 (n/a)	1.8-5.1 (n/a)	1.8-9.1 (n/a)	2.0-7.5 (n/a)	2.7-6.4 (n/a)

Total uranium results for all terrestrial biological and vegetal indicator samples (mBq/g of dry weight) and water samples (in mBq/l) 1996-2008.

Notes: × denotes that no sample of this type was collected.

For simplicity, results for both the total uranium activity and the isotopic $^{238}U/^{234}U$ ratio are reported as follows:

- The activity is reported first, and followed by the ratio in italics in bracket.
- Where the isotopic activities are below the Limit of Detection (LOD) for more than one isotope, the isotopic ratios could not be calculated and are reported as ‘n/a’.
- When only one result is reported for a sample type, the result is reported together with its uncertainty. When there are more than one result for a sample type, the range (min – max) is reported. The analytical error for these results, which is not reported here, is typically below 20%.

ANNEX A Issues to be considered when interpreting or comparing uranium data

A.1 There are a number of issues that may give rise to uncertainties when interpreting or comparing uranium data. These include:

- Analytical technique;
- Statistical variation;
- Spatial variability;
- Temporal variability; and
- Species variation (for plant and animal samples).

Analytical approaches

A.2 Soil sample results may be reported as either dry weight or wet weight depending on whether the masses of the samples were obtained prior to or after drying. This will have implications for comparison of results between the surveys at Kirkcudbright, which are reported as dry weight and other UK uranium in soil data, which may be reported as wet weight. Samples reported as dry weight will appear to have concentrations of uranium approximately 20% higher than those reported as wet weight (although this will depend on the moisture content).

A.3 For analysis techniques such as inductively coupled plasma mass spectrometry (ICP-MS) or alpha spectrometry, the uranium present in a sample may be extracted into solution by either leaching the soil samples or totally dissolving them. Total dissolution will give rise to higher uranium results because the analysis will include all uranium including that which is contained within the mineral grains. Leached samples, in comparison, will only contain uranium that is either easily dissolved or is adhered to the surfaces of mineral grains. This limitation is acceptable as any DU contamination which may be present at Kirkcudbright is likely to be leachable. Total sample analysis techniques such as gamma spectrometry will give results similar to those for total dissolution. Given the differences between the results for total analysis and leached analysis, care should be taken when comparing sets of data to ensure that either the same approach has been used or that the differences are appropriately discussed.

A.4 Uranium concentrations in plants may be affected by contamination of foliage or roots with dust or soil particles. Preparation of plants for analysis may or may not involve a washing stage. It is therefore important to be aware of the preparation approaches that have been applied when comparing the results of different plant analyses.

Statistical variations

A.5 There will be minor variations between the true uranium content of a sample and results produced by analysis. This variation is highlighted in the counting statistics for the

technique. The statistical uncertainties of laboratory results are likely to be small in comparison with the true variation in activity between samples.

Spatial variations

A.6 DU contamination will not be uniformly distributed within a sampling area. Any DU firing malfunctions will probably result in small fragments distributed over an area. Within that area, sampling at some locations would indicate contamination, whereas at others the soil would appear to be clean. Hence, the repeat sampling and analysis of soils from within an area may give rise to a significant degree of variation.

A.7 In addition to DU contamination due to firing at the range, there may be variations in uranium concentrations due to local anthropogenic or natural discharges. For example, natural uranium concentrations may be enhanced by the local application of phosphate based fertilise to agricultural land. Most of the phosphate fertiliser applied to clay loam soils in England in the last 100 years can be still be found retained in the top 23 cm layer of soils, with no increase at greater depths [A1]. Veins rich in uranium minerals occur naturally along the coast of the Solway Firth, such as uraninite found at Needle's Eye, approximately 24 km away from KTA on the north coast of the estuary. These features are thought to be present across the region [A2, A3], although this has not been studied specifically.

Temporal variations

A.8 There will be natural temporal variations in the uranium concentration and in the abundance of the various isotopes in the samples due to seasonal variations in rainfall. Rainfall can impact on dust re-suspension and deposition on grass as well as on the dissolution and migration of surface uranium down the soil profile. Some concentrations may be slightly elevated at the time of sampling, whilst others may be slightly below expected background levels.

A.9 The activities of samples from any particular sampling site may vary from year to year. This may relate to temporal changes in uranium concentration, but will also be affected by spatial variation (see above).

A.10 Plant uptake of radionuclides is affected by the period in the plant growing cycle. This is also mirrored in the animal uptake of radionuclides within their life cycle.

Species variations

A.11 Plant uptake of radionuclides is affected by the soil characteristics (uranium concentration and speciation as well as other soil physico-chemical characteristics) and varies with plant species. In general, leafy vegetables take up higher concentrations than fruit and grain crops. Uranium tends to be preferentially distributed in the leaves and stems rather than in the roots, fruits or seeds [A1].

A.12 Animal uptake of uranium is affected by their life habits, feeding patterns, physiology and the uranium concentration in their foodstuffs and the environment. Because it is clearly impractical to consider all species in a particular environment, the current approach in radiological dose assessments to non-human species is to advocate the use of reference organisms[A4]. For the terrestrial environment, animals that live both above and below

ground have been selected (earthworms, mice and foxes), as well as animals that live strictly above ground (deer and cattle). Although this environmental survey does not seek to compile a dose assessment for non-human species, wherever possible, the faeces of some of these animals have been collected and analysed.

Annex A References

A1. Mortvedt, JJ. (1992) Plant and soil relationships of uranium and thorium decay series radionuclides – A review, *Journal of Environmental Quality*, 23, 643.

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A3. Milodowski, A E. et al. (1990) Uranium-mineralised micro-organisms associated with uraniferous hydrocarbons in southwest Scotland. Nature, 347, 465.

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[web reference: http:\\www.erica-project.org].

ANNEX B Change in the $^{238}U/^{234}U$ activity ratio of a medium containing natural uranium with the addition of depleted uranium

Mass proportion of DU added*	Activity Concentration (mBq/kg)**			Ratio of total activity natural uranium to total activity	$^{238}U/^{234}U$ activity ratio
	U-238	U-235	U-234		
0	3.7×10^4	1.7×10^3	3.8×10^4	1.0	1.0
1	7.4×10^4	2.2×10^3	4.3×10^4	1.6	1.7
2	1.1×10^5	2.7×10^3	4.8×10^4	2.1	2.3
3	1.5×10^5	3.2×10^3	5.3×10^4	2.6	2.8
4	1.9×10^5	3.7×10^3	5.8×10^4	3.2	3.2
5	2.2×10^5	4.1×10^3	6.3×10^4	3.8	3.6
6	2.6×10^5	4.6×10^3	6.8×10^4	4.3	3.9
7	3.0×10^5	5.1×10^3	7.2×10^4	4.9	4.1
8	3.4×10^5	5.6×10^3	7.7×10^4	5.4	4.3
9	3.7×10^5	6.0×10^3	8.2×10^4	6.0	4.5
10	4.1×10^5	6.5×10^3	8.7×10^4	6.5	4.7
20	7.8×10^5	1.1×10^4	1.4×10^5	12.0	5.8
60	2.3×10^6	3.1×10^4	3.3×10^5	34.1	6.9
80	3.0×10^6	4.0×10^4	4.3×10^5	45.1	7.1
100	3.8×10^6	5.0×10^4	5.2×10^5	56.1	7.2
200	7.5×10^6	9.8×10^4	1.0×10^6	111.0	7.4
600	2.2×10^7	2.9×10^5	3.0×10^6	332.0	7.6
800	3.0×10^7	$3..9 \times 10^5$	3.9×10^6	442.0	7.6
1000	$3.73\ 10^7$	4.8×10^5	4.9×10^6	552.0	7.6

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* The value represents the additional mass of depleted uranium added (all radionuclides) relative to the original mass of natural uranium present (3 mg U/kg soil).

** Table assumes 3 mg U/kg of natural uranium present in soil in following proportion: ^{238}U (2.978 mg /kg); ^{235}U (0.022 mg /kg); ^{234}U (2e-04 mg /kg), prior to addition of DU.

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13. Abstract

Depleted uranium (DU) ammunition has been test fired at the Kirkcudbright Training Area (KTA) since 1982. Routine environmental monitoring has been carried out at KTA since 1980 to assess the environmental impact of the firings on the terrestrial and marine environments and any associated radiological hazards.

This report presents the findings of the terrestrial survey undertaken at KTA during 2008; the marine survey is reported separately in Part 2. No sample analysed was radioactive within the meaning of the Radioactive Substances Act 1993 nor exceeded the Generalised Derived Limits set by the Health Protection Agency. Potential exposure pathways and doses to critical groups, site personnel and members of the public have been assessed and are deemed to be insignificant compared to the local natural background.

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